# Review of dermal effects and uptake of petroleum hydrocarbons 

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#### Abstract

This report serves as an update to and an extension of a previous CONCAWE report on dermal absorption of petroleum hydrocarbons (Petroleum hydrocarbons: their absorption through and effects on the skin, CONCAWE Report 84/54, 1984).

To contribute to health risk assessments associated with dermal exposure, this report evaluates experimental data to determine the extent to which petroleum hydrocarbons pass through the skin. These data strongly suggest that dermal exposure to petroleum hydrocarbons, even following long-term exposures such as in occupational settings, will not cause systemic toxicity under normal working conditions and assuming an intact skin barrier. Skin contact with some petroleum products may cause skin irritation, leading to dermatitis, particularly after repeated or prolonged exposure. In addition to these irritating effects, the skin barrier function may be affected following repeated contact with petroleum hydrocarbons, making the skin potentially more susceptible to other irritants, sensitizing agents, and bacteria. In addition, the impaired skin barrier may lead to increased dermal penetration of hydrocarbons and other substances. To avoid this there is a need to minimise skin contact.


## KEYWORDS

Dermal exposure; absorption; permeability; flux; QSAR; skin; risk assessment; petroleum hydrocarbons

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## SUMMARY

Commercial petroleum products, such as gasoline, kerosene, and fuel oils, are complex mixtures consisting of hundreds of different hydrocarbons. Dermal exposure to petroleum hydrocarbons, for instance in the work place, may contribute to systemic levels of various aromatic and aliphatic hydrocarbon components. To assess the health risk associated with dermal absorption, in addition to external exposure (i.e. total exposure onto the skin), the extent to which petroleum hydrocarbons pass through the skin needs to be evaluated. Dermal absorption can be assessed either by measurements in in vivo in humans, or predicted by using animal, in vitro, and in silico models (e.g. Quantitative Structure-Activity Relationships or QSARs). One of the specific problems in assessment of the absorption rates of petroleum hydrocarbons is their lipophilicity which is usually expressed based on the partitioning coefficient of the material between $n$-octanol and water ( $\mathrm{K}_{\mathrm{ow}}$ ), for instance the $\log \mathrm{K}_{\text {ow }}$ values for toluene and hexadecane are 2.7 and 8.2, respectively (indicating that hexadecane has a much smaller water solubility than toluene. In particular for petroleum hydrocarbons, study design regarding the type of skin preparation, vehicle, and the way dermal absorption is expressed might strongly influence the outcomes. Furthermore, the vast majority of the compounds on which the QSARs are based have $\log \mathrm{K}_{\mathrm{ow}}$ values below 4.4 (i.e. relatively water soluble compounds)). In addition, these models are mainly derived from in vitro experiments obtained from aqueous solutions raising the question whether use of these models for the prediction of highly lipophilic petroleum hydrocarbons is justified since humans are rarely, if ever, exposed to aqueous solutions of petroleum hydrocarbons.

The main aim of this report is to critically evaluate the available experimental data on dermal absorption of petroleum hydrocarbons. Furthermore, permeability coefficient ( Kp ) and maximum flux data predicted by QSARs were compared with experimental values obtained from petroleum hydrocarbons in aqueous solutions, as neat individual hydrocarbons, and as the products themselves. The literature data on Kp and flux obtained from in vivo and in vitro studies have been evaluated with regard to the vehicle and the type of skin. For prediction of dermal absorption, two QSARs (Episuit-Dermwin and Skinperm) have been used.

The results indicate that the experimentally determined absorption from aqueous solutions were several orders of magnitude higher than the absorption after dermal exposure to either a neat chemical or a petroleum product such as a jet fuel. Furthermore, the experimental $\mathrm{K}_{\mathrm{p}}$ values show inverse proportionality with lipophilicity of the hydrocarbons which is in contrast with the trend obtained by both prediction models. The predicted $\mathrm{K}_{\mathrm{p}}$ values were similar for both models and showed a significant overprediction when compared with the $\mathrm{K}_{\mathrm{p}}$ values obtained in studies performed with petroleum products or neat solvents. It was found that there is a better agreement between predicted and experimental values for aqueous solutions. However, for the maximum flux, a decrease in maximum flux with increasing $\mathrm{K}_{\mathrm{ow}}$ is obtained for both models, which is similar to the trend obtained in experimental studies.

Overall, it is concluded that the absorption rates of petroleum hydrocarbons from aqueous solutions highly overestimated those from neat petroleum products emphasizing the need for conducting absorption studies under "in use" scenarios. Furthermore, the maximum flux that defines the highest dermal exposure risk for a chemical, and QSARs based on this approach, may offer a better approach to dermal risk assessment of highly lipophilic petroleum hydrocarbons.

## 1. INTRODUCTION

As the first organ in contact with the external environment, the skin can be frequently exposed to various liquid chemicals (by spills, splashes, immersion, or contacts with contaminated surfaces) or vapours. Absorption of a chemical into the skin may lead to local effects such as inflammation or sensitization or to systemic effects after subsequent uptake and distribution by the circulation. In a workplace situation absorption of hazardous substances through the skin can contribute considerably to the total systemic uptake or in some cases even become the main absorption route. To assess the health risk associated with dermal exposure, the extent to which a substance that is deposited onto the skin can be expected to pass through the skin needs to be evaluated.

In the first part of this document an overview has been given of the methodology used for the assessment of dermal exposure and absorption of chemicals, including experimental methods and predictive modelling. Furthermore, test methods for local skin effects such as irritation, allergic sensitisation and corrosion have been addressed.

In addition to methodological issues, this report presents recent literature studies on the dermal absorption and local skin effects of petroleum hydrocarbons.

Although dermal absorption of petroleum hydrocarbons can be determined in vivo in human skin, for ethical and technical reasons, various in vitro and animal models have been used to predict dermal absorption in humans. Recently, new guidelines for the conduct of these studies have been adopted by regulatory bodies. Among animal models, pig skin is considered the best choice since rat skin over-predicts dermal absorption through human skin. In in vitro testing the choice of receptor fluid, type of skin and vehicle have been shown to be particularly important factors for highly lipophilic compounds. For some hydrocarbons including benzene, toluene and xylene the absorption from aqueous solutions was several orders of magnitude higher than that seen after dermal exposure to either a neat chemical or a petroleum product e.g. jet fuel. Since petroleum products are complex mixtures with varying composition, experimental conditions must be relevant for the situation of the worker or consumer.

The most widely applied QSAR models for the prediction of dermal absorption, such as Episuit-Dermwin and Skinperm, relate the permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$ to some physicochemical parameters of the compound such as molecular weight (MW), water solubility and lipophilicity ( $\mathrm{K}_{\mathrm{ow}}$ ). One of the specific problems in the application of these models for the prediction of dermal absorption of hydrocarbons is that the vast majority of the compounds on which the QSARs are based have $\mathrm{P}_{\mathrm{ow}}$ ( $\log \mathrm{K}_{\mathrm{ow}}$ ) values below $\sim 4.4$. For the numerous constituents in petroleum products that have $\mathrm{P}_{\text {ow }}$ values greater than 4.4, the $\mathrm{K}_{\mathrm{p}}$ predictions may therefore be less reliable. For highly lipophilic hydrocarbons ( $\mathrm{P}_{\mathrm{ow}}>4$ ), the Skinperm and Dermwin models significantly overpredict the $\mathrm{K}_{\mathrm{p}}$. However, using the Dermwin program to predict the dermally absorbed dose per unit area per event (DAevent) may result in a more realistic estimate of the dermal absorption. Another problem in relation to hydrocarbons is that most predictive models are based on permeability data obtained from aqueous solutions, which were markedly higher than comparable data obtained with neat chemicals or lipophilic mixtures. In the case of petroleum hydrocarbons, the maximum flux provides a better estimate of dermal absorption than $\mathrm{K}_{\mathrm{p}}$. The maximum dermal flux can be estimated by multiplying the permeability coefficient obtained in water by its water solubilty. In the case of mixtures the
maximum flux of each constituent may be considered as a very worst case approach, since we are not dealing with neat solutions but compounds that represent only a fraction of the solution mixture.

In spite of all the shortcomings with the methods for estimating dermal absorption, the presently available models (Quantitative Structure-Permeation Relationships QSPeRs) can help in defining compounds (solutes) as good, bad or intermediate skin permeants.

Dermal exposure of chemicals can be quantified using various measurement techniques or predicted by using predictive models. The use of these methods in assessment of dermal exposure to petroleum hydrocarbons is hampered by the lack of method standardisation and validation.

An extensive literature review shows that individual petroleum hydrocarbons are able to penetrate into and through the skin when applied as pure substances, in mixtures or as commercial products. Generally, a higher absorption of aromatics (e.g. toluene, xylene, naphthalene) than of aliphatics (decane, dodecane, tridecane and hexadecane) has been reported. Available data on dermal absorption of individual hydrocarbons from petroleum products such as jet fuel, strongly suggest that dermal exposure to these products would not cause systemic toxicity under normal working conditions and assuming an intact skin barrier. Pre-exposure to petroleum products might lead to a higher absorption of the hydrocarbons implying an alteration of the skin barrier by repeated exposure. Individual hydrocarbons as well as petroleum products are also known to cause skin irritation, mainly due to the damaging effect on the skin barrier. In general, aromatic hydrocarbons are more irritating to the skin than aliphatic hydrocarbons. However, aliphatic hydrocarbons are retained longer in the skin due to their lipophilicity and have been shown to cause skin irritation which peaks around tetradecane (C14). Animal studies provide evidence that severe dermal irritation induced by repeated long-term exposure to petroleum fuels can contribute to the progression-promoting effect and development of skin tumours. Experimental data indicate that the tumour-promoting activity of alkanes is related to their chain length, with maximal activity found in C12-C14 alkanes. It is plausible to assume that repeated, long-term exposure to petroleum dermally irritating hydrocarbons may pose a carcinogenic risk, which is minimised if long-term irritation is prevented.

In contrast to the skin irritating effect, there is no evidence that petroleum hydrocarbons cause skin sensitization.

## 2. METHODS FOR THE ASSESSMENT OF DERMAL ABSORPTION

Dermal absorption can be assessed either by measurements in in vivo experimental exposure studies in humans, or can be predicted by using animal, in vitro and in silico models (e.g. QSPeRs).

This chapter is an overview of the experimental methods available for measuring dermal absorption.

### 2.1. EXPERIMENTAL METHODS

### 2.1.1. In vivo methods

The most appropriate approach for the studying of percutaneous penetration for risk assessment purposes would be to undertake studies in human volunteers. Since their use is limited and their conduct is closely regulated due to technical and ethical constraints (ICH, 1996; Declaration of Helsinki, 2004), an alternative approach to generate in vivo dermal absorption data is to use laboratory animal models. This section gives an overview of in vivo methodology for dermal absorption in both human volunteers and in laboratory animals.

### 2.1.1.1. In vivo methods in human volunteers

Human in vivo studies are considered as the "golden standard" against which all other alternative methods such as in vitro assays and animal models should be judged (Howes et al., 1996). The main advantage of human in vivo methods is that they allow for the determination of the extent of systemic absorption of the test substance in a natural, unaffected living system.

Generally in in vivo studies, the chemical is applied as neat or as a solution in an appropriate vehicle onto a designated surface area of the skin for a defined time period. The chemical is usually applied to the forearm or back skin. For vapours, whole body exposure can also be performed.

Dermal absorption can be assessed using different approaches. Common methods for determination of in vivo dermal absorption in humans include the measurement of the parent chemical and/or metabolite levels in biological material (e.g. blood, urine, and exhaled air), the microdialysis technique and tape stripping.

## Plasma and/or excreta measurements

The extent of dermal absorption of chemicals can be assessed by measuring the amount of the parent chemical or its metabolite in plasma, exhaled air or urine (Akrill et al., 2002; Brooke et al., 1998; Cardona et al., 1993). The amount of chemical determined after dermal exposure is compared to that after a reference exposure with a known input rate or dose such as intra-venous administration or inhalation (in case of volatile compounds). If the total amount of the chemical (e.g. total urine excretion) or the area under the blood/urine concentration-time profile (Area Under the Curve - AUC) of the two routes is compared, the dermal absorbed dose can be calculated from:

Absorbed dose dermal $=\left(\text { AUC or Excretion } \text { dermal } / \text { AUC or Excretion }{ }_{\text {ref }}\right)^{*}$ Dose $_{\text {ref }}$

From the absorbed amount, exposed skin area and exposure duration, the average rate of absorption into the skin throughout the exposure can be deduced:

Absorption rate dermal $=$ Absorbed dose dermal $/($ Area*Time $)$
Since this absorption rate is not constant during the exposure but dependent on the exposure duration, it cannot easily be extrapolated to other exposure scenarios. For more detailed dermal kinetics, concentration-time profiles have to be determined. Using appropriate kinetic analysis such as the (de)convolution method, dermal absorption rates as a function of time can be deduced from the concentration-time profiles (Opdam, 1991). The advantage of this approach is that, in addition to the average absorption rate into or through the skin, the maximum absorption rate can be deduced and, if a steady state absorption is reached, also the permeability coefficient ( $\mathrm{K}_{\mathrm{p}}$ ) (Opdam, 1991; Kezic et al., 2004). This method has been widely used for the determination of dermal absorption for solvents, drugs and other chemicals (e.g. Kezic et al., 2001; Jakasa et al, 2004).

The measurement of the parent chemical and/or its metabolite in plasma and excreta is of practical importance for risk assessment since it is expected to give an indication of the internal dose which would be expected to give a better correlation for any adverse health effects than external exposure (i.e. total exposure onto the skin). Especially in the case that dermal absorption contributes substantially to the total absorption, estimation of the internal dose by means of biological monitoring has to be preferred to environmental monitoring (since the latter implicitly assumes inhalation as the only route of uptake), particularly when specific markers are available. In addition to the occupational exposure levels for airborne exposure, their biological equivalents known as Biological Exposure Indices have been set for a number of chemicals (ACGIH, 2001).

## Microdialysis

Microdialysis is a technique that measures an amount of a dermally applied chemical in the extracellular space beneath the exposed skin site. Recent reviews offer an in-depth view of methodological aspects and applications of the microdialysis technique (de Lange et al., 2000; Joukhadar and Muller, 2005; Kreilgaard, 2000; Schnetz and Fartasch, 2001) The principle of the technique is based on passive diffusion of chemicals across the semi-permeable membrane of a microdialysis probe that is introduced into the subcutaneous tissue parallel to the skin surface. The probe is usually perfused slowly with a physiological solution to mimic blood flow. Molecules able to diffuse across the probe membrane can be analysed in the perfusate which is collected at certain time points. The microdialysis technique enables the determination of detailed kinetics without the necessity to perform a reference exposure (Benfeldt et al., 2007; Benfeldt, 1999). However, this technique also has limitations. Implantation of the probe can elicit a tissue reaction, which in turn can influence skin absorption (Anderson et al., 1996; Schnetz and Fartasch, 2001). Furthermore, only a fraction of the chemical that has penetrated the skin is recovered in the dialysate. The recovery efficiency depends on several experimental factors such as position of the tubing, physico-chemical properties of the chemical and the nature of the perfusate. (Kreilgaard, 2000; Schnetz and Fartasch, 2001). An additional limitation of the microdialysis technique is that it utilizes an aqueous perfusate and can therefore only dialyse water-soluble substances.
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## Tape stripping

The tape stripping method is based on determination of the amount of chemical in the consecutive layers of the stratum corneum ${ }^{1}$. After the end of a dermal exposure period, the stratum corneum of the exposed skin site is removed sequentially by adhesive tape. The amount of recovered substance in each tape strip is determined with an appropriate analytical technique. Depending on the exposure period, time of stratum corneum harvesting, and the part of stratum corneum which is taken for the analysis, different approaches are proposed (Chao et al., 2005; Dupuis et al., 1984; Mattorano et al., 2004).

In some studies, the amount of chemical is measured only in the superficial layers of the stratum corneum. It has been reported that the amount of the chemical in these stratum corneum layers was a good estimate of the total amount of the chemical absorbed into the systemic circulation (Chao et al., 2005; Dupuis et al., 1984; Mattorano et al., 2004; Rougier et al., 1983). The main problem with this methodology is the variability in the amount of the chemical recovered by each tape strip. The recovery is influenced by several factors; type of adhesive tape, vehicle in which the chemical is applied and the applied pressure on the tape prior to removal from the skin site. To avoid this source of variation, the amount of stratum corneum could be determined by measuring the weight of the stratum corneum in the strips or be estimated indirectly by, for example, the protein content or by assessing transepidermal water loss (Dreher et al., 1998; Jakasa et al., 2007; Pirot et al., 1997; Tsai et al., 2003; Weigmann et al., 2003).

Instead of measuring the total amount of the chemical in the stratum corneum, in some studies the concentration profile of the chemical across the entire stratum corneum depth has been determined. From the concentration-depth profiles, the diffusion coefficient and partition coefficient of the chemical between vehicle and the stratum corneum equation can be deduced which allows estimation of the permeability coefficient. (Alberti et al., 2001; Jakasa et al., 2007; Pirot et al., 1999; Reddy et al., 2002)

Although the tape stripping technique has certain advantages, there are critical points. Some authors use the number of consecutive tape strips instead of measuring the real stratum corneum depth thereby assuming that the amount of the removed stratum corneum by each strip is linearly proportional to the number of strips (Tsai et al., 1991). However, the amount of removed stratum corneum is known to vary considerably for different individuals and with the depth (Bashir et al., 2001; Tsai et al., 1991). The homogeneity of subsequent stratum corneum layers is also a point of concern: due to the furrows in the stratum corneum the amount of the chemical measured in one strip can come from different layers of the skin (van der Molen et al., 1997). Furthermore, the time taken to remove the entire stratum corneum can be critical for the determination of fast penetrating or volatile chemicals (Reddy et al., 2002).

## Spectroscopic methods

A variety of spectroscopic methods have been applied in studying in vivo penetration of chemicals, with the vast majority being based on infrared (IR) and Raman vibrational spectroscopy (Alvarez-Roman et al., 2004; Hanh et al., 2000; Touitou et al., 1998). The main advantage of these spectroscopic techniques is that

[^0]they are rapid, non-invasive, and some of them can provide real time data on chemical penetration through the skin.

The most frequently used IR technique for studying penetration is attenuated total reflectance Fourier transform infrared spectroscopy (ATR-FTIR) in which the spectra can be recorded directly from the skin, which is placed onto the ATR crystal. The amount of a test chemical taken up can be quantified as a function of time, by following its disappearance from the most superficial layers of the stratum corneum (Mak et al., 1990). Since the IR beam has a low penetration depth of only 1-2 $\mu \mathrm{m}$ (Laugel et al., 2001; Nothinger and Imhof, 2004), the measurements of IR spectra as a function of different depths are performed in combination with tape stripping. The penetrating substance is measured on the treated site in vivo and after each removal of the stratum corneum or indirectly on the tape strips (Alberti et al., 2001; Pirot et al., 1997). ATR-FTIR has been extensively used to quantify the absorption of various drugs, pesticides and other chemicals (Alberti et al., 2001; Ayala-Bravo et al., 2003; Carden et al., 2005; Moser et al., 2001; Pirot et al., 1997; Reddy et al., 2002, Touitou et al., 2000).

Rather recently other IR techniques have been introduced such as thermal emission decay FTIR (Notingher and Imhof, 2004), opto-thermal radiometry and mid-infrared spectroscopy (Xiao et al., 2001; Ring et al., 2006). These techniques show less sensitivity to small movements and roughness of the skin sample. In addition, the spectra can be recorded up to a stratum corneum depth of $10 \mu \mathrm{~m}$ without interference from deeper layers. Confocal Raman microspectroscopy represents another recent in vivo approach for non-invasive determination of a molecular profile at various depths in the skin (Caspers et al., 2001). So far, this technique has been used mainly to identify molecular structures of the skin, and only a few studies have determined the diffusion profile of a chemical in the skin (Caspers et al., 1998; Caspers et al., 2000; Caspers et al., 2002). The fact that the technique is completely non-invasive, that it allows depth-profiling and that its spectrum is highly molecule specific make this technique a promising tool for determination of percutaneous absorption. The main limitation of spectroscopic techniques is that in general the chemical has to have a specific absorption spectrum that is sufficiently distinct from that of the stratum corneum.

### 2.1.1.2. In vivo methods in laboratory animals

Although human skin is the best model for human risk assessment and the acceptance of the use of laboratory animals is decreasing, there are several reasons why animal in vivo models are still used for dermal risk assessment. Firstly, due to practical and ethical issues, conduct of exposure studies in humans is not always possible. Secondly, in vitro tests are not accepted in some countries, and data from animal studies, in particular those obtained from the rat are required for the regulatory risk assessment process. (US-EPA, 1998; Zendzian, 2000) Furthermore, in comparison with in vitro tests, an in vivo animal model uses a physiologically and metabolically intact system.

In addition to ethical issues, the disadvantage of using laboratory animals is that they have different skin permeability and systemic disposition properties compared to humans. The skin structure differs from one species to another (for example, in the epidermis the stratum corneum is thicker in species without hair), between different strains of the same species and even within the same species (for instance, differences in absorption for the various parts of the body) (EU DG SANCO, 2004). While rats, guinea pigs and rabbits have a more permeable skin than humans, the skin absorption properties of monkeys and pigs are reported to be
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more similar to that of humans (OECD, 2004c). The results of a species comparative study performed by Bartek et al. (1972) indicated that the dermal absorption increases in the following order:
man < pig < rat < rabbit.
It is only recently that the OECD adopted a guideline for in vivo dermal absorption (OECD, 2004a,b). This is similar to the guideline for Dermal Absorption Studies of Pesticides in the rat published by the US-EPA (1998); however, the standard OECD protocol requires considerably fewer animals, since the number of exposure intervals and the number of dose levels is lower than in the US-EPA (1998) protocol.

## Principles of the standard in vivo tests

The experimental protocols for in vivo methods have been described in detail in the adopted OECD and US-EPA guidelines, so in this review the principles and methods will only be discussed briefly (EHC, 2006; OECD, 2004a,b; US-EPA, 1986).

There are three classes of in vivo animal studies. Studies 1) in which a compound is measured in blood or excreta, 2) in which material is measured in the skin by biopsy or some other method, and 3) in which the compound is measured in all tissues (residue analysis). (EHC, 2006; OECD, 2004a,b; US-EPA,1986)

The test chemical is applied to the designated area of skin in an appropriate form, for a defined period of time. Ideally, a radiolabeled chemical is used.

## Selection of animals

The most commonly used animal in dermal absorption studies is the rat because this allows for direct comparison with other in vivo studies, which are generally performed in this species (EHC, 2006; Zendzian, 2000). It is generally recognized that rat overpredicts human skin absorption and will therefore lead to a conservative estimate of percutaneous absorption for risk assessment. The factor of difference (the ratio between the values obtained in rat and human skin) appears not to be determined by molecular weight, lipophilicity, or aqueous solubility of a chemical, although in general better agreement was found for less lipophilic compounds. (Van Ravenzwaay and Leibold, 2004 a,b; Vecchia and Bunge, 2005; ECETOC, 1993) Due to inconsistent differences in percutaneous absorption between rat and human skin it is not possible to derive a general adjustment factor for estimation of human percutaneous absorption. However, when percutaneous absorption data are available for rat in vivo and for rat and human skin in vitro, the in vivo dermal absorption through human skin may be estimated from the relationship proposed by Van Ravenzwaay and Leibold (2004 b):

Human in vivo $=($ rat in vivo $) \times($ in vitro human $) /($ in vitro rat $)$

## Test substance and application

The test substance can be applied neat, diluted or in formulated mixtures to represent as close as possible real human exposure. Vehicles other than water can be used, but their interaction with the substance and absorption characteristics should be known. The amount applied should mimic normal use.

## Exposure duration

The OECD guideline (2004 a,b) requires that duration of exposure mimics usual human exposure (typically 6-24 hours) whereas a 24 h termination time point would allow estimation of daily systemic exposure. In contrast, the US-EPA guideline prescribes evaluation of dermal absorption for a series of exposure durations (USEPA, 1998). After the end of exposure, the skin site is washed in a manner that reflects normal hygiene practice (using aqueous soap) (OECD, 2004 a, b). In the US-EPA (1998) protocol, the animals are sacrificed at this time point. In the OECD (2004b) protocol, the animals are housed individually in metabolism cages after the skin is washed. Excreta and (if appropriate) exhaled air are collected until sacrifice of the animals.

## Analysis of data and reporting

The amount of a test chemical determined in blood, excreta, expired air and carcass and the cage after washing enables determination of the amount absorbed at each time point. Usually the dermal absorption is expressed as the percentage of the applied dose. Additional studies can be performed to determine whether test compound remaining in the layers of the skin at the exposure site will be available for systemic absorption or not. In the case of infinite dose, steady state absorption rate and permeability coefficient are determined.

### 2.1.2. In vitro methods

In vitro methods for assessing dermal absorption are based on measuring the diffusion of a test chemical into and across excised skin to a fluid reservoir.

A variety of protocols and experimental conditions have been used for determination of percutaneous absorption in vitro and it is only recently that standard guidelines for conducting in vitro tests have been adopted (OECD, 2004 a,b). In line with the OECD guidelines the criteria for in vitro assessment of dermal uptake of cosmetic ingredients were defined and updated (SCCNFP, 2003; 2006). In addition to these documents, the US Environmental Protection Agency proposed rules for in vitro testing of 33 chemicals aimed at evaluating the need for "skin designations" for these chemicals (US-EPA, 2004).

The adopted guidelines allow for some flexibility with respect to, for example, the type and thickness of the skin sample, dose levels, composition of receptor fluid and vehicle. Flexibility in the choice of experimental design was needed to enable measurements to be made under circumstances which are as close as possible to the real life exposure situation. However, experimental conditions such as type and preparation of the skin and the physico-chemical properties of the receptor fluid are known to influence the predictive value of an in vitro test for the human (in vivo) situation (Jones et al., 2004; EHC, 2006; Williams, 2006; Jakasa et al., 2007). An additional challenge in the dermal risk assessment process is the interpretation of the results from the in vitro study. The data obtained from percutaneous absorption experiments are used in a number of ways for the calculation of dermal absorption parameters and expression of the permeation characteristics. In this respect the question whether or not material retained in the various skin layers should be included in the overall dermal penetration value is still a point of debate. The factors which might influence the predictive value of percutaneous absorption tests have been extensively addressed in recent monographs and reviews (Jones et al., 2004; EDETOX, 2004; EHC 2006; Williams, 2006; Jakasa et al., 2007).

## Principles of the in vitro tests

For an in vitro skin-penetration study a piece of excised skin is positioned between two chambers of a diffusion cell which may be of either of static or flow-through design (Figure 1). The test substance is applied either neat or dissolved in an appropriate vehicle to the skin membrane and remains in contact with the skin for a defined period of time before removal by an appropriate cleansing procedure. The lower chamber (''receptor') contains a receptor fluid that is designed to mimic the role of the blood in vivo. The receptor fluid, containing the test chemical that has passed through the skin, is collected once at the end of the experiment or, preferably, at regular intervals during the exposure period so that a penetration time profile may be constructed. The amount of chemical in the skin and in the receptor fluid can be measured using appropriate analytical techniques such as high performance liquid chromatography, gas chromatography or scintillation counting. The use of radiolabeled chemicals (usually ${ }^{14}$ Carbon or tritium) is preferred because of the high detection sensitivity, especially if chemical-penetration levels are very low.

Maintaining a constant temperature at $32 \pm 1^{\circ} \mathrm{C}$ (temperature of the skin surface) is important since the rate and the extent of skin absorption is temperature-dependent.

## Diffusion cells

Both, static and flow-through diffusion cells are considered suitable (OECD, 2004a,b). The main difference between these two types of diffusion cells is in the design of the receptor compartment. Static cells are usually designed for use with larger skin areas and are equipped with larger receptor chambers with the receptor fluid collected manually through a sampling port (Figure 1b). In static diffusion systems it is important to ensure that the test compound is sufficiently soluble in the receptor fluid throughout the exposure period (Brain et al., 1998a; OECD, 2004b; EHC, 2006).

There are different types of flow-through diffusion cells, but basically there is a continuous flow of receptor fluid through the receptor cell just beneath the donor chamber (Figure 1a). The receptor fluid containing a test chemical is transferred through tubing (usually teflon) into the test vials placed in an automatic collector enabling simultaneous collection from a number of cells.

Figure 1 Schematic flow-through and static diffusion cells for measuring dermal absorption.
(a) Flow-through diffusion cell

(b) Static diffusion cell


Flow-through diffusion cells have the advantage that they more closely mimic the in vivo blood flow. However, dilution of the sample collected over time, reduces the analytical sensitivity.

Several comparative studies showed no systematic differences in skin absorption obtained with static or flow-through diffusion cells (Bronaugh and Maibach, 1985; Bronaugh and Stewart, 1985; Clowes et al., 1994; Hughes et al., 1993). From a large multi-center comparison study on absorption of three model permeants: caffeine, testosterone and benzoic acid, using both types of diffusion cells, the design of these cells appeared not to be a significant source of variation (Van de Sandt et al., 2004). Using a standard silicone rubber membrane, Chillcot et al. (2005) also did not reveal a consistent pattern of differences between the results of static and flow-through diffusion cells.

## Some other systems

Although in vitro assays are commonly performed using diffusion cells, there are a few other systems that may be very useful for determination of percutaneous absorption. Recently a new in vitro technique based on thermal gravimetric analysis has been proposed which appears to be very suitable for measurement of dermal absorption of vapours of volatile compounds such as 2-propanol, methanol and toluene (Rauma et al., 2006).

The isolated perfused porcine skin flap (IPPSF) and isolated pig ear are other in vitro skin models. The primary advantage of these systems as compared with diffusion chambers is that they allow measurement of percutaneous absorption in a viable skin preparation which has a normal anatomical structure and a functional microcirculation (Riviere et al., 1996; 1995; Riviere and Monteiro-Riviere, 1991; de Lange et al., 1994). This model has been used for investigation of different classes of chemicals such as xylene, pesticides, etc.

## Receptor fluid

One of the prerequisites for an appropriate receptor fluid is sufficient solubility of the test chemical. Furthermore, the receptor fluid should not alter the barrier properties of the skin membrane. For water soluble compounds, the most frequently used receptor fluid is saline solution. To improve solubility of lipophilic compounds, bovine serum albumine (BSA), organic solvents or surfactants can be added to the receptor fluid (Bronaugh et al., 1999; Cross et al., 2003; de Lange et al., 1994). However, organic solvents and surfactants can cause damage of the skin barrier (Sartorelli et al., 2000). In general, the choice of the receptor fluid has a more pronounced effect on the dermal absorption of lipophilic compounds due to their limited solubility in most receptor fluids. However, some studies show that the presence of materials such as BSA in receptor fluid can significantly influence dermal absorption of relatively hydrophilic compounds, probably due to protein binding of the test chemical (Wilkinson and Williams, 2002).

## Skin type and preparation

OECD and SCCNFP guidelines for in vitro testing (OECD, 2004b; SCCNFP, 2003, 2006) allow for the use of human and animal skin, whereas US-EPA (US-EPA, 2004) prescribes the use of human cadaver skin only. In studies using animal skin, pig skin is preferred because it resembles best the permeation properties of the human skin (OECD, 2004; SCCNFP, 2006). Rat skin can also be used, but caution should be taken in the interpretation of the permeation data, since it is known that rat skin is more permeable than human skin (van Ravenzwaay and Leibold, 2004a; Vecchia and Bunge, 2005). Since differences in the percutaneous absorption of
human and rat skin are highly variable and not determined by the physico-chemical properties of a chemical, it is not possible to derive an adjustment factor for human percutaneous absorption from the rat data (van Ravenzwaay and Leibold, 2004a; Vecchia and Bunge, 2002a). Use of artificial skin is not considered appropriate by both SCCNFP and OECD due to insufficient barrier function compared with that of living skin (SCCNFP, 2003, 2006; OECD, 2004b).

For the in vitro assay human abdominal or breast skin can be used according to the SCCNFP and OECD guidelines, whereas US-EPA prescribes the use of abdominal skin only. When using pig skin, dermal absorption testing is usually performed with flank or ear skin, although skin from the back or limb is also allowed. In the case of the rat, dorsal and ventral skin may be used.

Skin samples used can be full-thickness (<1000 $\mu \mathrm{m}$ ), split-thickness (200-400 $\mu \mathrm{m}$ ) and epidermal membrane skin preparations (OECD, 2004b). Full thickness membrane includes, viable epidermis and dermis, whereas in split thickness skin the lower dermis has been removed. Both, the OECD and the SCCNFP guidelines recommend the use of split thickness skin, whereas the use of full thickness skin must be justified. In the US-EPA test rules (US-EPA, 1999, 2004) only splitthickness skin of 200 to $500 \mu \mathrm{~m}$ can be used. In addition to full thickness and split thickness preparations, epidermal membranes comprising the viable epidermis and the stratum corneum may be used, but the reason for this choice should be justified (OECD 2004b; SCCNFP, 2003, 2006).

The integrity of the skin sample should be determined prior to the experiment by measuring the penetration of a marker substance (e.g. tritiated water, testosterone) or by measuring TEWL (trans-epidermal water loss) or TER (transcutaneous electrical resistance).

The type of skin preparation has been reported as a major contributor to the variation in results of in vitro testing (Ramsey et al., 1994; van de Sandt et al., 2004; Wilkinson and Williams, 2006). Several comparative studies show that absorption of lipophilic compounds into the receptor fluid is significantly reduced when using full thickness skin as compared to split thickness skin due to the diffusion resistance of the hydrophilic dermis for lipophylic compounds (Wilkinson and Williams, 2002; Yourick et al., 2004; Cnubben et al., 2002; van de Sandt et al., 2004). However, some studies have demonstrated that hydrophilic compounds such as glycol ethers and compounds that can bind to the skin can be retained in the skin revealing a more complex relationship between skin thickness, lipophilicity and dermal absorption (Wilkinson and Williams, 2005).

## Vehicle

A test chemical can be applied onto the skin as neat, diluted (ideally with water), or as a formulation (OECD, 2004b; US-EPA, 2004). For lipophilic compounds, use of isopropyl myristate has been proposed (US-EPA, 2004). The physico chemical properties of a vehicle and the solute determine the partitioning of a test chemical in the stratum corneum and thereby the dermal absorption in vivo as well as in vitro. Furthermore, a vehicle can alter the composition and structure of the skin which might lead to enhancement or suppression of dermal absorption (Davis et al, 2002). The effect of the vehicle on percutaneous absorption has been demonstrated for compounds of different classes of compounds such as dodecane and naphthalene (Baynes et al., 2001), PAH (Sartorelli et al., 1999, 2001a), terpenes (Cal, 2006), benzene (Blank and MacAuliffe, 1985), toluene (Boman and Maibach, 2000), lindane (Dick et al., 2004), benzoic acid, caffeine and testosterone (Bronough an Franz 1986), glycol ethers (Traynor et al., 2007), and agrochemicals (van der

Merwe and Riviere, 2005). Due to the pronounced effect of a vehicle on percutaneous absorption, a test chemical should be applied in the same vehicle/formulation as for "in use" conditions (OECD, 2004 a,b; SCCNFP 2003).

## Dose

The dosing regime in in vitro assays should be as close as possible to the real life exposure. The in vitro testing can be performed under finite and infinite dose conditions. In the infinite dose regime, the concentration of a test chemical does not significantly diminish during the dermal exposure duration. The finite dosing regime often better reflects the actual "in-use" scenarios, e.g. short skin contact due to splashes or exposure to volatile compounds which evaporate easily from the skin (OECD, 2004). When exposure duration is sufficiently long to attain steady state absorption, experiments using infinite dose enable the determination of the permeability coefficient, $\mathrm{K}_{\mathrm{p}}$.

## Data analysis

After finite dosing, dermal absorption can be expressed as an absolute amount absorbed ( $\mu \mathrm{g} / \mathrm{cm}^{2}$ of skin surface), as the percentage of the applied dose (relative absorption), or as maximum absorption rate (flux, $\mu \mathrm{g} / \mathrm{cm}^{2} / \mathrm{h}$ ) attained. To calculate the relative absorption, the amount of a test chemical which is washed from the skin should be measured as well as the amount present in the different skin layers and in the receptor fluid. There is still some scientific debate whether the fraction of compound remaining in the skin after washing (rinsing) and at the end of the experiment should be considered as systemically absorbed or not (OECD, 2004b; EHC, 2006; SCCNFP, 2003,2006; CEFIC, 2004). This is particularly relevant for very lipophilic compounds which have a tendency to remain in the stratum corneum, but also for other chemicals which may be retained in the skin due to, for example, binding to skin proteins. According to the OECD Test Guidelines, the test substance remaining in the skin should be considered as absorbed unless it is demonstrated that absorption can be determined from receptor fluid only (OECD, 2004b). However, when the test substance remains in the skin at the end of the study, it may need to be included in the total amount absorbed. The guidelines adopted by the European Cosmetic, Toiletry and Perfumery Association (COLIPA) and of the European Commission's Scientific Committee on Cosmetic Products and Non-Food Products intended for Consumers (SCCNFP, 2003; 2006) state that the amount of a chemical present in the stratum corneum at the end of the exposure should not be considered as systemically available. Both guidelines consider the amount of a chemical in the epidermis and dermis and in the receptor fluid as systemically available.

For infinite dose applications, the maximum absorption rate (maximum flux) can be determined and in this case the relative amount absorbed is not relevant. The maximum absorption rate can be deduced from the slope of the cumulative absorption-time profile in its linear part (Figure 2). In the case of steady state absorption, the permeability coefficient ( $\mathrm{K}_{\mathrm{p}}, \mathrm{cm} / \mathrm{h}$ ) can be calculated from the maximal absorption rate. There is usually an interval between the time of applying the dose and the steady state being reached. This "lag time" can also be derived from a graph of the cumulative absorption versus time, and is the intercept (on the time axis) of the tangent to the linear part of the absorption profile.

Figure 2 Cumulative absorption of a test chemical in time. Absorption (y axis) is expressed as the amount penetrated the receptor fluid in mass per unit exposure area)


The use of in vitro dermal absorption data in the risk assessment
In vitro studies to determine dermal absorption of chemicals are increasingly used for regulatory purposes. Furthermore, in vitro data are used to develop QSARs. Several important factors should be taken into consideration when planning dermal absorption studies in vitro. Among these are the choice of vehicle and receptor medium, the choice of skin membrane, and the dosing regime. The experimental conditions regarding dose, vehicle and exposure duration should be aligned with real-life exposure scenarios. On the other hand, the choice of skin membrane and receptor fluid will mainly be determined by the physico-chemical properties of a test chemical. This will be more critical for very lipophilic substances because of their low solubility in most receptor fluids. Within the EU, there is a Technical Guidance Document (EC, 2002) which provides generic guidance on risk characterisation undertaken within the New Substances, Existing Substances Regulation (ESR) and Biocides Product Directive (BPD) legislation. In this document, the problem of choosing a skin reservoir for very lipophilic compounds is indicated, such that there is a justification for the inclusion of material residing within the skin layers in the overall estimate of skin absorption for such substances. By including the amount retained in the skin as absorbed, results from in vitro studies seem to correlate with those from in vivo experiments and support their use as a replacement for in vivo testing (EC, 2003; EHC, 2006; Yourick et al., 2004).

For risk assessment purposes, dermal absorption studies with human skin are preferable. When only rat in vitro dermal absorption studies are available, the most conservative approach would be to assume that human skin absorption would be equal to rat dermal absorption.

### 2.1.3. Methods used for petroleum hydrocarbons

The dermal absorption of petroleum hydrocarbons has been investigated using a variety of methods, including in vivo studies in human volunteers, animal and in vitro
models. For the detection of dermal absorption of hydrocarbons in petroleum products there are several approaches. The first approach involves addition of one or more radioactive constituent hydrocarbons as markers in the product and monitoring the penetration of radioactivity through the skin. For hydrocarbons which penetrate the skin in sufficient amounts to be detected by conventional techniques, radiolabeling is not necessary. Another approach to determine dermal absorption of hydrocarbons is to apply the individual chemical neat or as a (aqueous) solution. Use of various experimental designs concerning type of skin (full thickness vs. epidermal membrane), vehicle, receptor fluid, animal species, dosing conditions (finite vs. infinite dose) makes comparison of the results of the studies extremely difficult. Petroleum hydrocarbons are lipophilic chemicals and as discussed above, it can be reasonably expected that experimental conditions will have a major influence on their absorption. The differences in experimental design are probably one of the reasons for the enormous discrepancies in dermal absorption of various petroleum hydrocarbons reported in literature (see Chapter 4). In general, based on comparative studies for various classes of chemicals it may be assumed that rat studies will give a conservative estimate of human dermal absorption and that the difference will probably be higher for more lipophilic hydrocarbons. For toluene and xylene, rat skin was shown to be up to 10 times more permeable than human skin. For other petroleum hydrocarbons no comparative studies have been reported.

As discussed above, the choice of a vehicle might have considerable influence on the dermal absorption of a chemical. Dermal application of a lipophilic hydrocarbon in water or in other hydrophilic solvent will favour its partitioning into the lipophilic stratum corneum hence increasing its dermal absorption. Dermal absorption of benzene, toluene, and xylene from aqueous solution was up to 3 orders of magnitude higher in comparison to the dermal absorption from gasoline or Jet fuel. So, it should be realised that dermal absorption from one vehicle cannot be assumed to be predictive for that from a different matrix. Furthermore it should be kept in mind that most QSAR models used for the prediction of dermal absorption are based on experimental data obtained with aqueous vehicles. In addition to thermodynamic considerations of a vehicle resulting in altered partitioning of a penetrant into the skin, some vehicles can damage the skin barrier leading to higher absorption. Repeated exposure to jet fuels has been demonstrated to cause increased absorption of some aromatic and aliphatic hydrocarbons.

One of the major factors affecting in vitro dermal absorption results is the choice of receptor fluid. Hence, for highly lipophilic petroleum hydrocarbons sufficient solubility of a tested hydrocarbon in receptor fluid should be provided. As discussed above, skin reservoir effects can play a significant role for lipophilic compounds. For this reason it is important that the amount of a penetrated hydrocarbon is determined in different skin layers and included in the calculation of dermal absorption.

In conclusion, experimental conditions, among which the choice of vehicle seems to be of particular importance, can influence the results of dermal absorption studies by several orders of magnitude. This emphasizes the need for using exposure and dosing regimes which are closely related to the workplace situation and to further study the dermal absorption of complex mixtures.

### 2.1.4. Summary of the experimental methods for assessment of dermal absorption

In the assessment of the dermal absorption properties of specific compounds, in either neat or diluted form or in mixtures or formulations, both in vivo and in vitro studies can be applied. The most reliable dermal absorption data for health risk assessment are obtained from studies in human volunteers, although for technical and ethical reasons their conduct is limited and closely regulated. In vivo methods which are based on measurement of the chemical (or radioactivity) in plasma and/or excreta have an additional practical importance for risk assessment of dermal exposure. The levels of chemicals and metabolites give an indication of the internal dose and are as such can be a better correlate for adverse health effects than the external exposure. Recent developments in applying techniques such as stratum corneum tape stripping, FTIR and Raman confocal spectroscopy offer the possibility to study in vivo dermal absorption in humans in a minimally invasive manner. However, these techniques are not suitable for all classes of chemicals. The tape stripping technique is not well-suited for volatile and rapidly penetrating compounds and still needs thorough standardization and validation. The main limitation of spectroscopic techniques is that the chemical has to have a specific spectrum.

One of the most appropriate animal models for the assessment of human skin absorption is the domestic pig both in vivo and in vitro. Traditionally, for the purpose of risk assessment, dermal absorption is most often determined using the in vivo rat model. However, the majority of data in the published literature show that rat skin is more permeable to chemicals than human skin. Thus the systemic exposure of humans may well be overestimated if the estimation of dermal absorption is based on the results of an in vitro or an in vivo rat study. Typically, rat skin shows 2-fold higher permeability than human skin, however, for some chemicals this difference has been shown to be significantly higher, being the highest for lipophilic compounds (in some studies up to 40 -fold). Due to inconsistent differences in absorption between rat and human skin it is not practical to simply derive a general adjustment factor for estimation of human uptake. However, when absorption data are available for rat in vivo and for rat and human skin in vitro, the in vivo human uptake may be accurately estimated. In vitro studies to determine dermal absorption of chemicals are increasingly used for risk assessment of pesticides, biocides, cosmetic ingredients and industrial chemicals in the EU. Although recently in vitro skin absorption test guidelines have been adopted, there are still a number of issues that are not precisely defined, raising the questions of the way in which in vitro data should be used for dermal risk assessment. Since experimental conditions regarding dose, vehicle and exposure duration are known to influence dermal absorption, the experimental design should be governed by real-life exposure scenarios. In general, the agreement between in vitro and in vivo is better for hydrophilic than for lipophilic compounds. Most studies show that the use of full thickness skin results in a lower absorption of lipophilic compounds into the receptor fluid when compared with the results obtained with split thickness skin, indicating a reservoir effect in the skin for these compounds. The reservoir effect is also affected by the choice of the receptor fluid. Again, this will have larger impact for lipophilic substances because of their low solubility in most commonly used receptor fluids. To compensate for the amount of chemical retained in the skin at the end of the exposure period, recent guidelines propose inclusion of the amount retained in the epidermis and dermis as being absorbed as a conservative assumption. On the other hand, the amount of chemical present in the stratum corneum at the end of the experiment is considered as non-absorbed in most guidelines.

### 2.2. MODELS FOR THE PREDICTION OF DERMAL ABSORPTION

## Introduction

The rather cumbersome procedures needed in the assessment of the dermal absorption properties of hazardous chemicals and the vast amount of chemicals that remain to be tested has resulted in several studies to find ways in which the dermal absorption can be predicted from the intrinsic properties of the compounds of interest and the conditions under which the dermal exposure takes place.

In this chapter a brief introduction and overview will be presented of the various mathematical approaches that have been developed and are used to assess and predict the degree of dermal absorption. The models are based on the experimental data available for a large number of diverse compounds and knowledge of the structure and properties of the skin barrier. Recently comprehensive reviews and reports have been published in which the various aspects of these approaches are discussed in more detail (WHO, 2006; Jones et al., 2004; Yamashita and Hashida, 2003). Also in the text the reader is referred to comprehensive reviews on the various subjects.

First of all, the main aspects playing a role in dermal absorption will be discussed briefly. A general overview of the various aspects playing a role in percutaneous absorption has been presented in ECETOC Monograph No. 20 (ECETOC, 1993). The theoretical aspects of skin permeability in relation to percutaneous absorption studies have been outlined in detail by Dugard (1987).

In the dermal absorption and uptake process, the stratum corneum layer is generally assumed to form the main barrier, but other layers like the viable epidermis may in some cases (especially for highly lipophilic compounds) also play an important role (see e.g. Cleek and Bunge, 1993; Bunge et al, 1995). The partitioning between the external exposure medium and the skin, and the diffusion process through the skin barrier are the main two physicochemical processes underlying dermal absorption.

The mathematical models developed to describe dermal absorption are explicitly or implicitly based on these physicochemical processes and furthermore on a more detailed description and understanding of the structure and properties of the skin. (See e.g Cleek and Bunge, 1993; Bunge et al., 1995; Frash and Barbero, 2003; Kasting 2001; Kasting and Miller, 2006; Anissimov and Roberts, 1999, 2001; Kruse et al., 2007).

The transport properties through the skin are usually described by the permeability coefficient $\mathrm{K}_{\mathrm{p}}$ which, according to Fick's first law, is defined as:
$K_{p}=J_{s s} /\left(C_{d}-C_{r}\right)$
where $J_{s s}$ is the steady state flux (i.e. the maximum flux after reaching equilibrium) through the skin and $\mathrm{C}_{\mathrm{d}}$ and $\mathrm{C}_{\mathrm{r}}$ are the concentrations of the permeant in the donor (exposure vehicle) and receptor (systemic circulation or receptor cell in case of in vitro studies), respectively. For a diffusion-based model in which the stratum corneum is assumed to represent the barrier for absorption, $\mathrm{K}_{\mathrm{p}}$ may be rewritten as:
$\mathrm{K}_{\mathrm{p}}=\mathrm{K}_{\mathrm{sc} / \mathrm{d}} \cdot \mathrm{D}_{\mathrm{sc}} / \mathrm{h}_{\mathrm{sc}}$
where $h_{s c}$ is the apparent thickness of the stratum corneum, $D_{s c}$ is the effective permeant diffusivity in the membrane and $\mathrm{K}_{\text {sc/d }}$ is the partition coefficient between
the stratum corneum and the vehicle. Usually, the concentration $\mathrm{C}_{\mathrm{r}}$ (i.e. in the systemic circulation or receptor cell) is assumed to be zero or at least negligible compared to $\mathrm{C}_{\mathrm{d}}$ (often referred to as the sink condition). Then the steady state flux $J_{s s}$ becomes:
$\mathrm{J}_{\mathrm{ss}}=\mathrm{K}_{\mathrm{p}} . \mathrm{C}_{\mathrm{d}}$
If the concentration in the vehicle $\left(\mathrm{C}_{\mathrm{d}}\right)$ does not decrease during exposure (so-called infinite dose) the permeability coefficient can be calculated from the donor concentration and the steady state flux. If, on the other hand, the concentration in the donor fluid decreases due to uptake in the skin (finite dose condition), the maximum (steady state) flux may only be reached for a limited time, and may not be maintained or even reached at all. This hampers a proper determination of the permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$.

In the first period of dermal exposure there is no equilibrium flux across the skin barrier (no steady state absorption). This non-steady state period is characterized by the lag time ( $\mathrm{t}_{\mathrm{lag}}$ ), which for a single homogeneous barrier (e.g. the stratum corneum) is given by:
$\mathrm{t}_{\mathrm{lag}}=\left(\mathrm{h}_{\mathrm{sc}}\right)^{2} /\left(6 . \mathrm{D}_{\mathrm{sc}}\right)$
The lag time can be also derived from the data generated in a dermal absorption study by determining the time intercept of the tangent to (or practically the linear portion of) the cumulative absorption time course (see Figure 3).

Figure 3 Cumulative absorption through a skin area (A). From the tangent to the absorption curve the (maximum) steady state flux $\left(\mathrm{J}_{\mathrm{ss}}\right)$, the permeability coefficient and the lag time are derived as indicated in the figure.


Use of a finite dose (see above), and also the fact that due to longer lag times the steady state has not (yet) been reached may in practice hamper a straight-forward interpretation of the experimental observation in terms of $\mathrm{K}_{\mathrm{p}}$.

In the dermal permeation models $\mathrm{K}_{\text {sc/d }}$ is usually related to the octanol-water partition coefficient, while $D_{s c}$ is assumed to depend predominantly on the size of the molecule.

For the modelling of dermal absorption, two main approaches can be distinguished: QSARs and mathematical models in which the partitioning and transport processes involved in the dermal absorption are described (Fitzpatrick et al., 2004).

In the QSAR, or more specifically QSPeR, approach relationships are established between the permeability properties (usually the permeability coefficient, $\mathrm{K}_{\mathrm{p}}$ ) and the physicochemical properties of the permeant. To derive these relationships databases containing experimentally determined $K_{p}$ values for molecules with a range of physicochemical properties (molecular size and lipophilicity) are used. (Vecchia and Bunge, 2002a; see also section 2.2 .5 below). The relationships can therefore be used to predict the permeability of untested compounds.

In a second approach, various mathematical models incorporating a more detailed description of the skin structure, the related barrier properties and the resulting transport processes are developed. This approach may result in further insight on the mechanism of skin permeation. The models can be used in the analysis, interpretation and extrapolation of results from dermal absorption experiments, such as time courses of permeation and amount of absorption.

In the following sections the two approaches indicated above (QSPeR and mechanistically based mathematical models) will be reviewed and discussed further. In addition, in the literature some special models for the prediction of dermal absorption are presented, which although related to the previously mentioned approaches, will be described in a separate section (see Section 1.2.4).

### 2.2.1. QSPeR methods

QSPeRs for dermal absorption are normally based on statistically determined correlations between the physicochemical properties of a range of permeants and solvents and their experimentally determined (steady state) permeability coefficients (Geinoz et al., 2004, Potts and Guy, 1992).

QSPeRs have been the subject of several comprehensive reviews (Geinoz et al, 2004; Moss et al., 2002; Walker et al., 2003; Vecchia and Bunge, 2002a,b). Many of the published QSPeRs are based on a set of permeability coefficients for 94 compounds, published by Flynn (1990).

Based on the Flynn dataset, Potts and Guy (1992) established a QSPeR using a combination of the octanol-water partition coefficient ( $\mathrm{K}_{\mathrm{ow}}$ ) and molecular weight (MW) -and thus indirectly the molecular volume - as physicochemical descriptors. It has the following form:
$\log \left(\mathrm{K}_{\mathrm{p}}\right)=0.71 \log \left(\mathrm{~K}_{\mathrm{ow}}\right)-0.061 \mathrm{MW}-2.44$ (with $\mathrm{K}_{\mathrm{p}}$ in units of $\mathrm{cm} / \mathrm{h}$ )
Following this study several molecular descriptors in combination with various permeability databases have been used in studies aimed at optimization of the

QSPeRs. A selection of these studies is given by Fitzpatrick et al. (2004). An example of this more complex approach is presented by Patel et al. (2002). They recently developed a QSPeR that contained, in addition to terms for hydrophobicity ( $\mathrm{K}_{\text {ow }}$ ) and molecular size (MW), two additional calculated descriptors, which were identified to be the most significant in a group of 169 physicochemical descriptors, calculated for compounds in the dataset. The resulting QSPeR has the following form:
$\log \left(K_{p}\right)=0.652 \log \left(K_{o w}\right)-0.0603$ MW - 6.23 ABSQon - 0.313 SsssCH - 2.30
with

ABSQon the sum of absolute charges on oxygen and nitrogen, and
SsssCH the sum of E-state indices
The various attempts to arrive at an optimal QSPeR may be severely hampered by the quality of the databases. In a recently published report, based on a Workshop organized by CEFIC, the importance of the quality of databases comprising measured and well defined dermal absorption data is emphasized, as a key first step in QSPeR development (Jones et al., 2004). The OECD test guidelines for in vitro and in vivo dermal absorption experiments (OECD, 2004 a,b,c) leave room for variation in the experimental conditions. For further development and improvement of QSPeRs an optimal data base containing uniformly obtained (in vitro) results is essential. Therefore it is important to establish and follow strict experimental protocols, as recently proposed in the CEFIC Workshop meeting (Jones et al., 2004).

Another problem with most available QSPeRs is that the resulting predicted permeability coefficients $\left(\mathrm{K}_{\mathrm{p}}\right)$ are not directly applicable in risk assessment, since the permeability coefficients are derived using studies that are (in theory) based on maximum flux (steady state) data obtained from infinite dose in vitro studies. However, realistic risk assessment scenarios usually correspond to finite dose conditions and often exposures for limited periods (see e.g. Jones et al., 2004).

A further limitation to the application of available QSPeRs lies in the fact that they are based on data obtained using aqueous solutions of single compounds. This hampers their application for compounds present in mixtures and other solvents and for neat (liquid) chemicals.

Recently some critical evaluations of QSPeRs were performed. Geinoz et al. (2004) suggested that these models are more reliable when their application is confined to well-defined chemical classes, and that their applicability is restricted by the limited range of polarity and size of the permeants on which they are based. Bouwman (2006) presented an evaluation of 14 QSARs for skin penetration. Many models showed poor external predictivity. Only the models of Patel et al. (2002) for $\mathrm{K}_{\mathrm{p}}$ and Magnusson et al. (2004a) for Jmax provided reasonable predictions for liquids tested neat. They concluded that the values obtained may not be truly applicable for risk assessment.

### 2.2.2. Mechanistic models

In mechanistically based mathematical models for dermal absorption the transport of a chemical through the skin is related to its structure, the physicochemical
properties of the permeant in relation to the various compartments (layers) in the skin, the vehicle used and, in the case of in vitro experiments, also the experimental conditions, such as vehicle volume, receptor volume/flow etc.

In the models that have been developed the skin is assumed to consist of one or more layers, usually the stratum corneum and the viable epidermis. The two main processes in the dermal absorption process are:

- the partitioning of the permeant between the vehicle / donor compartment, the various skin layers, and receptor compartment / systemic circulation
- the diffusion through the various skin layers

The processes are described by mathematical equations that are solved in various ways. Reviews of mechanistically based mathematical models in percutaneous absorption have been presented by Roberts et al. (1999), McCarley and Bunge (2001) and Roberts and Anissimov (2005).

Some characteristic examples of (recently) developed mathematical models for dermal absorption are described briefly as follows:

Bunge and coworkers (Cleek and Bunge, 1993; Bunge and Cleek, 1995; Bunge et al., 1995) developed a model which included consideration of the barrier effect of the viable epidermis for lipophilic compound. They also presented a strategy for estimating the compound specific parameters when using aqueous vehicles. The model was included in documents by the US-EPA (1992). US-EPA (2004) introduced a modification to this model that takes into account losses of lipophilic compounds, which are absorbed in the stratum corneum and lost via desquamation. Furthermore, it was used in estimating the dermal absorption by the DERMWIN module of the Episuite program (Dermwin, 2007).

Anissimov and Roberts developed a pair of models for analysing the effects of flow rate, receptor sampling rate, and viable epidermis resistance for a constant donor concentration (Anissimov and Roberts, 1999, 2000) and the effects of finite vehicle volume and solvent deposited (Anissimov and Roberts, 2001).

Recently a detailed dermal permeation model was published in which, in addition to finite/infinite dosing and non-steady state situations, the effect of repeated exposures and other occupationally relevant exposure scenarios could be simulated (Kruse et al., 2007). This model and a new implementation of the finite dose model, based on the publications of Anissimov and Roberts (1999, 2000, 2001), showed good agreement. The models were used for a detailed analysis and interpretation of in vitro dermal permeation data (Kruse et al., 2007).

Kasting developed models for the analysis of the kinetics of finite dose dermal absorption of the non-volatile compound vanillylnonanamide (Kasting, 2001) and for volatile compounds (Kasting and Miller, 2006).

Frasch and Barbero modelled the steady state flux by describing the stratum corneum barrier as a brick and mortar structure and as a more complex irregular structure. The results compared well with those obtained with a homogeneous stratum corneum layer (Frasch and Barbero, 2003).

For the analysis of dermal absorption data in human volunteers that were exposed to jet fuel and prediction of blood concentrations following exposure, Kim and
coworkers (Kim et al., 2006a) developed a dermatotoxicokinetic model, which included 5 compartments (surface, stratum corneum, viable epidermis, blood and storage). The rate constants in the model were determined by fitting to the data. In this approach no physicochemical parameters (diffusion and partition coefficients) were determined.

Other examples of mechanistic modelling are: Roy et al. (1996), van der Merwe et al. (2006) and Auton et al. (1994).

The mechanistic modelling approach may result in further insight in the mechanism of skin permeation. The models can be used in the analysis, interpretation and extrapolation of results of dermal absorption experiments, such as the time course of permeation and amount of absorption. Furthermore, using these models, the effects of finite and infinite dosing can be evaluated. They may also be helpful in the extrapolation of experimentally obtained results to other (occupationally relevant) exposure conditions and scenarios and in the optimization of experimental conditions.

### 2.2.3. Default values

A third approach is the use of default-values for percutaneous absorption, which are based on theoretical considerations that there should be an optimum in $\log \mathrm{K}_{\mathrm{ow}}$ and a maximum in molecular weight for facilitating percutaneous absorption. De Heer et al. (1999) proposed the following criteria to discriminate between chemicals with high and low dermal absorption:

1. $10 \%$ dermal absorption for chemicals with a molecular weight $>500 \mathrm{Da}$ and $\log \mathrm{K}_{\text {ow }}$ smaller than -1 or higher than 4, otherwise
2. $100 \%$ dermal absorption is used.

In the absence of (reliable) experimental dermal absorption data, this approach is used by the European regulatory authorities involved in risk assessment (EC, 2003, 2004). If the risk assessment outcome based on these default values indicates that the exposure level is acceptable/tolerable, no additional quantitative assessment of the dermal absorption is needed. The advantage of the use of default values is the practical applicability of this approach. On the other hand it is generally considered to result in very conservative (worst case) estimates of dermal absorption.

### 2.2.4. Specific approaches

In the predictive modelling of dermal absorption some alternative approaches can be distinguished that cannot directly or uniquely be categorized as QSPeR methods or mechanistic models and therefore are discussed in a seperate section.

## Maximum flux

The worst case dermal exposure risk for a chemical is defined by its maximum flux ( $\mathrm{J}_{\max }$ ) through the skin. Since the maximum dermal flux should occur at the solubility limit of the chemical, it can be estimated as the product of the solubility limit of the chemical in water and its permeability coefficient from water (Roberts et al., 2002). Maximum flux values can also be estimated by using QSAR relations. Magnusson et al. (2004a) showed that from analysis of a database containing $J_{\max }$ values of 278 compounds, MW turned out to be the main determinant for the dermal flux.
$\log J_{\max }=-4.52-0.0141$ MW (with $J_{\max }$ in units of $\mathrm{mol} /\left(\mathrm{cm}^{2} . \mathrm{h}\right)$
Maximum dermal flux QSARs apply to all vehicles that do not affect the skin integrity. Predictions of rates below the maximum flux can be made on the basis of concentrations expressed in terms of fractional solubilities of a given vehicle (Roberts et al., 2002).

## Neural networks

Degim et al. (2003) described the use of artificial neural network (ANN) modelling for the prediction of dermal absorption. Their work is based on the use of partial charge, $\log \mathrm{K}_{\text {ow }}$ and MW data as input factors into an ANN for the prediction of permeability coefficients. The ANN model successfully predicted experimentally determined permeability coefficients. No direct relationship was observed between descriptors and permeability, indicating the existence of a complex relationship between the structure of the permeant and its dermal permeation.

## Skinperm

In the Skinperm model developed by Wilschut and ten Berge (Wilschut et al., 1995) the skin barrier is considered to be formed by a protein layer in the stratum corneum, a lipid layer of the stratum corneum and an aqueous layer below the stratum corneum, through which the substance has to permeate into the capillary bed before entering the bloodstream. The two layers in the stratum corneum are assumed to form parallel routes in the permeation process. The overall permeation coefficient from aqueous solution is expressed mathematically as a complex of permeation coefficients representing the permeation through the sub-parts of the skin barrier. The values of the five regression coefficients in the model were derived from a dataset of 123 measured permeability coefficients for 99 different chemicals, and includes the Flynn data set (Wilschut et al., 1995). The model can be used to estimate the dermal uptake from both aqueaous solutions and air for various exposure durations. An actualized version of the model (Skinperm 3.3) is available (ten Berge, 2007).

### 2.2.5. Databases

In the development of QSPeRs the availability of good quality skin permeability data sets is essential. In this section a brief overview of some relevant data sets will be presented. One of the first skin permeability databases used in the development of several QSPeRs was published by Flynn (1990) and contained 97 permeability coefficients for 94 compounds.

For homologous or closely related series of molecules, several permeability data sets have been published:
n-Alkanols:

Ethyl ether, 2-butanone, 1-butanol, 2-ethoxyethanol, 2,3-butanediol:

Phenolic compounds:
Glycol ethers:
Aromatic amines:
Para-substituted phenols:
PAHs:

Scheuplein and Blank (1971, 1973); Flynn and Yalkowsky (1972); Wiechers (1989)

Blank et al. (1967)
Roberts et al. (1977)
Dugard et al. (1984)
Levillain et al. (1998)
Hinz et al. (1991)
Roy et al. (1998); van Rooij et al. (1995)

List in adapted form taken from WHO (2006)
In the context of the EDETOX project (EDETOX, 2004) a well documented database was set up, including in vivo and in vitro percutaneous absorption and distribution data (Soyei and Williams, 2004). The database is accessible via http://www.ncl.ac.uk/edetox.

Recently Vecchia and Bunge (2002a,c) presented detailed summary reviews of published data sets of skin permeation coefficients and stratum corneum/water partition coefficients.

In some recent publications the data sets that were used in the development of QSPeRs have been critically reinvestigated and reanalysed (see e.g. Moss et al., 2002; Patel et al., 2002; Fitzpatrick et al., 2004; and WHO, 2006 - chapter 10).

A maximum flux database for 278 compounds has recently been described and is available as supplemental material from the Journal of Investigative Dermatology (online) at: http://www.nature.com/jid/journal/v122/n4/suppinfo/5602287s1.html?url=/ jid/journal/v122/n4/full/5602287a.html. The criteria for inclusion of these compounds are included in the accompanying paper (Magnusson et al., 2004a).

### 2.2.6. Summary of the models for the prediction of dermal absorption

The results of a large number of human in vitro studies have been used to derive the chemical permeability coefficients that form the basis for the development of various QSPeRs. The most widely applied QSPeRs relate the $K_{p}$ to simple physicochemical characteristics for the compounds, such as the MW and the lipophilicity (logK ${ }_{\text {ow }}$ ). The QSPeRs are usually based on data obtained with aqueous vehicle solutions, and therefore these QSPeRs should in first instance only be used for an indication of the permeability properties from aqueous solutions. Furthermore, the permeability coefficients only relate the concentration in the vehicle on the skin to the dermal permeation rate for steady state and infinite dose conditions.

In practice the QSPeR approach provides predictions of limited accuracy. In part this may be due to the quality of the underlying databases. A more standardized approach to the performance of human in vitro studies, like that proposed by the OECD (OECD, 2004c), may help to reduce the variation in the key data used for the QSARs. However, in interlaboratory studies, it has proved difficult to remove all variations.

The permeation data on which the QSPeRs are based are mostly derived by using human skin samples obtained from abdomal skin and breast corrections. Because of the inherent variability in the barrier properties between different skin areas, one must be aware that the QSPeRs based predictions may either over- or underestimate the skin permeability for specific skin areas.

One of the specific problems in applying the available QSPeRs and programs for the prediction of $\mathrm{K}_{\mathrm{p}}$ and dermal absorption, such as Episuit-Dermwin and Skinperm that are implicitly using QSARs, is that the vast majority of the compounds on which the QSARs are based have $\mathrm{P}_{\text {ow }}$ values below about 4.4. For the numerous constituents in the petroleum products that have $\mathrm{P}_{\text {ow }}$ values greater than 4.4, the $\mathrm{K}_{\mathrm{p}}$ predictions may therefore be less reliable. In this respect it must also be noted that accurate determination of the $\mathrm{K}_{\mathrm{p}}$ is in practice very difficult to achieve for these type of compounds (high log $\mathrm{K}_{\mathrm{ow}}$ ) because infinite dose is difficult to achieve when using limited aqueous vehicle volumes (Kruse et al., 2007).

In general the QSPeRs and modelling programs based on these QSPeRs (Skinperm, Dermwin) predict an increase in $\mathrm{K}_{\mathrm{p}}$ with increasing $\mathrm{K}_{\mathrm{ow}}$ (at fixed MW).

This may result in predicted $K_{p}$ ranging from $1-10 \mathrm{~cm} / \mathrm{h}$ for $\log \mathrm{K}_{\mathrm{ow}}>6$. This is in sharp contrast with practical and experimental observation, which show that highly lipophilic compounds hardly permeate the skin (but stay in the lipophilic skin reservoir), and theoretical considerations on the hydrophilic barrier which the viable epidermis presents to lipophilic compounds (Cleek and Bunge, 1993). In the Dermwin program the dermally absorbed dose per unit area per event (DAevent) based on the approach of Cleek and Bunge (1993) may be used to obtain a more realistic idea of dermal absorption and the absorption rate per exposure event.

Even if the QSPeRs might be improved by using more recent and better data, they may only apply to one or a few situations (e.g. infinite dose and steady state) and complex extrapolation procedures should be performed to deal with actual exposure conditions.

In particular, application of the $\mathrm{K}_{\mathrm{p}}$ may be a problem in the case of relatively long dermal permeation lag times (e.g. in case of high MW molecules) and/or short (acute) exposure durations. In these cases the total penetration (amount absorbed into the skin) will be much larger than that based on the $\mathrm{K}_{\mathrm{p}}$ predictions, while the permeation rate (absorption rate through the skin into the system) will be much lower than predicted from the steady state permeability coefficient.

The expression of toxicity is a function of the dose at the target site. This dose may be the concentration ( $\mathrm{C}_{\max }$ at the target site) or the total amount delivered (for example. AUC at the target site) of the parent compound or a relevant metabolite. With respect to assessing the systemic risk of dermal permeation, two parameters may be distinguished: the permeation (absorption) rate, which will determine the $\mathrm{C}_{\text {max }}$ (and acute toxic effects), and the total amount systemically absorbed during a certain period (e.g. 24 hours).

This aspect may also be important when comparing the effects of acute and longer duration systemic exposures via inhalation and through the skin. Even assuming identical fractional uptakes via the two routes, the dermal uptake will generally be slower, but it may continue longer due to the lag time and the presence of a skin reservoir (especially for lipophilic compounds).

For investigating the various problems in the application of QSPeR results, mechanistically based mathematical models for dermal permeation may present a helpful tool, although one must be aware of the limitations of their use. Further, in the interpretation, extrapolation and application of experimentally obtained dermal absorption data, mechanistic models describing the dermal permeation process may prove to be useful, especially for proper retrieval of descriptive parameters such as lag time and permeability coefficient, extrapolation between infinite and finite doses, and the simulation of the permeation after occupationally relevant exposure scenarios (e.g. short duration, intermittent, repeated exposures) based on limited experimental data.

For neat solvents and possibly also for non-aqueous mixtures of poorly watersoluble, lipophilic compounds the $\mathrm{K}_{\mathrm{p}}$ 's derived from experimental flux data can probably not easily be compared with QSPeR predicted $\mathrm{K}_{\mathrm{p}}$ 's, which are based on data obtained with aqueous solutions, since the actual driving force for permeation is based on the activity and not on the concentration.

In that case thermodynamic arguments (activity, water solubility) may help to extrapolate from aqueous to non-aqueous exposure conditions. In this regard, it is the concept of maximum flux that defines the highest dermal exposure risk for a chemical, and QSARs based on this approach (Magnusson et al., 2004a) may offer a better approach to dermal risk assessment.

Since neat liquids and saturated solutions will have the same (maximum) thermodynamic activity, the dermal flux should be the same for neat liquids and saturated solutions, provided the vehicle or the neat chemical does not affect the barrier properties of the skin (CEFIC, 2004; Frasch, 2007). The maximum dermal flux of a chemical should thus occur at its solubility limit, and can therefore be estimated by multiplying its permeability coefficient obtained in water by its water solubilty (WHO, 2006). This approach may be especially useful for petroleum products and especially neat petroelum hydrocarbons.

In the case of mixtures, the maximum flux of each constituent may be considered as a very worst case approach, since we are not dealing with neat solutions but compounds that represent only a fraction of the solution mixture.

If the outcomes still pose a problem, the calculated (maximum flux) may be corrected by a factor representing the fractional amount of the compound in the solution. In this approach the difference in lipophilicity between the neat compound and the mixture (determining the partitioning of the compound into the skin) will be ignored.

It may be concluded that whilst the concept of maximum flux may seem to be rather attractive, in practice application of this will be quite complex, especially for the limited concentrations of individual compounds in (complex) mixtures.

A serious problem in the development and use of mathematical models is the effect of the permeant or the vehicle on the skin barrier properties. For solvents prolonged contact with the skin may result in penetration into the skin and thereby gradually change both the structure and the physicochemical properties of the skin. Contact with highly lipophilic solvents or permeants may even result in damage or destruction of the skin barrier. The time course of these effects is difficult to predict and to model in a quantitative way. Moreover the effects may be related to individual susceptibility. This will seriously hamper extrapolation of dermal permeation data between short and prolongued exposure.

In spite of all the shortcomings in the methods for estimating dermal absorption, the presently available models (QSPeRs) can help in the categorisation of compounds (solutes) as good, bad or intermediate skin permeants (Magnusson et al., 2004b; WHO, 2006).

In the absence of detailed exposure data and information on the skin permeation properties, a more generic and conservative approach, like that proposed by De Heer et al. (1999), may be followed to arrive at a worst case estimate of the 'safe' exposure level.

### 2.2.7. Comparison of experimental data on dermal absorption of petroleum hydrocarbons with model predictions

One of the main obstacles to wider application of predictive modelling for dermal absorption is their limited evaluation and validation with experimental data. In this section, the experimental data on percutaneous absorption of petroleum hydrocarbons presented in Chapter 4 are compared with the values predicted using the SKINPERM and DERMWIN models. The data on physico-chemical properties needed for modelling and the predicted values are shown in Tables I and II. As illustrated in Figure 4, the experimental $\mathrm{K}_{\mathrm{p}}$ is inversely proportional to the lipophilicity of the petroleum hydrocarbons (expressed as $\log \mathrm{K}_{\text {ow }}$ ), which is in contrast to the trend obtained from both prediction models. The predicted $\mathrm{K}_{\mathrm{p}}$ values were similar for both models and showed a significant overprediction when compared with experimentally derived $\mathrm{K}_{\mathrm{p}}$ values obtained in studies performed with petroleum products or neat solvents. There is a better agreement between predicted and experimental values for water solutions, which is to be expected since both models are based on studies with aqueous solutions (Figure 4, closed circles). However, for the maximum flux, which is calculated using $K_{p}$ values predicted by the DERMWIN model (Flux $\max =K_{p} x$ water solubility), a decrease in maximum flux with increasing $\mathrm{K}_{\text {ow }}$ is obtained. This is consistent with the trend obtained in experimental studies (Figure 5). Interestingly, six of the data points clearly deviate from other values. These points represent the values of aromatic compounds with 3 or more rings (anthracene, phenanthrene, fluorine, pyrene, chrysene and benzo[a]pyrene), which are characterised by very low water solubility.

In conclusion, this result shows the advantage of using maximum flux instead of $K_{p}$ in the risk assessment of highly lipophilic hydrocarbons.

Figure 4 Relationship between the $K_{p}$ and $\log K_{\text {ow }}$. The values represent the experimental data presented in Chapter 4 and the values predicted by the SKINPERM and DERMWIN models


Figure 5
Relationship between the maximum dermal flux and $\log \mathrm{K}_{\text {ow }}$ predicted by the DERMWIN model


Table I
Physico chemical properties of selected hydrocarbons used by the predictive models SKINPERM and DERMWIN

| Chemical | CAS number | $\begin{gathered} \text { MW } \\ \text { Da } \end{gathered}$ | Vapour pressure Pa $\left(25{ }^{\circ} \mathrm{C}\right)$ | Water solubility mg/l | $\operatorname{logK}_{\text {ow }}$ SKINPERM | $\log _{\text {ow }}$ DERMWIN |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Hexane | 110-54-3 | 86.20 | 20000 | 10 | 3.50 | 3.90 |
| Heptane | 142-82-5 | 100.20 | 6130 | 3.4 | 4.66 | 4.66 |
| Nonane | 111-84-2 |  |  |  |  | 5.65 |
| Decane | 124-18-5 |  |  |  |  | 5.01 |
| Undecane | 1120-21-4 | 156.31 | 55 | 0.257 | 5.74 | 5.74 |
| Dodecane | 112-40-3 | 170.34 | 18 | 0.0037 | 6.10 | 6.10 |
| Tridecane | 629-50-5 |  |  |  |  | 6.73 |
| Tetradecane | 629-59-4 |  |  |  |  | 7.20 |
| Hexadecane | 544-76-3 |  |  |  |  | 8.20 |
| Benzene | 71-43-2 | 78.00 | 12700 | 1780 | 2.13 | 2.13 |
| Trimethylbenzene | 25551-13-7 |  |  |  |  | 3.42 |
| Ethylbenzene | 100-41-4 | 106.17 | 1280 | 169 | 3.15 | 3.15 |
| Toluene | 108-88-3 | 92.14 | 3800 | 526 | 2.73 | 2.73 |
| xylene $_{\text {izomer mixture }}$ | 1330-20-7 | 106.00 | 1330 | 200 | 3.00 | 3.12 |
| Anthracene | 120-12-7 | 178.24 | 0.06600 | 0.04340 | 4.45 | 4.45 |
| benzo[a]pyrene | 50-32-8 | 252.00 | 0.00000073 | 0.00162 | 6.13 | 6.13 |
| Pyrene | 129-00-0 | 202.00 | 0.0008 | 0.135 | 4.88 | 4.88 |
| Chrysene | 218-01-9 |  |  |  |  | 5.81 |
| Fluorine | 86-73-7 |  |  |  |  | 4.18 |
| Phenanthrene | 85-01-8 |  |  |  |  | 4.46 |
| Naphthalene | 91-20-3 | 128.00 | 4.0 | 30 | 3.37 | 3.30 |
| methylnaphthalenes | 1321-94-4 |  |  |  |  | 3.87 |
| dimethylnaphthalenes | 28804-88-8 |  |  |  |  | 4.31 |
| trimethylnaphthalenes | 28652-77-9 |  |  |  |  | 4.81 |

Table II Permeability coefficient $\left(\mathrm{K}_{\mathrm{p}}\right)$ and maximum flux predicted by the SKINPERM and DERMWIN models

| Chemical | $K_{p}(\mathrm{~cm} / \mathrm{h})$ <br> SKINPERM | $K_{p}(\mathrm{~cm} / \mathrm{h})$ <br> DERMWIN | $\operatorname{logK}_{\text {ow }}$ SKINPERM | $\operatorname{logK}_{\text {ow }}$ DERMWIN | Maximum flux SKINPERM | $\begin{aligned} & \text { Maximum } \\ & \text { flux } \\ & \text { DERMWIN } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Hexane | 0.241 | 0.334 | 3.50 | 3.90 | 0.002410 | 0.00576 |
| Heptane | 1.37 | 0.949 | 4.66 | 4.66 | 0.004650 | 0.003370 |
| Nonane |  | 3.230 |  | 5.65 |  | 0.001480 |
| Decane |  | 0.931 |  | 5.01 |  | 0.001170 |
| Undecane | 3.49 | 2.540 | 5.74 | 5.74 | 0.000897 | 0.000653 |
| Dodecane | 5.13 | 3.730 | 6.10 | 6.10 | 0.000019 | 0.000410 |
| Tridecane |  | 8.520 |  | 6.73 |  | 0.000234 |
| Tetradecane |  | 15.200 |  | 7.20 |  | 0.000140 |
| Hexadecane |  | 52.500 |  | 8.20 |  | 0.000048 |
| Benzene |  | 0.0207 | 2.13 | 2.13 | 0.048800 | 0.041400 |
| Trimethylbenzene | 0.0274 | 0.0944 |  | 3.42 |  | 0.01130 |
| Ethylbenzene |  | 0.0739 | 3.15 | 3.15 | 0.016500 | 0.016900 |
| Toluene | 0.0978 | 0.0453 | 2.73 | 2.73 | 0.031600 | 0.026000 |
| Xylene ${ }_{\text {iisomer mixture }}$ | 0.0601 | 0.0704 | 3.00 | 3.12 | 0.0152 | 0.017100 |
| Anthracene | 0.106 | 0.225 |  | 4.45 |  | 0.000155 |
| Benzo[a]pyrene |  | 1.240 |  | 6.13 |  | 0.000013 |
| Pyrene | 0.281 | 0.324 |  | 4.88 |  | 0.000073 |
| Chrysene | 1.49 | 1.030 | 4.45 | 5.81 | 0.000012 | 0.000027 |
| Fluorine | 0.399 | 0.171 | 6.13 | 4.18 | 0.000002 | 0.000229 |
| Phenanthrene |  | 0.229 | 4.88 | 4.46 | 0.000054 | 0.000155 |
| Naphthalene |  | 0.0694 |  | 3.30 |  | 0.009860 |
| Methylnaphthalenes |  | 0.145 |  | 3.87 |  | 0.005890 |
| Dimethylnaphthalenes | 0.10 | 0.244 |  | 4.31 |  | 0.003620 |
| Trimethylnaphthalenes |  | 0.454 | 3.37 | 4.81 | 0.003010 | 0.002170 |

## 3. METHODS FOR THE ASSESSMENT OF DERMAL EXPOSURE

### 3.1. EXPERIMENTAL METHODS

### 3.1.1. Sampling methods

Contamination or exposure of the skin surface may occur by a number of mechanisms. Substances can be deposited on the skin directly from the air or due to contact of the skin with the contaminated surface. It may also arise from intentional or accidental immersion into the liquid or powder or due to spillage onto the skin during handling. Additionally, the processes such as evaporation of the contaminant from the skin or uptake of a contaminant through the skin make the evaluation and quantification of skin exposure even more complex (Schneider et al., 1999).

The methods for the assessment of dermal exposure can generally be divided into three groups: removal methods, collection (surrogate skin) methods and fluorescent tracer techniques. Despite the existence of several comprehensive guidelines (HSE, 1999; OECD, 1997; US-EPA, 1986; WHO, 1982) the main problem with application of the methods for measuring skin exposure is that they are not yet properly standardized nor validated due to (1) lack of consensus regarding the differences in the underlying assumptions for the various methods, (2) lack of emphasis on the development of a basic model of dermal exposure and (3) lack of consistent terminology (Schneider et al., 1999). This also hampers the comparison of results obtained using the various methods.

### 3.1.2. Removal methods

Removal methods are widely used and their advantage lies in the low cost and ease of application. In this technique, a substance deposited on the skin may be removed chemically (hand washing and rinsing) or mechanically (skin wiping, tape stripping) and subsequently determined by appropriate analytical method.

## Hand washing/rinsing and wiping

Hand washing and wiping are common methods for the assessment of dermal exposure because of their accessibility and ease of use. The methods have been reviewed by Brouwer et al. (2000a). The washing method is based on washing or rinsing the hands in a bag/bottle containing an appropriate solution of solvent or detergent in a routine washing procedure for a certain period of time at the end of exposure. The amount of solution should be adequate for sampling the entire hands. In this way the chemical is removed from the surface of the skin, with the solvent or detergent acting to enhance removal (Tucker et al., 2001; Brouwer et al., 1998; Fenske et al., 1999). The hand wiping method is based on removal of the contaminant from the surface of the skin (usually limited by a template) with wipes usually soaked in solvent, water or mixtures that are not damaging to the skin. While the hand washing method allows exposure of the whole hands to be estimated (expressed in units of mass per body part), hand wiping samples discrete/limited surface areas enabling expression of the mass removed from the skin in terms of mass per surface area e.g. $\mu \mathrm{g} / \mathrm{cm}^{2}$.

To assess the contamination level quantitatively, the sampling efficiency must be known. For that purpose two approaches have been used: mass balance and direct
spiking (Fenske and Lu, 1994; Fenske et al., 1998; Brouwer et al., 2000a) whereby the first approach mimics the transfer of a non-liquid contaminant and the latter the transfer of a liquid contaminant. The sampling efficiency, defined as the percentage of a contaminant removed from the skin, depends on several parameters such as time of residence, level of skin loading, method of contamination, skin condition, solvent type, and number of consecutive washings or wiping, etc. For example, time of residence of a substance on the skin is an important variable for contaminants that penetrate the skin rapidly or bind to skin components which can result in a reduced recovery and an underestimation of the actual exposure. This has been demonstrated for chlorpyrifos and captan where prolonged residence times resulted in decreased sampling efficiency (Fenske and Lu, 1994; Fenske et al., 1998).

Fenske et al. (1999) compared rates of exposure of the hand by hand washing and hand wiping sampling method. They observed on average a 6 fold difference between the two methods, with hand wiping being lower. Incomplete removal by hand wiping was also demonstrated by McCurdy et al. (1994) and Fogh et al. (1999).

## Tape stripping

Tape stripping is a non-invasive sampling technique used to assess dermal exposure. Generally, after the end of exposure, one or multiple layers of the uppermost layer of the skin, the stratum corneum, are removed by adhesive tape and analyzed for the chemical. The method has been applied to study the dermal exposure to multifunctional acrylates (Surakka et al., 1999; Nylander-French, 2000), metals (Cullander et al., 2000) and jet fuel (Mattorano et al., 2004; Chao et al., 2005). It was suggested that the washing technique could be a more accurate method than the tape stripping technique due to the variation caused by the operator performing the sampling (Roff et al., 2001; see also chapter in vivo methods). Mattorano et al. (2004) showed, in accordance with the results obtained by another technique (Fenske and Lu, 1994; Fenske et al., 1998), that the residence time in the skin dramatically influences the sampling efficiency for naphthalene. For jet fuel increasing the time of residence from 5 to 20 minutes resulted in a marked decrease from $69.8 \%$ to only $0.9 \%$. According to the authors this indicates that jet fuel rapidly penetrates through the stratum corneum.

### 3.1.3. Surface collection methods

Collection methods, also called surrogate skin techniques, refer to placing a collection medium on the skin surface or clothes for a certain period of time and its subsequent analysis for chemical content. The methods include patch sampling, whole body dosimetry and glove collection. The main assumption is that the collection medium/item captures and retains the chemicals in the same way as the skin but this is not necessarily the case since they do not have same properties as the skin. Another limitation of this technique is that it measures mass of the chemical deposited on the skin and not the concentration (e.g. $\mathrm{mg} / \mathrm{ml}$ ) which is the driving force for diffusion of the chemical through the skin (Cherrie and Robertson, 1995).

## Patch sampling and whole body dosimetry

Patch collection and whole body dosimetry techniques have been reviewed by Soutar et al. (2000). Both methods measure the amount of a contaminant on the skin surface expressed per unit area. Typically, the dimension of the patches is approximately $10 \mathrm{~cm} \times 10 \mathrm{~cm}$ although smaller patches have also been used. The area monitored represents around $3-8 \%$ of the body surface depending of the
number of patches used. The exposure is estimated by extrapolation of the patch area to the body surface. Another assumption which may lead to errors is that the patches represent the body surface contamination and hence that exposure is uniform over the whole body surface area. The various protocols use different standard body surface areas (OECD, 1997; US-EPA, 1986; WHO, 1982). The main drawback of the estimation of exposure is that extrapolation can lead to under- or over-estimation, errors being larger for smaller patches. Tannahill et al. (1996) observed differences of more than two orders of magnitude in a comparison of patches and oversuits. Whole body dosimetry, which measures the contaminant deposited on clothing overcomes this problem since the whole area is represented and no scaling is required. Additionally, it does not rely on an assumption that the distribution of the contaminant is uniform. However, a disadvantage is the need to use large volumes of extracting solvent resulting in only a small concentration of the contaminant which might present a problem for chemical analysis.

An additional problem is the use of different materials as a collection media, such as cotton, polypropylene, polyester, flannel and cellulose paper which hampers good comparison of studies. Care should be taken to prevent the saturation of the patches or garment. An assessment of the sampling efficiency should be performed prior to any field survey, as in case for removal techniques.

Sampling of volatile substances presents a problem in exposure measurement. In an attempt to overcome this problem, Cherrie and Robertson (1995) proposed the usage of a patch sampler that will collect the contaminant by diffusion. The sampler, comprising an adsorbent and semi-permeable membrane, should mimic the process of uptake through the skin thereby providing estimates of exposure that are "biologically relevant".

Vacuuming techniques, mostly used for sampling from contaminated surfaces other than skin, have been regarded as poor collection technique due to their low removal efficiency (Byrne, 2000; Glass et al., 2003). In a study by Lundgren et al. (2006) on the dust deposition of cornstarch and wheat flower on the skin three methods were compared: vacuuming sampling, tape stripping and patch sampling. They reported good agreement between vacuum sampling and tape stripping and slight overestimation for patch sampling compared to tape stripping. This was explained by particles remaining stuck on the glue.

## Glove collection

Absorbent gloves can be used to collect the contaminant by replacing or covering the protective gloves or placement under protective gloves. They are easily used, especially in field conditions, and they efficiently absorb the contaminant, which will otherwise be absorbed into the skin. Still, it is reported that gloves usually overestimate the exposure compared to, for example, hand washing and wiping (Fenske et al., 1999). Gloves should not become saturated, and to avoid this, it is recommended to regularly replace them (OECD, 1997). The method cannot be used for reliable estimation of whole body exposure due to the non-uniform distribution of contaminant on the skin and the large extrapolation factors needed.

### 3.1.4. Other procedures

## Fluorescent tracer method

The fluorescence tracer technique is based on the inherent property of a substance to emit visible light when illuminated with ultraviolet radiation. It has been used to
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assess surface deposition of naturally fluorescing substances, like polycyclic aromatic hydrocarbons (Ness, 1994). Since the procedure can easily locate the contamination, it enables the demonstration of non-uniform exposure, which contributes substantially to the uncertainty around the level of dermal exposure and finally risk assessment (Fenske et al., 1990). Combination of the fluorescent tracer technique with image processing (Video Imaging Technique to Assess Exposure, VITAE) provides a quantitative method which enables simultaneous assessment of the amount of a contaminant on the surface of skin and the area of exposed skin (Cherrie et al., 2000). The method has also been used to assess the efficiency of protective garments (Fenske et al., 2002). Semi-quantitative methods using a visual score system, which are based on visual observation and scoring of both the extent and intensity of the fluorescent tracer on the skin are also available (Fenske et al., 1988; Aragon et al., 2006).

Since the number of naturally fluorescing substances is relatively small, it is common to add a fluorescing substance as a tracer to follow the dispersion of a contaminant. The disadvantage of such an approach lies in the assumption that the added fluorescent tracer shows the same pattern of deposition, dispersion and retention characteristics as the contaminant. This method ignores the loss of contaminant from the surface of the skin by evaporation or by washing. Additionally, the fluorescent tracer can also bind to the skin which makes it less easily removed (Cherrie et al., 2000). The results of a fluorescence tracer study have been compared to those using chemical analysis; in the first study (analysis of coveralls) a large variation was reported between methods, while in the second study (washing of the skin) the correlation was good (Roff et al., 1997; Brouwer et al., 2000b). In both studies it was concluded that at low concentration exposure levels the fluorescence tracer is not as good as chemical analysis. Roff et al. (2001) also compared the fluorescence tracer method with the rinsing method and showed good agreement between these two completely different methods. In a very recent study, dermal exposure to semi-synthetic metal working fluids using a surrogate skin method (pads) and VITAE was measured (van Wendel de Joode et al., 2007). The study showed that dermal exposure levels estimated by VITAE were lower, by a factor of 3 , when compared to pads.

## Conceptual model

It has been suggested that a conceptual model proposed by Schneider et al. (1999) to assess dermal exposure could act as a starting point for the development of a sampling strategy and form the basis for improved comparability between studies (Vermeulen et al., 2000; Soutar et al., 2000). The model describes exposure as the result of mass transport between six identified compartments where the hazardous substance may reside (source, air, surface contaminant layer, outer and inner clothing contaminant layer and skin contaminant layer). The proposed assessment/measurement strategy is based on a tiered approach which follows the protocol presented in European Standard EN 689 (CEN, 1995). In a first step, all sources of potential exposure should be identified by preparation of lists of all chemical substances present in the workplace with their corresponding toxicological information. In the second tier an evaluation of workplace factors including tasks, work patterns and techniques, production processes, sources of direct contact, spilling, splashing and emission to air, safety precautions and procedures should take place. The third tier should consist of a structured approach to assess the exposure. If dermal uptake cannot be ruled out a basic survey should be made with the purpose of providing quantitative information about the distribution and level of dermal exposure (Schneider et al., 2000). The ideal basic survey would consist of a dermal uptake monitor with corresponding dermal occupational exposure limits (DOEL). Different approaches to develop DOEL values for occupational practice
have been suggested, however, no consensus has yet been reached concerning establishment of these values (Bos et al., 1998; Nielsen \& Grandjean, 2004).

### 3.1.5. Dermal exposure measurement of petroleum hydrocarbons

There are very few dermal exposure studies concerning petroleum hydrocarbons. Van Wendel de Joode et al. (2005) evaluated charcoal cloth pads for the assessment of dermal exposure to benzene and toluene in workers performing four different jobs in a petrochemical plant. Activated charcoal pads with an area of 4 cm x 3 cm were worn on the wrist of the hand of preference during a full-shift of 8 hours. The determined recoveries of charcoal pads ranged from 85 to $100 \%$ for both benzene and toluene. The average amount of benzene deposited on the pad was between 0.46 and $2.37 \mu \mathrm{~g} / \mathrm{cm}^{2} / 8 \mathrm{~h}$ and for toluene was between 0.69 and $11.99 \mu \mathrm{~g} / \mathrm{cm}^{2} / 8 \mathrm{~h}$. Mattorano et al. (2004) investigated dermal exposure to neat jet fuel using an adhesive tape stripping technique. The subjects ( $n=22$ ) were exposed on the ventral surface of each arm (exposure area $10 \mathrm{~cm}^{2}$ ) for $5,10,15$ and 20 minutes. After the end of exposure, adhesive tapes were applied on the exposed site. Three consecutive strips were taken. As a marker of exposure naphthalene was determined in the tape strips. On average the first tape strips taken 5 minutes after exposure removed $21.6 \mu \mathrm{~g} / \mathrm{cm}^{2}$ of naphthalene. The amount recovered with first tape strip decreased to $0.24 \mu \mathrm{~g} / \mathrm{cm}^{2}$ after 20 minutes of exposure. In the second tape strip the $0.15 \mu \mathrm{~g} / \mathrm{cm}^{2}$ of naphthalene was removed after 5 minutes decreasing to $0.022 \mu \mathrm{~g} / \mathrm{cm}^{2}$ after 20 minutes of exposure. The data indicate that naphthalene penetrates rapidly into the deeper layers of the stratum corneum. In the second study of the same group (Chao et al., 2005) dermal exposure of a USAF fuel-cell worker $(\mathrm{n}=124)$ to jet fuel performing different jobs was investigated. Exposure was assessed by using tape stripping technique and naphthalene was measured in tape strips as a marker of exposure. The subjects were monitored using several body regions with the greatest potential for exposure: forehead, neck, shoulders, arms, hands, legs, knees, feet and buttocks. Three successive tape strips were taken from each body region. Overall, dermal exposure to naphthalene was different among the sampled regions with the highest being on the buttocks of one subject. The authors calculated a whole body exposure to naphthalene of $2.02 \mu \mathrm{~g} / \mathrm{m}^{2}$. Significant differences were observed between the high-exposure group ( $4.19 \mu \mathrm{~g} / \mathrm{m}^{2}$ ) and the low-exposure group ( $0.34 \mu \mathrm{~g} / \mathrm{m}^{2}$ ), while there was no significant difference between the low-exposure and the medium-exposure $\left(0.48 ~ \mu \mathrm{~g} / \mathrm{m}^{2}\right)$ groups.

### 3.1.6. Summary of the experimental methods

Various methods have been developed and used for measuring dermal exposure. In essence, two main approaches can be distinguished: removal methods and collection methods. The various aspects of these approaches have been described and discussed in this chapter, and some specific examples for petroleum hydrocarbons have been presented. Additionally methods using fluorescent tracers were discussed and compared with other exposure assessment methods. The fluorescent tracer method has been used to demonstrate the non-uniformity of exposure, and to assess the efficiency of protective garments. The technique may be especially useful for petroleum products containing naturally fluorescing (polycyclic) aromatic hydrocarbons.

The methods and results may be applied in two ways.

1. For monitoring dermal exposure during specific occupational situations and/or when handling specific product types
2. To create a database of exposure data that can be used in creating and validating more generalized approaches (exposure models) for the assessment of dermal exposure.

The first application will especially be aimed at maintaining and/or improving measures for keeping dermal exposure at a level generally considered to be safe (occupational hygiene).

In the second application, exposure models can be developed for the assessment of dermal exposure under various conditions. Provided they are properly validated, these models may then be used reliably in risk assesssment procedures for dermal exposure (combined with assessment procedures for dermal absorption as appropriate).

In practice it turns out that current methods for measuring skin exposure are, for several reasons, not yet properly standardized nor validated (Schneider et al., 1999). This will hamper both comparison of results obtained by the various methods and an objective evaluation of results in the context of risk assessment.

For petroleum hydrocarbons, the scope of occupational activities in the petroleum industry and the size of the database needed to enable the development and validation of exposure models may turn out to be too limited at this moment. In particular, the specific physicochemical properties that must be taken into account when designing and applying techniques for measuring exposures. In this respect it is important to recognise that petroleum products are (generally) characterized by a high lipophilicity, are often of high volatility, and the presence of different hazardous compounds (both aliphatic and aromatic). Also the exposure data that are already available must be critically evaluated with respect to the specific properties of the petroleum products in relation to the measuring techniques and materials used. In this respect, the use of specific compounds (e.g. naphthalene) in complex mixtures as markers for the total exposure (or absorption) should be viewed very critically, because of the possible differences (e.g. volatility, lipophilicity) with other compounds in the complex hydrocarbon mixture.

Finally, it must be bourne in mind that dermal exposure data may eventually be used to assess the amount that is systemically absorbed via the skin. In this respect it will be important to know not only the total amount deposited on the skin, but also the physical state (solid / liquid, neat / solution, type of vehicle / solvent), its concentration in the matrix (solution), the skin area exposed, the volume per surface area, the duration of exposure, the rate of flux through the skin and also the (various) anatomical sites of exposure.

### 3.2. MODELS FOR THE PREDICTION OF DERMAL EXPOSURE

Models have been developed for the assessment of occupational exposure particularly in cases when no measurements or very few data are available. There are only a limited number of models existing for dermal exposure e.g. EASE (ECB, 1996), EUROPOEM (EUROPOEM, 1996), RISKOFDERM (EU RISKOFDERM project, $5^{\text {th }}$ framework program, project QLK4-CT-1999-01107), and DREAM (van Wendel-de Joode et al., 2003). Although the validity of these models has not been extensively studied (Marquart et al., 2003), in recent years some studies aiming at validating these models have been published (Hughson and Cherrie, 2001; van Wendel de Joode et al., 2005a,b).

### 3.2.1. Estimation and assessment of substance exposure (EASE)

EASE is a general model for prediction of occupational exposure applicable to a wide range of substances. It was originally developed by the UK Health and Safety Executive (HSE, 1996) for both dermal and inhalation exposure assessment of new chemicals prior to their introduction to the market. The model has been also used for purposes not originally envisaged: e.g. it is used to fill data gaps present in exposure measurement of existing substances and applied for regulatory risk assessment (Creely et al., 2005, Cherrie et al., 2003). At the time of writing this report, the EASE model 2.0 for dermal exposure was available and a prototype version of model 3.0 was under construction (Tickner et al., 2005).

The structure of this model is based on a series of basic criteria, adapted from an inhalation exposure model, representing the overall situation in the workplace: physical state, pattern of use and control, where the latter two parameters are given in a simple form, due to lack of reliable data on dermal exposure. A more detailed discussion of these basic criteria is available in a comprehensive report on EASE model 2.0 by Cherrie et al. (2003). The model predicts the potential exposure of hands and forearms only and is expressed as a mass per unit area exposed per day (e.g. $\mathrm{mg} \mathrm{cm}^{-2}$ day $^{-1}$ ), since it was assumed that that hands and forearms would be the most commonly exposed body parts and this represents a skin area of about $2000 \mathrm{~cm}^{2}$ (Tickner et al., 2005).

Although there were a number of published validation studies for the inhalation part of EASE (e.g. Bredendiek-Kämper et al., 2001; Hughson and Cherrie, 2005; Devillers et al., 1997; Hughson and Cherrie, 2000; Llewellyn et al., 2001, Mäkinen et al., 2002), there is only one validation study published for the dermal part (Hughson and Cherrie, 2001). Although this study seems less relevant in the context of this report because of the type of material investigated (zinc compounds), it is included here as a validation result of EASE for the dermal route. Hughson and Cherrie (2001) investigated the workplace dermal zinc exposure and compared it to the values obtained by the EASE model. Measurements were performed on subjects that were producing or working with zinc oxide, zinc powder and zinc dust in different workplaces. Dermal exposure was estimated by collecting dust from the different parts of the body with wet wipes (forearm, hands, forehead and chest) but only exposure data for hands and forearms were compared to values predicted by EASE. Furthermore, the tasks carried out in the factories were categorized in terms of the following three different EASE endpoints: non-dispersive use with intermittent direct handling, wide dispersive use with intermittent direct handling and wide dispersive use with extensive direct handling. The survey was divided into phases according to the sampling procedure: in the Phase 1 only the back of the hands were sampled and in Phase 2 both the palms and the back of the hands were sampled. The measured levels of zinc exposure for all categories and both phases ranged from 6 to $439 \mathrm{~m} \mathrm{~cm}^{-2}$. When the exposure was predicted by EASE it ranged from $100-15000 \mu \mathrm{~g} \mathrm{~cm}^{-2}$, which clearly overestimated the dermal exposure by a factor of about 50. The interesting point is that the EASE predictions increased in line with the average measured exposures, which is similar to the validations of EASE performed for inhalation exposure scenarios (Devillers et al., 1997; Cherrie and Hughson, 2005). There are no validation data that show an agreement of EASE predictions with experimental/measured data on dermal exposure for liquids and non-metallic solids (Cherrie et al., 2003). The main limitation of the EASE model is that it only gives estimates of dermal exposure for hands and forearms, disregarding the level of dermal exposure for other areas of the body. The model also ignores the influence of hand/forearm washing, evaporation of contaminant from the skin as well as use of protective equipment (Tickner et al., 2005).
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### 3.2.2. Dermal Exposure Assessment Method (DREAM)

DREAM is a model to evaluate occupational dermal exposure in a semi-quantitative manner (Wendel-de Joode et al., 2003) The aim was to provide an initial assessment of dermal exposure to liquids as well as solids, by estimating the potential and actual exposure, to give insight into the dermal exposure distribution over different body parts and finally to identify exposure routes. Additionally, by including job and task ranking, it provides information on potential measurement strategies and helps in determining what, where and who to measure (van Wendel de Joode et al., 2003). It is largely based on a conceptual model developed by Schneider et al. (1999).

A DREAM assessment consists of two parts: the inventory part addresses general information and possible dermal exposure determinants, as identified in the conceptual model of Schneider et al. (1999). It consists of six modules (company, department, agent, job, task and exposure) structured hierarchically in the form of a questionnaire to be filled by an occupational health professional. The evaluation part determines the potential dermal exposure (i.e. the predicted contamination of clothing and uncovered skin due to three different (consecutive) exposure routes emission, transfer and deposition) and actual dermal exposure (i.e. the exposure on skin) for nine body parts. It is organized in such a way that each estimate is determined by a set of underlying variables, from a total of 33 variables. According to the authors, the advantage of DREAM lies in its ability to document the decisions made by the investigator in a structured way (van Wendel-de Joode et al., 2003). In order to assess the reliability of the model the authors of DREAM conducted a series of studies on dermal exposure where inter-observer agreement, the effect of the individual observer on dermal exposure estimates for different tasks and comparison of inter-observer agreement for ranking of body parts according to their exposure level were assessed (van Wendel-de Joode et al., 2005a). The study showed good reproducibiliy for inter-observer agreement, covering a broad range of tasks comprising exposure to liquids, solids and vapours. Furthermore, to assess the accuracy of their method they compared DREAM estimates with quantitative dermal exposure measurements in six occupational settings (van Wendel-de Joode et al., 2005 b ). The results showed that the accuracy of DREAM estimates varied within and between surveys. The limitation of the study lies in the fact that it was performed only for liquids. Another limitation is the inability to verify the accurate assessment of spatial variability. This was due to the lack of sufficient measurements on body parts other than hands.

### 3.2.3. RISKOFDERM

RISKOFDERM is an EU funded project which aims at (1) developing a toolkit for the assessment of dermal exposure in small and medium-sized enterprises and (2) development of dermal exposure models for regulatory risk assessment (Warren et al., 2003, 2006, van Hemmen, 2004; van Hemmen et al., 2003; Marquart et al., 2006). A series of papers have been published presenting the main framework of the toolkit (Goede et al., 2003; Marquart et al., 2003; Oppl et al., 2003; Warren et al., 2003). The toolkit is based on theoretical approaches for the following determinants: processes and tasks, substance and product characteristics, and situations and conditions. It assumes that operator exposure is influenced mainly by the tasks (Warren et al., 2003) for which six categories or DEO (dermal exposure operation) units are distinguished. Each of these categories has several modifiers divided into three groups (substance-specific modifiers, workplace-related modifiers and control measures) and to each group separate scaling factors are assigned
representing three different mechanisms of dermal exposure (direct contact, surface contact and deposition from an aerosol); this is explained in detail in the conceptual model by Schneider et al. (1999). This division of groups will lead to derivation of default exposure values for each task, which can be revised as new quantitative data become available on dermal exposure (Marquart et al., 2006). Development of dermal exposure models for regulatory risk assessment applies two approaches; a first is derivation of default values for each of six designated DEO-s (Marquart et al., 2006) and the second is a model based on statistical relationships between measured data and potential exposure determinants for each of the six DEO-s (Warren et al., 2006). Although, default values derived for potential dermal exposure of the hands are considered useful for similar scenarios and substances used, the uncertainty in the assessment using those values cannot be quantified (Marquart et al., 2006), The aim of the project was to design a model for each of the dermal exposure situations, but this is not yet possible due to insufficient reliable and precise data (Warren et al., 2006).

### 3.2.4. European predictive operator exposure model (EUROPOEM)

EUROPOEM is a harmonized, predictive model for operator exposure during application, based on the results of field studies. It is used only in the registration process for agricultural pesticides in the European Union. The predictions obtained from this model are compared to acceptable operator exposure level (AOEL, derived from relevant toxicological data) to assess whether the recommended application technique for the proposed pesticide is safe. When the ratio of exposure and AOEL is less then 1 the exposure scenario is considered acceptable. In the case of exposure and AOEL ratios exceeding 1, according to the tiered approach, this will lead to a more-detailed exposure assessment. The database was developed using only those studies performed according to the OECD guidance document (1997) unless they have been done before the Guidance document came into effect. Those data were also evaluated in order to have a representative set of data in the database. The exposure data can be calculated and presented either as the amount of active substance handled, or as the amount of formulation or spray volume per unit of time (van Hemmen, 2001).

### 3.2.5. Summary of models for prediction of dermal exposure and their applicability to petroleum products

In this chapter various models for the prediction of dermal exposure have been described and discussed namely EASE, DREAM, RISKOFDERM and EUROPOEM.

Although many efforts have been made in the development and optimization of these models, their specific application for the estimation of dermal exposure to petroleum hydrocarbons seems to be limited.

Schneider et al. (1999) proposed a conceptual model, that has been used in the development of the DREAM and RISKOFDERM models. At this time the practical application of these models for petroleum products in occupational settings has not been widely practiced. Further development and validation directed to these types of products and industries is needed.

Although EASE is a widely applied model for the prediction of occupational exposure, application of this model for dermal exposure asssessment of petroleum products is severely hampered by the lack of relevant validation data. There is only one validation study available for the dermal exposure part, but this concerns zinc
compounds (zinc oxide, zinc powder and zinc dust), which are not considered to be directly relevant to petroleum products. A further draw back is the fact that the EASE model provides only estimates of dermal exposure for hands and forearms and does not take into consideration the effects of hand/forearm washing, evaporation from the skin and the use of protective equipment. In particular evaporation from the skin may be an important factor for certain (volatile) compounds in petroleum products.

EUROPOEM is a data-based model for the prediction of operator exposure during mixing and application of agricultural pesticides. Therefore it can be concluded that application in petroleum industries is unlikely to be useful and hence it is not considered further.

Overall it is concluded that the available models for prediction of dermal exposure are not (yet) useful for providing a reasonably accurate dermal exposure prediction for pretroleum products. Therefore one should either use a worst case approach or perform actual occupational exposure studies and/or invest in further development and validation of the models presented above.

## 4. LITERATURE STUDIES ON DERMAL ABSORPTION OF PETROLEUM HYDROCARBONS

### 4.1. INTRODUCTION

The purpose of this brief review is to summarize relevant data on the dermal absorption of petroleum hydrocarbon products and their constituents as well as a critical evaluation of the available data. The data taken into consideration cover various aliphatic and aromatic hydrocarbons present in petroleum products that may come in contact with skin, either as individual chemicals or as part of the mixtures and commercial products.

### 4.2. PROCEDURE

To achieve the objective the following four steps were performed.

## Step 1:

Identification of relevant petroleum hydrocarbons and available literature data.

| Aliphatic hydrocarbons | Aromatic hydrocarbons |
| :--- | :--- |
| - Hexane | - Benzene |
| - Heptane | - Toluene |
| - Nonane | - Xylene |
| - Decane | - Methylbenzene, ethylbenzene, |
| - Undecane | trimethylbenzene |
| - Dodecane | - Naphthalene |
| - Tridecane | - Methylnaphthalene, dimethylnaphthalene, |
| - Tetradecane | trimethylnaphthalene, tetramethylnaphthalene |
|  | - Pyrene and benzo[a]pyrene |

## Step 2:

A searching of scientific bibliographic databases to identify publications that were not covered in a previous CONCAWE report on dermal absorption of petroleum hydrocarbons (Petroleum hydrocarbons: their absorption through and effects on the skin, CONCAWE, 1984).

Databases and web pages searched were:
www.pubmed.com
www.scholargoogle.com
www.google.com
Keywords used for the search of databases were:
"Name of the pure chemical" + percutanoeus absorption
"Name of the pure chemical" + percutaneous penetration
"Name of the pure chemical" + dermal absorption
"Name of the pure chemical" + dermal penetration
"Name of the pure chemical" + dermal permeation
"Name of the pure chemical" + skin absorption
"Name of the pure chemical" + skin penetration
"Name of the pure chemical" + skin permeation
Additionally, instead of "Name of the pure chemical" the following keywords were used: petroleum, petroleum hydrocarbons, kerosene, jet fuel, coal tar (although not a petroleum product), gasoline, aliphatic hydrocarbons, aromatic hydrocarbons and PAHs or polycyclic aromatic hydrocarbons.

The search was also performed by replacing "Name of the pure chemical" with CAS numbers of pure chemicals.

## Step 3:

A critical evaluation of the available data.
Questions addressed were:
i. Is the exposure performed to pure chemicals or to a mixtures or commercial products?
ii. Are the data and the manner in which they were obtained scientifically valid?
iii. For better comparison all the parameters were transformed into uniform units (i.e. $\mathrm{K}_{\mathrm{p}}$ in $\mathrm{cm} / \mathrm{h}$, flux in $\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}$ ).

## Step 4:

The analyzed data were summarized in standardized tabular formats (per study and per compound) to enable an effective use of the available information.

### 4.3. OVERVIEW OF THE STUDIES ON DERMAL ABSORPTION OF HYDROCARBONS

In the following tables, an overview has been given of the experimental studies on dermal absorption of petroleum hydrocarbons.

In Tables 1-20 listed in Appendix 1 of this report summary data on dermal absorption parameters for individual hydrocarbons have been presented.

In a second set of tables (Tables 1-49 listed in Appendix 2 of this report) a more detailed description of experimental conditions and interpretation of the outcomes is presented.

### 4.4. SUMMARY OF LITERATURE STUDIES ON DERMAL ABSORPTION OF PETROLEUM HYDROCARBONS

- There is substantial evidence from both human and animal in vitro as well as in vivo studies that aliphatic and aromatic compounds reviewed in this report are able to penetrate and permeate the skin when applied as pure substances, in mixtures or as commercial products.
- Generally, higher absorption of aromatics (e.g. toluene, xylene, naphthalene) than aliphatics (decane, dodecane, tridecane and hexadecane) has been reported.
- Aliphatic hydrocarbons show higher tendency to remain in the skin which might have influence on local effects such as irritation.
- Presented data on dermal absorption of individual hydrocarbons from petroleum products such as jet fuel, strongly suggests that dermal exposure to these products would not cause systemic toxicity under normal working conditions and assuming an intact skin barrier.
- Only 6 of 49 dermal absorption studies described in this report were done with chemical vapours and these were all done in vivo. These studies revealed the limited significance of dermal exposure to vapours of hexane, benzene, toluene, and xylene. The absorption of vaporous hydrocarbons is negligible in comparison to that of liquid hydrocarbons. Even when whole-body dermal exposure is considered, dermal uptake contributes only a few percent of the total (inhalatory + dermal) body burden.
- For some hydrocarbons, including benzene, toluene and xylene, the absorption from aqueous solutions was several orders of magnitude higher that the absorption after dermal exposure to either a neat chemical or a petroleum product e.g. jet fuel. This emphasises the need for conducting absorption studies under 'in use" scenarios. In the application of QSPeRs ( $\mathrm{K}_{\mathrm{p}}$ ) it should be kept in mind that for their development permeability coefficients are used that are obtained with aqueous vehicles.
- A few comparative studies showed that rat skin is more permeable than human skin. For toluene and xylene this difference was roughly 10 -fold. Pig skin permeability was similar to that of human skin.
- Almost all experimental studies were performed in "normal" untreated skin and no data were found for compromised skin. In two studies skin was used that was treated with baby oil, moisturizer, insect repellant and sunscreen cream or jet fuel before exposing to the compounds.
- Most of the in vitro studies were performed with split thickness skin.
- Pre-exposure to jet fuel led to higher (2- to 4- fold) absorption of the aromatic and some aliphatic hydrocarbons implying alteration of the skin barrier probably due to lipid extraction from the stratum corneum.
- Dermal absorption can significantly be affected by viscosity of a petroleum product. In general, higher viscosity leads to reduce dermal absorption.

Some gaps in our knowledge were identified:

- The lack of dermal absorption studies in compromised skin.
- The need for information on the effect of repeated exposure on the dermal absorption of petroleum hydrocarbons.
- The effect of the vehicle composition (mixture, aqueous neat liquid) and additives on dermal absorption.
- The effect of the concentration of individual compounds in complex petroleum product on their permeability (and flux).


### 4.5. USE OF DERMAL ABSORPTION STUDIES IN RISK ASSESSMENT: EXAMPLE OF BENZENE

To illustrate the way in which dermal absorption data can be used for risk assessment the contribution of dermal uptake to the total uptake (inhalatory + dermal uptake) has been estimated for benzene. Benzene has been taken as an example for several reasons. First, unlike the other key constituents, benzene is classified as a human carcinogen.. Secondly, dermal flux of benzene is one of the highest among petroleum hydrocarbons and reliable experimental data on dermal absorption of benzene is available in the literature for both aqueous solutions and petroleum products. Furthermore, for benzene there is an occupational limit value for indoor air concentration enabling comparison of inhalatory and dermal uptake routes in an occupational context.

Dermal intake of benzene resulting from skin exposure to a petroleum product (gasoline) was estimated using experimental results based on the study of Adami et al. (2006).

The experimentally determined skin flux amounted to $1.99 \mu \mathrm{~g} \mathrm{~cm}{ }^{-2} \mathrm{hr}$ ( $\mathrm{K}_{\mathrm{p}} 43.8 \mathrm{~cm} \mathrm{hr} 10^{-5}$ )

The following exposure scenario has been assumed:
The concentration of benzene in gasoline is: $0.7 \%$ (average of 3 gasoline samples).
Duration of skin exposure is: 60 minutes (cumulative exposure over an 8 -hour shift).
The exposed skin area: $1000 \mathrm{~cm}^{2}$.
Air concentration: $3.2 \mathrm{mg} \mathrm{m}^{-3}$ corresponding to the occupational exposure limit value in most countries at the time of writing this report. Furthermore, this value is expected to represent a "reasonable worst case" in the petroleum industry.

Assuming a respiration rate of $0.6 \mathrm{~m}^{3} \mathrm{hr}$, the estimated inhalatory uptake of benzene would be:
$0.6 \mathrm{~m}^{3} \mathrm{hr} \times 8 \mathrm{~h} \times 3.2 \mathrm{mg} \mathrm{m}^{-3}=15.4 \mathrm{mg}$
Therefore dermal uptake based on experimental data would be:

$$
1.99 \mathrm{~g} \mathrm{~cm}^{-2} \mathrm{hr} \times 1000 \mathrm{~cm}^{2} \times 1 \mathrm{~h}=1.99 \mathrm{mg}
$$

Dermal uptake $/($ dermal + inhalatory uptake $)=1.99 / 1.99+15.4=11.4 \%$
The comparison of dermal and inhalatory uptake shows that dermal exposure to petroleum products can lead to uptake of benzene. However, it has to be emphasized that the exposure scenario used in this example was rather unrealistic that is a 1-hr exposure over a skin area of $1000 \mathrm{~cm}^{2}$ does not occur under normal working conditions. However, this exposure scenario has been proposed by ECETOC for assignement of the skin notation (ECETOC, 1993).

## 5. METHODS FOR THE ASSESSMENT OF LOCAL SKIN EFFECTS

Skin exposure to petroleum products can lead to a variety of skin reactions varying in intensity from skin dryness and skin irritation to irreversible skin changes such as skin tumours (Koschier, 1999, Nessel et al., 1999a,b; Ritchie et al., 2003; Jia et al., 2002). Also individual aliphatic and aromatic hydrocarbons are known to be able to induce skin irritation; several of these compounds were positive in skin irritation testing and are classified as skin irritants.

This section addresses the mechanisms of skin local effects including irritation, corrosion and sensitization and the methods for the determination of these local effects

### 5.1. LOCAL SKIN EFFECTS

### 5.1.1. Skin irritation

Skin irritation encompasses a broad range of sensory and visible effects including dryness, fissuring, erythema, and oedema, which occur as a result of local inflammatory processes following single or repeated contact of the skin with chemicals (Maibach and Coenraads, 1995; Weltfriend et al., 1996). Acute irritant contact dermatitis is characterized predominantly by inflammation, while chronic irritant contact dermatitis is characterized predominantly by hyperproliferation and transient hyperkeratosis (Corsini and Galli, 2000). Skin irritation is a complex phenomenon that involves resident epidermal cells, dermal fibroblasts, and endothelial cells as well as invading leukocytes, particularly T-lymphocytes, interacting with each other under the control of a network of cytokines, neuropeptides, and eicosanoids. Keratinocytes presumably play an important role in the pathophysiology via generation of signals leading to attraction of leukocytes (Fuchs et al., 2001). The underlying pathophysiological mechanism by which a chemical induces an inflammatory response is not fully understood; however the fact that chemicals of different structure and physico chemical properties can cause skin inflammation implies that different pathways are involved. One of mechanisms by which a chemical can cause skin irritation is damage to a skin barrier for example by extraction of the lipids from the Stratum corneum by lipophilic solvents such as aliphatic and aromatic hydrocarbons. Increased water loss due to skin barrier perturbation can initiate release of pro-inflammatory cytokine IL-1a, IL-1 $\beta$ and tumour necrosis factor (TNF) alpha which induces secondary molecular responses and an inflammatory cascade (Wood et al., 1992; Nickoloff and Naidu, 1994; Corsini and Galli, 1998; Welss et al., 2004). The damage to the skin barrier might further facilitate the penetration of larger amounts of skin irritants and other chemicals which normally would not be able to penetrate normal intact skin. Chemicals which are able to penetrate the stratum corneum may elicit toxicological effect by damaging viable epidermal cells. In response to damage keratinocytes release IL1a, which essentially is a primary event in skin defence. IL-1a stimulates keratinocytes and fibroblasts to produce and release more IL-1 $\alpha$ and other proinflammatory cytokines and chemokines such as IL-1ß, IL-6, and tumor necrosis factor (TNFa) as the first step in the inflammatory cascade. These cytokines and chemokines in turn induse production of a wide array of other inflammatory meadiators and adhesion molecules resulting in the recruitment and proliferation of leukocytes at the site of inflammation (Nickoloff and Naidu, 1994; Corsini and Galli, 2000; Welss et al., 2004; Homey et al., 2006).
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Furthermore, contact irritants can induce skin inflammation by generation of reactive oxidative species in the skin which might activate transcription factors inducing synthesis of proinflammatory cytokines, dysregulate redox-sensitive signal transduction pathways, and trigger cytotoxicity and apoptosis (Corsini and Galli 1998; Fuchs 2001; Allen and Tresini, 2000; Coleman, 2003; Welss et al., 2004).

Acute inflammation is a rapid self-limiting process; however, by repeated exposures it may be maintained for a prolonged time and/or become chronic (Lisby, 2006). In addition to development of chronic contact dermatitis, repeated irritation can lie at the basis of the progression-promoting effect in the development of skin tumours. There is increasing evidence that severe, long-term dermal irritation, necrosis and degeneration induced by petroleum fuels maybe integrally related to possible tumorgenesis (Ingram and Grasso, 1991; Freeman et al., 1993; Walborg et al., 1998; Nessel et al., 1998, 1999a,b; Nessel, 1999). Although mechanisms for this effect are not yet clear, it has been hypothesized that epidermal hyperplasia with repeated episodes of damage and regeneration can initiate tumor induction due to oncogene activation possibly stimulated by release of oxidative enzymes from inflammatory cells.

### 5.1.2. Skin corrosion

As defined by the United Nations (UN) Globally Harmonised System for the Classification and Labelling of Chemical Substances and Mixtures (GHS) skin corrosion refers to the production of irreversible damage to the skin, manifested as visible necrosis through the epidermis and into the dermis, following the application of a test material.

### 5.1.3. Skin sensitization

Dermal exposure to contact allergens induces highly regulated cellular and molecular events that result in the stimulation of cutaneous immune responses and the acquisition of skin sensitization. In contact allergic reaction, a distinction should be made between induction (sensitization) and effector (elicitation) phases (Rustemeyer, 2006). The induction phase includes the events following a first contact with the allergen and is complete when the individual is sensitised and capable of giving a local immune response. The effector phase begins upon elicitation (challenge) and results in a clinical manifestation (Rustemeyer, 2006). Most contact allergens are small, chemically reactive molecules with a molecular weight less than 500 Da (Bos and Meinardi, 2000). Upon penetration through the epidermis, a chemical may react with protein or peptide, and then subsequently be processed by epidermal Langerhans cells (LC), a step which is critical for development of sensitization. Allergen-carrying LC become activated and travel to the draining lymph nodes: the site of proliferation for specific T-cells. Migration and maturation of LC following skin sensitization is mediated by release of locally available cytokines and chemokines such as IL-1 $\beta$, TNF- $\alpha$ and IL-18 (Ryan et al., 2007a). Sensitizing potential is dependent on several factors including the molecule's capacity to penetrate the Stratum corneum, lipophilicity and chemical reactivity. Two other factors which further contribute to the allerginicity are proinflammatory activity and capacity to induce maturation of Langerhans cells. There are clearly dose-response relationships for both induction and elicitation phases which are dependent on the potency of the chemical allergen but also on external factors such as influence of vehicle or formulation. (Ryan et al., 2007 b)

### 5.2. METHODS FOR THE ASSESSMENT OF LOCAL SKIN EFFECTS

### 5.2.1 Test methods for skin irritation and skin corrosion

The assessment of the irritancy/corrosion potential of chemicals is important when establishing the hazard classification of industrial chemicals and for safetyassessment purposes. Various in vivo and in vitro tests which will be addressed in this section are used to determine the irritation and corrosion hazard resulting from the contact of skin with chemicals. Due to complex mechanisms and different pathways of skin irritation, different endpoints are often needed to evaluate the irritation potential of chemicals

At present, validated in vivo and vitro tests for irritation and corrosion tests are available, and some of them are accepted by regulatory bodies (OECD TG 404, 1999, 2002; US-EPA 1996). The OECD guideline 404 (OECD, 1992, 2002) covers the assessment both of skin corrosivity (classified as R34 or R35) and skin irritation (classified as R38) based on in vivo rabbit skin tests. The criteria for R35 (causes severe burns) indicate that the substance will cause full thickness skin destruction within 3 min ; R34 (causes burns) indicates that full thickness skin destruction follows an exposure of up to 4 h . For R38 (irritating to skin), a substance will cause significant inflammation of the skin, which persists for at least 24 h after a semioccluded exposure of up to 4 h .

The OECD has indicated that in vitro toxicity tests can be accepted for regulatory purposes only after a successful experimental validation study (Liebsch and Spielmann, 2002). Two in vitro models for skin corrosion have formally been accepted by regulatory bodies; the Transepicutaneous resistance test (TER), based on measuring of the electrical resistance in ex in vivo rat skin and the Human Skin Model Test using a reconstituted human epidermal model (TG 431) (OECD, 2004c). In addition to the TER test and human skin models, the Corrositex membrane barrier test has been proposed for assessing the dermal corrosion hazard potential of chemicals (TG 435) (OECD, 2006).

Although several in vitro test methods for assessment of skin irritation are under evaluation, at the time of writing this report, none of them have been approved as a full alternative for in vivo tests. Some of these methods have been shown to be useful in studying skin irritation of various types of substances for the purpose of internal testing, risk assessment and screening (Faller, 2002). Moreover, some of these tests have passed the last stage of evaluation and have been submitted for formal approval by the regulatory bodies (Cotovio et al., 2005; ESAC, 2007; Kandarova et al., 2005).

The principles of common in vivo and in vitro tests will be addressed below.

## In vivo methods for skin irritation and corrosion

The in vivo rabbit skin test for irritation and corrosion has served as the basis for international regulatory requirements for the testing of chemicals (OECD, 1992, 2002). This test consists of the topical application of substances to the albino rabbit's dorsal skin and the subsequent assessment of induced skin damage. The duration of the study should be sufficient to evaluate the reversibility or irreversibility of the effects observed. A scoring system based on the intensity of erythema and oedema enables products to be classified from non-irritant to very irritant. The dermal irritation scores should be evaluated in conjunction with the nature and severity of lesions, and their reversibility or lack of reversibility.

As rabbit skin is physiologically different from human skin, there has been criticism that the test cannot accurately predict irritation potential in humans (Robinson et al., 2000; Robinson and Perkins, 2002).

For chemicals that lack other toxicities (e.g. mutagenicity, sensitization, corrosivity etc.) 4-h patch test with human volunteers has been developed (Basketter et al., 2004; Robinson et al., 2001). To avoid subjective scoring of swelling or redness, alternative parameters have been introduced such as cutaneous blood flow, as measured by Laser Doppler Flowmetry, infrared detection of skin temperature, and skin thickness assessment. In addition to methodology used for hazard identification, skin irritation has been studied in experimental studies in volunteers. Besides measurements of visible skin effects such as erythema and oedema, the profile of cytokines and other inflammatory mediators have been measured in human volunteers by using punch biopsies, suction blister techniques, and Stratum corneum tape stripping techniques (Ulfgren et al., 2000; Perkins et al., 2001; de Jongh et al., 2006).

## In vitro prediction of skin irritation and corrosion

The development of in vitro alternatives for skin irritation and corrosion has been driven by scientific and ethical considerations (Botham et al., 1998). Due to a limited supply of fresh human skin, a variety of in vitro systems have been developed, including animal ex-in vivo models and keratinocyte cultures. The Skin Irritation Function Test (SIFT) (Heylings et al., 2003) and the Pig Ear Test (PET) (Fentem et al., 2001) are two examples of ex vivo animal models (mouse and pig, respectively). Keratinocytes can be cultured as a monolayer or as stratifying layers, that can also be grown in conditions where the cultured cells reconstruct the basis of an epidermis with three dimensional organization and production of a cornified barrier when the surface of the culture is exposed to the air. Keratinocytes monolayers have been widely used to study dermal irritancy. However, due to the absence of the principle barrier, the Stratum corneum, use of this model leads to higher exposure of keratinocytes (Ponec and Kempenaar, 1995, Coquette et al., 2003).

Currently, the most relevant in vitro tests for regulatory purposes are based on reconstituted human skin equivalents. The reconstituted human epidermis is a multilayered human skin grown in the laboratory which mimics the biophysical properties of in vivo human epidermis. The reconstructed epidermis model presents the following mechanistic advantages: (1) it mimics the architecture of a normal epidermis, where the presence of the stratum corneum is obtained through an imersion step (required for keratinocyte differentiation); (2) it can be used to assess biological parameters underlying various types of damage; and (3) it permits the topical application of chemicals or finished products directly onto the stratum corneum (Borlon et al., 2007). Different end-points can be assessed including tissue viability, histological evaluation, the release of inflammatory mediators and expression of genes taking part in stress response such as cell signalling, inflammation, protein metabolism, etc.

In vitro skin corrosion and in vitro skin irritation assays differ significantly in their test design concerning exposure pattern, follow up time and investigated end point. In comparison to in vitro skin corrosion test, the design of in vitro skin irritation protocols enables long-term development of the cell damage or tissue recovery after exposure to a test chemical. Corrosive materials are identified by their ability to produce a decrease in cell viability by using an appropriate test such as MTT-assay (Mosman, 1983). The cut-off percentage cell viability value distinguises corrosive from non-corrosive test materials.

Unlike corrosion, skin irritation encompasses a vast range of severity. In contrast to in vivo irritation tests which primarily rely on visible changes in the skin such as erythema and erythema, irritation in vitro tests assess various biomarkers for different end-points such as cytotoxicity, morphological changes, differential expression of genes, release of inflammatory cytokines (IL-1 and TNF- $\alpha$ ) and neuropeptides (substance P (SP)) (Coquette et al., 2003; Tornier et al., 2006; Chateerjee et al., 2005). Cytotoxicity is known to trigger irritation processes, and can therefore be a first common event for irritants. Cell cytotoxicity assays such as cell viability (e.g. MTT conversion and the membrane integrity - such as neutral reduptake or LDH release) are amongst the most common in vitro endpoints used to predict the toxicity of a substance in cell culture. Although there is a correlation between irritation potential and cell viability, measurment of cytotoxicity alone does not always discriminate between irritants and non-irritants (Fentem et al., 2001) and it has been recommended that more specific biomarkers, that have different mechanistic response, should be incorporated. For this purpose, the release of a membrane damage marker, adenylate kinase (AK), and of cytokines IL-1a and IL-8 were also investigated. Combining these endpoints, a simple two-tiered strategy was developed, with the MTT assay as the first, screening stage. This resulted in a clear increase in sensitivity, and a fall in the false-positive rate (Cotovio et al., 2005).

Today several reconstructed human skin models are commercially available, of which the EPISKINTM model (EPISKIN-SNC, France), the EpiDermTM model (MatTek Corporation, MA, USA) the SkinEthicTM model (SkinEthic Laboratories, France), EST-1000TM (Cell Systems GmbH, Germany) and Phenion TM (Phenion GmbH, Germany) are the most used.

For skin corrosion, in addition to Human Epidermal Model, two other tests are available which have been formally accepted as in vitro alternatives for in vivo skin tests. The Transcutaneous Electrical Resistance (TER) Assay (OECD, 2004) is based on an irreversible loss of normal stratum corneum integrity and function measured as a reduction in the inherent TER below a predetermined threshold level. A test material is topically applied for up to 24 hours to the epidermal surfaces of skin discs obtained from humanely killed young rats. A validation study and other published studies have reported that the TER assay is able to reliably discriminate between known skin corrosives and non-corrosives, and there was good agreement between the skin corrosivity classifications derived from the in vitro and in vivo data assessed under OECD guideline 404 (OECD, 2002). However, the TER test does not allow the sub-categorisation of corrosive substances. Another in vitro corrosivity test is CORROSITEX in which the potential corrosivity of a test material is assessed by measuring the time required for a chemical to 'breakthrough' a bio-barrier membrane (a reconstituted collagen matrix). The average breakthrough time of four replicates is used to determine whether a chemical is corrosive or not.

### 5.2.2. Test methods for skin sensitization

Under current regulations, the identification of skin sensitization hazard is assessed through in vivo testing. The animal test methods used in harmonised classification of substances, according to their potential to cause skin sensitisation, are the guinea pig maximization test (GPMT), the Buehler test, and the local lymph node assay (LLNA) (ref OECD 406, OECD 429, EU B.6, EU B.42, UNECE GHS). A substance that is classified as a skin sensitizer carries the Risk Phrase R43 according to EC criteria. A substance is either a skin sensitiser or not and there is no subdivision based on potency.

Substances are classified as skin sensitizers if, in properly conducted tests, at least 30 \% of animals show a positive response in a GPMT and $15 \%$ in a Buehler test (Buehler, 1965). For the LLNA a positive response is defined if at least one concentration induces a 3-fold or greater increase in proliferation in draining lymph nodes compared with concurrent vehicle-treated controls (the EC3 value).

The guinea pig maximisation test (Magnusson and Klingman, 1970) and the occluded patch test (Buehler, 1965) are the most widely applied animal models. In these tests, groups of animals (guinea pigs) are topically exposed to the test material. In some tests, an adjuvant is also administered to enhance immune responses. Elicitation of cutaneous hypersensitivity is determined as a function of challenge-induced erythema and/or edema. Sensitizing potential is judged on the basis of the frequency of specific reactions induced by this challenge. One of the drawbacks of the guinea pig tests is that it doesn't provide information on relative potency (ECETOC 2000). In this regard, important progress has been made by development of the LLNA (Kimber and Basketter, 1992; Gerberick et al., 2007; Basketter et al., 2007). This test is based on the events induced during the induction phase of skin sensitization. Skin sensitizers are identified as a function of their ability to provoke lymphocyte proliferation in the draining lymph nodes following topical exposure of mice to chemicals. This method, which has fully been validated as an alternative to guinea-pig tests, can be used not only for hazard identification but also for assessment of the relative skin sensitizing potency of a substance (ECETOC, 2003; Basketter et al., 2007; Gerberick et al., 2007).

### 5.2.3. Models for the prediction of skin sensitization

In vitro and in silico models provide an alternative approach for the assessment of skin sensitization potential but as yet there are no validated methods which are acceptable for regulatory purposes. For the development of in vitro tests, changes in expression of cell surface markers (e.g., CD54, CD80, CD83, CD86) and chemokine receptors (e.g. CXCR4, CCR5 and CCR7) have been investigated by using peripheral blood-derived dendritic cells (Rustemeyer et al., 2003; Aeby et al., 2004; Staquet et al., 2004; Boisleve et al., 2004). However, to establish this as a validated in vitro assay for skin sensitization, further optimization and development are needed.

Another potential alternative approach to skin sensitisation hazard identification is the use of QSARs. QSARs for skin sensitisation comprise so-called 'local' models and 'global' (or general) models (Patlewicz et al., 2007). Local models are typically characterised by a chemical class or by a single chemical mechanism of action. Global models are based on skin sensitisation datasets comprising diverse chemical structures corresponding with a number of mechanisms of action. Expert predictive systems are built upon experimental toxicity results with prediction rules derived from the data. The rules may be based on statistical inference and take the form of QSARs (e.g. TOPKAT, MultiCASE), or based on expert judgment taking the form of SARs describing reactive chemistry (e.g. Derek for Windows (DfW)) or they may be a hybrid of the two approaches (e.g. Tlssue MEtabolism Simulator (TIMES)). At present, none of the existing approaches seemed to be suitable for routine prediction of skin sensitisation hazards of chemicals. Further work is needed before validated QSARs for this endpoint are available (Li et al., 2007).

### 5.3. SUMMARY OF THE METHODS FOR THE ASSESSMENT OF LOCAL EFFECTS OF PETROLEUM HYDROCARBONS

In vivo methods to assess skin irritation include measurements of skin barrier (e.g. TEWL), skin inflammation (erythema and oedema), and release of inflammatory mediators. At present, with the exception of corrosion, no alternative in vitro tests for skin local toxicity are available for regulatory purposes, although some of reconstructed human skin models have passed the last evaluation stage and have been submitted for formal approval by regulatory bodies. Although these models are one of the most promising in vitro systems that have the potential to replace animal assays for assessment of skin corrosion and irritation it must be mentioned that skin equivalents differ from normal human skin in some characteristics. Most models consist of keratinocytes and the other skin cells which play a critical role in inflammatory response such as Langerhans cells are absent. Moreover, the absence of blood-derived and resident leukocytes reduces the complexity of the observable cytokine network (Coquette et al., 2003). Furthermore, skin equivalents showed approximately 10-30-fold higher permeability as compared to normal skin which might lead to an over-prediction of irritants due to the higher penetration rate of applied substances (Perkins et al., 1999). It seems also that the reconstructed epidermal models are more sensitive to some families of compounds (Tornier et al., 2006). To date, alternative approaches to the in vivo skin tests have proved largely successful at identifying severe irritants, but are generally poor at discriminating between agents with mild to moderate irritant potential. Identification of new more specific markers of irritation will improve the applicability of in vitro irritation test. For skin sensitization the LLNA has been shown to be a valuable alternative to in vivo animal models for hazard identification and also for assessment of relative skin sensitizing potency of a substance.

For asessment of local toxicity of petroleum hydrocarbons various in vivo and in vitro methodologies have been used. For the regulatory purpose, usually in vivo rabbit skin tests have been conducted. For research purposes, in vitro systems based on reconstructed epidermis are increasingly being used, but using different toxicological endpoints. One of the concerns about using these systems is their limited barrier function. Since various studies have shown the importance of local tissue concentration, this should be considered in the design of the experiment and interpretation of the results. Another significant problem with tissue culture systems is the necessity for solubilizing the lipophilic hydrocarbons into aqueous media which can be attained by addition of various solubilizers. Furthermore, some of petroleum hydrocarbons are volatile and can evaporate rapidly from the exposure medium affecting the local concentration at the site of action. It is clear that all these experimental factors might influence the outcomes of the study.

## 6. LOCAL SKIN EFFECTS OF PETROLEUM HYDROCARBONS

The primary irritation hazard of a number of petroleum hydrocarbon has been assessed. As seen from Table III, a number of petroleum hydrocarbons were classified as skin irritants (R38) as of the time of writing this report, although in some cases this is based on structure-activity considerations rather than actual laboratory data (Table III).

Table III Results of Dermal Irritation
(Draize Skin Irritation Tests) of Key Gasoline Constituents

| Substance | Summarised Result |
| :--- | :--- |
| n-Hexane | Classified Xi; R38 |
| Cyclohexane | Not irritating to rabbit skin when applied <br> under non-occlusive patches, but more <br> severe and persistent scores when a <br> chamber was utilised to prevent evaporation. <br> (classified Xi;R38) |
| Benzene | Grade two erythema which progressed to <br> grade 3 |
| Heptane | No data (classified Xi;R38) |
| Methyl Cyclohexane | Produces local irritation and thickening with <br> repeated dermal application |
| Toluene | Grade 2 erythema observed which did not <br> clear within 7 days. |
| Octane | Classified Xi;R38 <br> No data |
| Ethyl Benzene | Considered "moderately irritating to the skin" |
| Xylenes | Classified Xi;R38 |
| Tri-methyl Benzene | Classified Xi; R38 |
| n-Propyl Benzene | Considered "slightly irritating" (isopropyl <br> benzene as a surrogate) |
| Naphthalene | Naphthalene is considered to be a "slight" <br> skin irritant and is not classified. |

The mechanism of the irritation has been studied in more detail for several individual hydrocarbons and petroleum products. The most investigated individual hydrocarbons are aromatic solvents toluene and xylene. Prolonged skin contact with these aromatic hydrocarbons can lead to burning sensation, erythema, delipidization, and dermatitis (see for review Rowse et al., 2004). Occluded exposure to toluene, for as little as 10 min , has been documented to significantly increase the blood flow in vivo in humans (Mahmoud and Lachapelle, 1985). This is in agreement with human volunteer study of Kezic et al. in which skin irritation was observed after short-time ( 3 min ) exposure to neat toluene resulting in enhanced absorption rates (Kezic et al., 2001). In another study in human volunteers, dermal application of toluene caused erythema, increase in TEWL and skin blood flow and a decrease in skin moisture content (Frosch and Kurte, 1994). Toluene is often used in experimental irritation studies as a model skin irritant. The degree of induced irritancy is preparation-, dose- and exposure-time dependent with symptoms ranging from very faint erythema to severe irritation in humans (Frosch and Kurte, 1994).

Also animal studies show that these aromatic solvents may act as skin irritants. Histological changes in guinea pig skin due to exposure to toluene after a 2-h occlusive application showed degenerated epidermal cells such as pyknosis of nuclei and perinuclear oedema. (Mahmoud et al., 1984)

In a recent study (Ahaghotu et al., 2005) using hairless rats in vivo the effect of methyl substitution of benzene on the skin irritation was assessed by measuring TEWL, erythema, skin histopathology and expression of interleukin-1a (IL-1 a) and tumor necrosis factor- $\alpha$ (TNF- $\alpha$ ) in the skin and blood. Rats were exposed dorsally to neat chemicals (dose of $15 \mu$ every 2 hours for 8 hours a day for four days over an area of $3 \mathrm{~cm}^{2}$ ). Test chemicals used were benzene, xylene and tetramethylbenzene (TMB). As measured by TEWL all chemicals significantly disrupted skin barrier compared to controls (at 24 h TEWL values for controls, benzene and xylene, and TMB were $5.59 \pm 0.52,11.3 \pm 1.66,11.14 \pm 0.90$ and 8.15 $\pm 1.00 \mathrm{gm}^{-2} \mathrm{hr}$, respectively). Although TMB induced relatively lower TEWL than xylene and benzene at 24 h , it exceeded benzene and xylene at 32 h and remained at a higher level throughout $104 \mathrm{~h}\left(16.82 \pm 2.42 \mathrm{~g} \mathrm{~m}^{-2} \mathrm{hr}\right)$. Unocclusive exposure resulted also in mild skin irritation; xylene induced slightly higher irritation then benzene and TMB ( $0.8 \pm 0.20,0.4 \pm 0.24$ and $0.40 \pm 0.16$, respectively). As irritation increased with time during the application, TMB demonstrated highest irritation score of $2.6 \pm 0.16$ at 104 h . Both TEWL and erythema demonstrated increases with the lipophilicity or methyl substitution of the aromatic hydrocarbons. Skin histopathology revealed only slight swelling and disruption of the stratum corneum due to exposure to benzene, while besides swelling and disruption, exposure to xylene induced granulocyte infiltration in the epidermis and infiltration of homogenous eosinophilic material at epidermal-dermal separation which is in agreement with the findings of Gunasekar et al. (Gunasekar et al., 2003). Most severe damage was induced by TMB. Additionally, TMB induced high increase of mast cells in the areas of inflammation which is thought to be an immunological reaction due to exposure to TMB. High levels of IL-1a were only found in the blood suggesting rapid clearance of the cytokines from the skin. Almost 12 fold levels of IL-1 $\alpha$ by TMB indicate severe irritation response of the skin. In contrast, TNF- $\alpha$ was only found in the skin indicating its local accumulation induced by irritation and inflammation (being highest for TMB; 11-fold in comparison to benzene and xylene; 5 - and 6-fold, respectively).

The effects of short-term and long-term occlusive and unocclusive dermal exposure to xylene and benzene on the skin irritation response, skin moisture content, erythema and cytokine/chemokine levels were investigated in hairless rats in vivo (Chatterjee et al., 2005). The dorsal skin of rats was exposed occlusively (short-term exposure) to $230 \mu \mathrm{l}$ of xylene or benzene over an area of $1.04 \mathrm{~cm}^{2}$ for 1 hour. Unocclusive (long-term exposure) repeated exposure was performed by application of $15 \mu \mathrm{l}$ of xylene or benzene every 2 h for 8 h a day, for four days over an area of 3 $\mathrm{cm}^{2}$. The measurements were performed at different time points before, during (only unocclusive exposure) and after the end of exposure. For the assessment of biomarkers in skin and blood, blood samples were collected at 24 and 104 h . As measured by TEWL both chemicals significantly disrupted the skin barrier upon occlusive and unocclusive exposure (TEWLs at 1 h under occlusive conditions were $11.3 \pm 1.66$ and $15.96 \pm 2.21 \mathrm{~g} \mathrm{~m}^{-2} \mathrm{hr}$ and at 24 h under unocclusive conditions were $11.3 \pm 1.66$ and $11.14 \pm 0.90 \mathrm{~g} \mathrm{~m}^{-2} \mathrm{hr}$ for benzene and xylene, respectively; control TEWLs were $6.20 \pm 0.45$ and $5.59 \pm 0.52 \mathrm{~g} \mathrm{~m}^{-2} \mathrm{hr}$ for occlusive and unocclusive exposure, respectively). Skin moisture content due to occlusive benzene exposure was lowered after 7 h being significantly lower than control at 8 h and 24 h , while xylene induced consistently lower skin moisture content. Prolonged unocclusive exposure to benzene and xylene induced a low level of moisture content throughout
the experiment (104 h). The effect of the short-term occlusive exposure to benzene and xylene showed that xylene is more irritating than benzene to hairless rat skin (at 24 h score was $1.8 \pm 0.2$ and $1.2 \pm 0.2$ for xylene and benzene, respectively). Unocclusive long term exposure to benzene and xylene induced mild irritation within 8 h which steadily increased with time and reached maximum score of $1.8 \pm 0.2$ and $2.0 \pm 0.32$ for benzene and xylene, respectively. Occlusive exposure induced 2.4and 2.7 -fold levels of IL-1a, while unocclusive exposure induced 3.7 and 3.9 -fold increased levels of IL-1a for benzene and xylene, respectively. Exposure to benzene and xylene induced similar increase of TNF- $\alpha$ in the skin: 2.3- and 2.5 -fold for occlusive exposure and 6.2- and 5.7 -fold for unocclusive exposure. Both occlusive and unocclusive exposure elevated monocyte chemoattractant protein-1(MCP-1) levels in the skin approximately by 1.7 fold compared to the control. Presented data showed that repeated long-term, low-level exposure is more harmful than short-term, high-level exposure in respect to benzene and xylene. Furthermore, there was a significant difference in irritation potential between occlusive and unocclusive exposure.

Gunasekar et al. (2003) studied histological and molecular changes in the skin that may reflect induced dermal irritation as a response to exposure to neat m-xylene. Hairless rats were exposed to m-xylene on their backs over an area of $4.9 \mathrm{~cm}^{2}$. The skin sites were excised at zero, one, two, four and six hours after the beginning of the exposure. As molecular marker of inflammation interleukin-1a (IL-1a) and inducible nitric oxide synthase (iNOS) proteins were measured. Local skin damage due to exposure to m-xylene was suggested by accumulation of homogenous eosinophilic material at the areas of epidermal-dermal separations. Exposure to mxylene also induced increase of IL-1a (50-85\% above the control samples, maximum at 1-2 hours) and iNOS protein (15-60\% above control skin samples with a maximum at 4 hours).

Rogers et al. (2001) studied hairless rats ( $6-10$ per group) in vivo whether oxidative species and low molecular weight (LMV) DNA could be detected in the skin following dermal exposure to m-xylene. Hairless rats were exposed to m-xylene over an area of $4.9 \mathrm{~cm}^{2}$ for 1 hour. The skin sites were excised at zero, one, two, four and six hours after the beginning of the exposure. Oxidative species, determined by measuring the oxidation of $2^{\prime}, 7$ '-dichlorofluorescin diacetate, were found in both exposed and unexposed skin being higher in exposed sites (1.4-2.0 fold greater than unexposed site). There were also high levels of low molecular weight DNA present in both exposed and unexposed skin being significantly higher in the skin excised after two, four and six hours (110.9-588.5 ng LMV DNA/ $\mu \mathrm{g}$ genomic DNA).

In another study (Singh et al., 2003) the local skin effect of dermal exposure to xylene, heptane and hexadecane was investigated in vivo in weanling pigs. Weanling pigs were exposed to neat chemicals for 24 h over an area of $1 \mathrm{~cm}^{2}$. Measurements of TEWL and erythema were performed at zero h, $1 \mathrm{~h}, 2 \mathrm{~h}$ and 24 h after the end of exposure. Heptane caused significant increase in TEWL (2-fold compared to control) while hexadecane and xylene did not produce any significant changes in TEWL at any time points after the end of exposure. Slight erythema was observed at all chemically treated sites and it was completely resolved at sites treated with hexadecane and xylene, while elevated erythema persisted at sites treated with heptane 24 h after the end of exposure. None of the investigated hydrocarbons caused edema.

The effect of unocclusive dermal exposure to nonane, dodecane and tetradecane on skin irritation was investigated in hairless rats in vivo by measurement of
erythema and TEWL. Furthermore, expression of IL-1a, TNF- $\alpha$ and monocyte chemoattractant protein-1 (MCP-1) were determined in the skin and blood at the end of dermal exposure (Babu et al., 2004). The chemicals were applied unocclusively every 2 h for 8 h a day for four days over an area of $3 \mathrm{~cm}^{2}$. All the chemicals demonstrated a significant increase in TEWL during the exposure period; the increase in the TEWL was in the following descending order: tetradecane> dodecane> nonane. Also the erythema scores increased with increase in the molecular weight. Nonane and dodecane caused moderate erythema while tetradecane produced severe erythema. The expression of IL-1a in the blood and TNF- $\alpha$ in the skin was higher for tetradecane followed by dodecane and nonane. These results are in contrast with those reported by the same research group (Babu et al., 2004a) in which single occlusive application of these chemicals produced higher irritation by nonane than dodecane and tetradecane as assessed by TEWL and erythema. Tetradecane did not show any visible signs of skin irritation and also did not produce any significant difference in IL-1 $\alpha$ and MCP-1 as compared with the control. Nonane significantly increased the expression of IL-1a, TNF- $\alpha$ and monocyte chemoattractant protein-1 (MCP-1) in skin and blood as compared to the control. Tetradecane and dodecane did not elicit IL-1a release either in the skin or in blood, whereas nonane exposure showed higher IL-1a levels in blood. The expression of TNF- $\alpha$ was higher in response to all the chemicals but significantly more with nonane than dodecane or tetradecane.

Expression of a number of genes involved in inflammatory and growth-related pathways has been studied in rats in vivo after brief dermal exposure to JP-8 and 4 of its constituents: undecane (UND), tetradecane (TET), trimethylbenzene (TMB) and dimethylnaphtalene (DMN) by using a microarray technique (McDougal and Garrett, 2007). Results of this study revealed that all investigated chemicals caused responses that may result in irritation. UND and TMB showed the greatest number of gene changes, more than twice as many as DMN, and about 10-fold more than TET. UND caused changes in development, morphogenesis, and cell differentiation whereas TMB caused dramatic transcript decreases in metabolism categories. Differences in gene expression between investigated compounds showed not to be caused by different epidermis concentrations.

Release of neuropeptide $P(S P)$ as a proinflammatory biomarker was measured after topical occlusive exposure to xylene in rats in vivo by using a microdialysis technique (Fulzele et al., 2007). Occlusive exposure to xylene produced significant SP release after 2 and 3 h following removal of occlusion, however no significant induction of SP was observed during the period of occlusive treatment.

An in vivo study was conducted with Yucan minipigs to evaluate dermal toxicity of neat JP-8, nonane or toluene (Kanikkannan et al. 2001a). TEWL, skin capacitance (moisture content), erythema and edema were evaluated before and after treatment. Application of toluene increased the TEWL by 1.4 times at 2 h after the removal of patches and the TEWL value remained high until 24 h though it was statistically insignificant compared to base line level ( $p>0.05$ ). Exposure of pig skin to nonane caused a gradual increase in the TEWL during the 24 h post application period and the TEWL was about 2 times higher at 24 h than the baseline value ( $p<0.05$ ). Both toluene and nonane caused slight erythema. Application of toluene did not cause a significant edema whereas nonane caused a slight edema, which gradually decreased after 24 h following removal of the patches.

Muhammad et al. (2005b) assessed the irritation caused by JP-8 and eight aliphatic hydrocarbons (nonane, decane, undecane, dodecane, tridecane, tetradecane, pentadecane, hexadecane) and 6 aromatic hydrocarbons (ethyl benzene, o-xylene,
trimethyl benzene, cyclohexyl benzene, naphthalene, dimethyl naphthalene) which were topically exposed to pigs for 1 day and with repeated daily exposures for 4 days. Erythema, epidermal thickness, and epidermal cell layers were determined. Erythema and epidermal hyperplasia were pronounced after 4 days of repeated application of the aliphatic hydrocarbons. Aromatic hydrocarbons such as ethyl benzene, o-xylene, trimethyl benzene, cyclohexyl benzene, naphthalene, and dimethyl naphthalene did not produce any macroscopic or significant microscopic changes in epidermal thickness or lesions after 1 or 4 days of in vivo exposures. Morphological observations revealed slight intercellular and intracellular epidermal edema in 4-day exposures with the aliphatic hydrocarbons. Ultrastructural studies showed that jet fuel hydrocarbons-induced cleft formation within intercellular lipid lamellar bilayers of the stratum corneum. The degree of damage to the skin was proportional to the length of in vivo hydrocarbons exposures. A hydrocarbon-specific response was demonstrated by epidermal thickness and the number of epidermal cell layers, with tridecane and tetradecane having the greatest proliferative effect followed by JP-8 and pentadecane after 4 days of in vivo exposures. The short chain aliphatic hydrocarbons such as nonane, decane, and undecane produced only mild erythema after 4 days of in vivo exposures. These findings are in accordance with Brown and Box (1970) who studied the skin irritancy of alkanes and reported that $n$-decane was slightly irritant with some epidermal thickening, while $n$ tetradecane was more irritant with epidermal thickening. The same research group (Muhammad et al., 2005b) reported that individual hydrocarbons may cause lipid extraction from the stratum corneum, as studied with Fourier Transform Infra Red (FTIR) spectroscopy and transmission electron microscopy (TEM) (Monteiro-Riviere et al., 2004).

In several in vitro studies, various biomarkers of skin irritation were investigated after exposure to aliphatic and aromatic hydrocarbons providing more insight in the molecular mechanisms of the inflammatory effects.

Oxidative damage from dermal exposure to the vapours of xylene and toluene was investigated in human skin in vitro (Costa et al., 2006). Skin integrity was assessed by measuring TEWL and percutaneous penetration of a model compound 1,2,4trimethylbenzene. In addition to skin viability as assessed by the MTT test, the activity antioxidant systems, including glutathione transferase, catalase and superoxide dismutase, were determined. As biomarkers of oxidative stress, products of lipid and protein peroxidation were measured in the skin. 8-h dermal exposure to vaporous xylene and toluene significantly reduced tissue viability when concentrations higher then $10^{4} \mathrm{ppm}$. The overall trend of the results indicates that both solvents can damage human skin even at concentrations comparable to their TLV, and that oxidative stress has a role in these effects.

The toxicity of the JP-8 components m-xylene, 1-methylnaphthalene (1-MN), and $n$ nonane was studied in the murine keratinocyte line (Rogers et al., 2004). The viability of keratinocytes was assessed by measuring reduction of MTT and was expressed as the $\mathrm{EC}_{50}$ value (chemical concentration at which $50 \%$ cell viability was observed). The viability of keratinocytes exposed to $m$-xylene, 1-MN, and n-nonane decreased with increasing chemical concentration. The results show that m-xylene is the most potent, followed by n-nonane, and $1-\mathrm{MN}$ as the least potent, supporting earlier findings that the aromatic hydrocarbons cause greater direct cytotoxicity than aliphatic hydrocarbons (Chou et al., 2003).

The cytotoxic effect of m-xylene in rat dermal fibroblast and the role of oxidative stress were studied by Coleman et al. (2003). Viability of fibroblast in dermal equivalents was assessed using MTT assay. Levels of two cellular antioxidants,
endogenous catalase and endogenous thiols, were assessed spectrophotometrically and by using Ellman's reagent, respectively. The results show that m-xylene induces cytotoxicity as assessed by decreased viability of the fibroblast with calculated $\mathrm{EC}_{50}$ value to be $1481.1 \pm 88.2$ and $930 \pm 32.5 \mu \mathrm{~g} \mathrm{~m}$ xylene/g cell at 1 and 4 hour exposure, respectively. The observed $\mathrm{EC}_{50}$ at 1 hour exposure was about $60 \%$ higher than the $\mathrm{EC}_{50}$ at 4 hour exposure. M-xylene was also shown to promote decreases in cellular antioxidant levels in time- and dosedependent manner. The maximum observed decrease in thiol level was about 20 and $40 \%$ (at 1 and 4 hour exposure, respectively) and in catalase activity was about 10 and $55 \%$ (at 1 and 4 hour exposure, respectively).

Chou et al. (2002) investigated the effect of chain length of individual aliphatic hydrocarbons on cell toxicity and release of IL-8 in human epidermal keratinocytes (HEK) in vitro. 10 aliphatic hydrocarbons with carbon (C) length ranging from 6 to 16 were applied neat on HEK for 1,5 and 15 mins. All treatments caused significant dose dependent increase in cell mortality which corresponded with the decrease in carbon chain length. However, the increase in IL-8 release showed a peak in response around C9-C13. Higher cytotoxicity by shorter-chain aliphatic hydrocarbons did not correlate with IL-8 release revealing different mechanisms by which these compounds exert toxicity in cell cultures.

In another study from the same research group (Chou et al., 2003) in vitro cutaneous toxicity of 9 individual aromatic hydrocarbons was assessed in HEK cells. The cytotoxicity was evaluated by determining the dose causing $50 \%$ mortality $\left(\mathrm{LD}_{50}\right)$ and by the highest non-cytotoxic level (HNTL) (5\% of HEK cells mortality). In addition, IL-8 release at selected dose was measured. The increase in cytotoxicity was correlated to the number and the size of the side-chains attached to the aromatic ring. $\mathrm{LD}_{50}$ rank order potency was cyclohexylbenzene > trimethylbenzene > xylene > dimethylnaphthalene > ethylbenzene > toluene> benzene. At the $L D_{50}$ dose level ethyl benzene, cyclohexyl benzene (which induced significantly the highest IL-8 level compared to other aromatic hydrocarbons), dimethyl naphthalene and methyl naphthalene significantly induced IL-8 release while other aromatic hydrocarbons induced equivalent or slightly higher IL-8 levels compared to controls. The study showed that ability to cause HEK cell death does not entirely correspond to ability to induce IL-8 release which suggests that different mechanisms of action are responsible for skin toxicity and irritation.

Moloney and Teal (1988) investigated a structure activity relationship for neat $n$ alkanes in a mouse ear edema model to study the mechanism of cumulative irritancy. Animals were repeatedly applied with $5 \mu$ of $n$-alkanes in hexane to the whole of the pinna of the ear for zero, $8,24,32,48,56,72$ and 80 h over four days. Ear edema was quantified by measuring of the thickness of the tip of the pinna at zero, $24,48,72$ and 96 h after the beginning of the exposure. Control exposure was conducted with pure hexane and resulted in no edema. Dodecane had no effect on the pinna thickness, while tridecane induced significant increase at 96 h . Tetradecane induced the greatest response, while hexadecane, octadecane and eicosane produced progressively diminished responses. Octadecane was the only $n$-alkane which induced a significant response already at 24 h . At 96 h the treated sites were exposed to hydrocortisone which revealed a positive correlation between ear thickness and induced permeability of the ear to hydrocortisone. In an additional cross-over experiment mouse pinna was initially treated with tetradecane and further treated with tetradecane, or dodecane or hexane. Continued treatment with tetradecane again resulted in substantial increased of ear thickness and with dodecane in only modest increase of the ear thickness while hexane treatment resulted in no increase of the thickness of the ear.

Allen et al. (2001) studied the effect of individual aliphatic hydrocarbons (undecane, dodecane, tridecane and hexadecane) on IL-8 production by normal human epidermal keratinocytes (NHEK) in vitro after exposure for 24 h . The results revealed that in vitro, individual aliphatic hydrocarbons are capable of inducing IL-8 release from keratinocytes at concentrations that were determined to be subtoxic to cellular membranes. There appear to be differences among each hydrocarbon with respect to their effects on IL-8 release. At subtoxic doses IL-8 release was increased in decreasing intensity as follows: tridecane (C13) > hexadecane (C16) > dodecane $(\mathrm{C} 12)>$ undecane $(\mathrm{C} 11)>$ control. However, these results should be interpreted with caution because these differences might be attributed not only to different effects of aliphatic hydrocarbons but also to variations in complexation or differences in their release from the introducing complex with $\alpha$-cyclodextrins influencing the actual concentration.
lyadomi et al. (2000) investigated the irritancy properties of several aliphatic and aromatic solvents including toluene, $m$-xylene, and $n$-hexane using a mouse ear thickness model. All investigated solvents produced an increase in ear thickness; the strongest effect was observed for toluene and $m$-xylene. Toluene showed a clear dose-response relationship, however at concentrations under $30 \%$ toluene did not produce signs of skin irritancy. The same author reported plasma extravasation and inflammation in the abdominal skin of hairless rats induced by topical exposure to toluene, $m$-xylene and toluene (lyadomi et al., 1998).

Yang et al. (2006) studied cytotoxicity and proinflammatory activity of 10 aliphatic (C6, C8-C16) and 7 aromatic (benzene, ethylbenzene, trimethylbenzene, toluene, xylene, cyclohexylbenzene and dimethylnaphthalene) hydrocarbons in HEK. The study was designed to investigate any interactions among alophatic and aromatic hydrocarbons in order to understand how these compounds cross-react in a mixture situation similar to the actual composition of jet fuels. Two parameters representing direct cytotoxicity to the living cells (HEK mortality) and the activation of proinflammatory pathways (IL-8) were evaluated. The application solution was dermally applied as a mixture of aliphatic or aromatic solvents or as a mixture of JP8 fuel and a single hydrocarbon. The results suggested that middle-ranged aliphatic components (C10-C13) are likely to be the principal hydrocarbons dictating HEK toxicity, while C14-C16 may be the most important components in composing a less dermatotoxic fuel. Increase of aliphatic components in JP-8 generally reduced HEK mortality in a dose-dependent manner. Interestingly, when aliphatic and aromatic hydrocarbons were dosed in combination, the mixture may exhibit a different cytotoxicity or the degree of toxicity may change significantly. Whether or not the changes follow certain patterns or changes in a predictable way is not clear. The results suggested that the significance of aromatic components in a complex mixture like JP-8 seemed to be well below that of the aliphatic components in terms of dermatotoxicity.

## Dermal sensitization studies of petroleum hydrocarbons

Petroleum hydrocarbons which have been tested for allergic contact dermatitis either in laboratory animals or in human volunteers do not appear to be contact sensitizing agents (CONCAWE 2009).

### 6.1. SUMMARY OF THE LOCAL SKIN EFFECTS OF PETROLEUM HYDROCARBONS

There are several lines of evidence that show dermal exposure to both aliphatic and aromatic carbons leads to local skin irritation. Development of a local skin effect is dependent on the local concentration of a chemical in the skin and its intrinsic potential to cause effects on skin structure and induce an inflammatory response. While aromatic hydrocarbons penetrate the skin better than aliphatics, the aliphatic components are absorbed into and remain in the skin to a greater extent than the aromatics which can cause their gradual accumulation in the skin in repeated exposure. In general, aromatic compounds are more irritating to the skin than aliphatic compounds (Boman 1996; Hoekstra and Phillips 1963; Klauder and Brill 1947, Yang et al., 2006). The aromatic hydrocarbons are more potent in causing keratinocyte cell death; methyl substitutions make benzene rings more cytotoxic (Ahaghotu et al. 2005) and reactivity increases with the number of the side chains attached to the aromatic ring. On the other hand, the aliphatic hydrocarbons are more potent at inducing the release of proinflammatory cytokines. In general, straight chain hydrocarbons are more irritating than branched hydrocarbons with the same number of carbon atoms. The skin irritating potential of aliphatic hydrocarbons increased with increase in the molecular weight and paralled their affinity to the stratum corneum. The ranking order of irritant potential of aliphatic hydrocarbons reported in various studies seemed to be dependent on experimental conditions such as occlusion which might influence partitioning and local bioavailability of a chemical. Although several studies showed that for a series of hydrocarbons (C6C16) irritancy peaked around tetradecane (C14) (Brooks and Baumann 1956; Brown and Box 1970; Moloney and Teal, 1988), a recent study identified such middle ranged aliphatics as the most cytotoxic (Yang et al., 2006).

There is increasing evidence that severe, dermal irritation induced by long-term or repeated exposure to certain hydrocarbon fuels can be the basis of the progressionpromoting effect in the development of skin tumours. The tumour-promoting activity of alkanes is related to their chain length, with maximal activity found in C12-C14 alkanes.

In contrast to skin irritating effects, there is no evidence that petroleum hydrocarbons cause skin sensitization.

## 7. CONCLUSIONS: HEALTH RISK OF DERMAL EXPOSURE TO PETROLEUM HYDROCARBONS

### 7.1. SYSTEMIC UPTAKE AND EFFECTS

Dermal uptake of hydrocarbons can occur after skin contact with a petroleum product in either liquid or vapour form. Studies on the dermal absorption of individual hydrocarbons in petroleum products show a clear decrease of dermal absorption with increasing lipophilicity and molecular weight. Accordingly, higher absorption of aromatics (e.g. toluene, xylene, trimethylbenzene, naphthalene) than aliphatics (decane, dodecane, tridecane and hexadecane) has been reported. Further, the long chain aliphatics (e.g. tetradecane, pentadecane) penetrated in smaller amounts than shorter chain aliphatics.

Available data on the dermal absorption of individual hydrocarbons from petroleum products suggest that dermal exposure to these substances, even following longterm exposures such as in an occupational setting, will not cause systemic toxicity under normal working conditions and assuming an intact skin barrier. Furthermore, in most exposure scenarios, volatile hydrocarbons (such as aromatics) will easily evaporate from the skin, reducing dermal exposure and uptake. Dermal absorption of vapours can be considered negligible, as, even in the worst case scenario where total body exposure is assumed, dermal uptake will be much lower than respiratory uptake at identical air concentrations. Repeated skin contact with petroleum products may however lead to an impaired skin barrier and consequently enhanced uptake of hydrocarbons. To avoid this there is a need to minimise skin contact through use of protective clothing and gloves. For the purposes of quantitative risk assessment from dermal exposure, data on (or estimates of) external exposure (i.e. concentration of a hydrocarbon in petroleum product, duration of exposure, exposed skin area) and absorption rate of hydrocarbons are needed. In Figure 6, and Tables IV and V, a scheme for risk assessment of dermal exposure and the potential sources of data needed are presented.

### 7.2. LOCAL SKIN EFFECTS

Skin contact with certain petroleum substances or products may cause skin irritation, leading to dermatitis, particularly after repeated or prolonged exposure, which is assumed to be caused by aliphatic and aromatic hydrocarbons constituents. Aromatic solvents tend to be more irritant than aliphatic compounds. A number of petroleum hydrocarbons are classified as skin irritants, according to EU criteria.

In addition to the irritation effects of petroleum hydrocarbons, the skin barrier function may be affected following repeated contact, making the skin more susceptible to other irritants, sensitizing agents, and bacteria and also enhance the dermal penetration of other subtances. Furthermore, there is increasing evidence that severe, dermal irritation induced by long-term or repeated exposure to certain hydrocarbons can contribute to the progression-promotion effect and the development of skin tumours. The tumour-promoting activity of alkanes is related to their chain length, with maximal activity found in C12-C14 alkanes.

Petroleum hydrocarbons which have been tested for allergic contact dermatitis do not appear to be contact sensitising agents.

Figure 6 Health risk assessment of dermal exposure to petroleum hydrocarbons


Table IV
Tiered approach for data generation on skin absorption

| Increasing |
| :--- |
| Uncertainty |

$\left.\begin{array}{|l|l|l|}\hline \begin{array}{l}\text { Data on dermal } \\ \text { absorption }\end{array} & \text { Predictive value } & \text { Remarks } \\ \hline \begin{array}{l}\text { In-vivo human } \\ \text { Occupational } \\ \text { studies } \\ \text { Volunteer studies }\end{array} & & \begin{array}{l}\text { Preferred data for risk assessment } \\ \text { (gold standard) }\end{array} \\ \hline \text { In-vivo animal } & \begin{array}{l}\text { Pig skin is the best } \\ \text { model animal for human } \\ \text { skin } \\ \text { Rat skin overestimates } \\ \text { human dermal } \\ \text { absorption }\end{array} & \begin{array}{l}\text { Application of rat skin data results } \\ \text { in a conservative risk assessment }\end{array} \\ \hline \begin{array}{l}\text { In-vitro } \\ \text { Human or animal } \\ \text { skin }\end{array} & \begin{array}{l}\text { Lower prediction value } \\ \text { for highly lipophilic } \\ \text { chemicals }\end{array} & \begin{array}{l}\text { When full thickness skin used, the } \\ \text { amount of a chemical in all skin } \\ \text { layers should be taken for the } \\ \text { calculation } \\ \text { Results obtained from dermal } \\ \text { absorption studies with petroleum } \\ \text { products are favoured above those } \\ \text { carried out in aqueous solutions } \\ \text { which overestimate dermal } \\ \text { absorption }\end{array} \\ \hline \begin{array}{l}\text { QSAR } \\ \text { Skinperm } \\ \text { Dermwin }\end{array} & \begin{array}{l}\text { Limited predictivity } \\ \text { Not sufficiently validated }\end{array} & \begin{array}{l}\text { Kp significantly overpredicts dermal } \\ \text { absorption of highly lipophilic } \\ \text { petroleum products (log Kow>4) } \\ \text { Maximal flux based on permeability }\end{array} \\ \text { 100\% } \\ \text { absorption }\end{array} \quad \begin{array}{l}\text { coefficient, Kp and maximal } \\ \text { solubility of a hydrocarbon in water } \\ \text { gives better estimate than (Kp) } \\ \text { alone } \\ \text { Skinperm calculates dermal } \\ \text { absorption to both, vapours and } \\ \text { liquids. Furthermore it takes into } \\ \text { account evaporation from the skin } \\ \text { and calculates maximal flux based } \\ \text { on maximum water solubility }\end{array}\right\}$

Table V Tiered approach for data generation on skin exposure

| Increasing Uncertainty | Source of data on dermal exposure | Remarks |
| :---: | :---: | :---: |
|  | Field studies Experimental studies | Limited standardisation and evaluation of the methods used |
|  | Exposure modelling e.g. <br> EASE <br> DREAM <br> RISKOFDERM | Not sufficiently validated <br> Most models do not take into account protective clothing and evaporation from the skin <br> Most models show overprediction of exposure |

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## 9. ABBREVIATIONS

| ACGIH | American Conference of Governmental Industrial Hygienists |
| :---: | :---: |
| AHC | Aliphatic hydrocarbons |
| AK | Adenylate kinase |
| ANN | Artificial neural network |
| ANOVA | Analysis of variance |
| AOEL | Acceptable operator exposure level |
| ASGDI | Atmospheric sampling glow discharge ionization |
| ATR-FTIR | Attenuated total reflectance-Fourier transform infrared spectroscopy |
| AUC | Area under the curve |
| BPD | Biocidal products directive |
| BSA | Bovine serum albumin |
| C | Carbon |
| CEFIC | European chemical Industry council/Conseil Européen de l'industrie chimique |
| CEN | The European committee for standardization |
| COLIPA | European cosmetic toiletry and perfumery association |
| D | Diffusion coefficient ( $\mathrm{cm}^{2} / \mathrm{h}$ ) |
| DEO | Dermal exposure operation |
| DIEGME | Diethylene glycol-monomethyl ether |
| DMN | Dimethylbenzene |
| DMSO | Dimethysulfoxide |
| DOEL | Dermal occupational exposure limit |
| DREAM | Dermal exposure assessment method |
| EASE | Estimation and assessment of substance exposure |
| ECB | European chemical bureau |
| ECD | Electron capture detection |
| ECETOC | European centre for ecotoxicology and toxicology of chemicals |
| ECVAM | European Centre for Validation of Alternative Methods |
| EDETOX | Evaluation and Predictions of Dermal Absorption of Toxic Chemicals |
| EHC | Environmental health criteria |
| EPA (USA) | Environmental protection agency |
| ESAC | ECVAM Scientific Advisory Committee |
| ESR | Existing substances regulation |
| EUROPOEM | European predictive operator exposure model |
| FDA (USA) | Food and drug administration |
| FID | Flame ionization detection |
| FoD | Factor of difference |
| FPD | Flame photometric detection |
| FTIR | Fourier Transform Infra Red |
| GC | Gas chromatography |
| GHS | Globally harmonized system for the classification and labeling of chemical substances and mixtures |
| GPMT | The guinea pig maximization test |
| HEK | Human epidermal keratinocytes |
| HNTL | Highest non-cytotoxic level |
| HPLC | High performance liquid chromatography |
| HSE (UK) | Health and safety executive |


| ICH | International conference on harmonization |
| :---: | :---: |
| IL | Interleukin |
| iNOS | inducible nitric oxide synthase |
| IPPSF | Isolated perfused porcine skin flap |
| IR | Infrared |
| K | Partition coefficient |
| $\mathrm{K}_{\mathrm{p}}$ | Permeability coefficient ( $\mathrm{cm} / \mathrm{h}$ ) |
| LC | Langerhans cells |
| LLNA | Local lymph node assay |
| LMV | Low molecular weight |
| $\log \mathrm{K}_{\text {ow }}$ | Lipophilicity of the petroleum hydrocarbons |
| LSD | Linear system dynamics |
| MS | Mass spectrometry |
| MW | Molecular weight |
| MTT | MTT [3-(4,5-dimethylthiazol-2-yl)-2,5-diphenyl tetrazolium bromide] |
| NHEK | Normal human epidermal keratinocytes |
| OECD | Organization for economic cooperation and development |
| PAH | Polycyclic aromatic hydrocarbons |
| PBPK | Physiologically based pharmacokinetic model |
| PET | Pig Ear Test |
| Pow | $\log \mathrm{K}_{\text {ow }}$ |
| PPM | Part per million |
| RSD | Relative standard deviation |
| QSAR | Quantitative structure activity relationship |
| QSPeR | Quantitative structure permeability relationship |
| SAR | Structure activity relationship |
| SCCNFP | The scientific committee on cosmetic products and non-food products intended for consumers |
| SD | Standard deviation |
| SE | Standard error |
| SEM | Standard error of the mean |
| SIFT | The skin irritation function test |
| $\mathrm{t}_{\text {lag }}$ | Lag time (h) |
| TED | Thermal emission decay |
| TEM | Transmission electron microscopy |
| TER | The transcutaneous electrical resistance assay |
| TET | Tetradecane |
| TEWL | Trans-epidermal water loss |
| TMB | Trimethylbenzene |
| TNF- $\alpha$ | Tumor necrosis factor |
| UN | United Nations |
| UND | Undecane |
| USAF | United States Air Force |

VITAE Video imaging technique to assess exposure
Volatile organic compounds
WBC White blood cells
WHO
World health organization

## APPENDIX 1 SUMMARY DATA ON DERMAL ABSORPTION PARAMETERS FOR INDIVIDUAL HYDROCARBONS

Aliphatic hydrocarbons

| Application of hexane | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Hexane vapour | In vivo, rats |  | 0.0065 | $0.031 \pm 0.004$ | $\begin{aligned} & \text { McDougal et al, } \\ & 1990 \end{aligned}$ |
| Hexane vapour | In vivo, human volunteers |  |  | $0.0051 \pm 0.0036$ | Kezic et al, 2000 |
| Table 2: Heptane - summary |  |  |  |  |  |
| Application of heptane | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vivo pigs | 0.18 |  |  | Singh et al., 2003 |
| Jet fuel (JP-8) | In vitro human cadaver skin In vitro pig skin |  | $\begin{gathered} \times 10^{-6} \\ 2.669 \pm 0.577 \\ 4.540 \pm 0.550 \end{gathered}$ | $\begin{gathered} \times 10^{-5} \\ 10.65 \pm 2.31 \\ 18.22 \pm 2.20 \end{gathered}$ | Singh et al., 2002 |


| Table 3: Nonane - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of nonane | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vitro human cadaver skin In vitro pig skin |  | $\begin{gathered} \times 10^{-3} \\ 0.637 \pm 0.058 \\ 0.477 \pm 0.025 \end{gathered}$ | $\begin{aligned} & \times 10^{-3} \\ & 0.0724 \\ & 0.0541 \end{aligned}$ | Kanikkannan et al., $2001^{\circ}$ |
| Jet fuel (JP-8) | In vitro rat skin |  | $\begin{gathered} \times 10^{-3} \\ 0.384 \pm 0.240 \end{gathered}$ | $\begin{aligned} & \times 10^{-3} \\ & 0.042 \end{aligned}$ | $\begin{aligned} & \text { McDougal et al, } \\ & 2000 \end{aligned}$ |
| Jet fuel (JP-8) | In vitro pig skin After 1 day of pre-exposure Control Pre-exposed to jet fuel (JP-8) After 4 day of pre-exposure Control Pre-exposed to jet fuel (JP-8) |  | $\begin{gathered} \times 10^{-3} \\ 0.03 \pm 0.01 \\ 0.08 \pm 0.01 \\ \\ 0.03 \pm 0.01 \\ 0.07 \pm 0.01 \\ \hline \end{gathered}$ | $x 10^{-3}$ $0.002 \pm 0.001$ $0.005 \pm 0.001$ $0.002 \pm 0.000$ $0.005 \pm 0.000$ | $\begin{aligned} & \text { Muhammad et al, } \\ & 2005 \end{aligned}$ |
| Jet fuel (JP-8) | In vitro pig ear skin JP-8 + 100 <br> JP-8 + BHT <br> JP-8 + MDA <br> JP-8 + 8Q405 |  | $\begin{gathered} \times 10^{-3} \\ 0.395 \pm 0.007 \\ 0.396 \pm 0.014 \\ 0.451 \pm 0.031 \\ 0.461 \pm 0.033 \end{gathered}$ | $\begin{gathered} \times 10^{-4} \\ 0.4489 \end{gathered}$ | Kanikkannan et al., 2001b |

Table 4: Decane - summary

| Table 4: Decane - summary |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :---: |
| Application of <br> decane | In vivo/in vitro <br> animal-human | \% dose <br> absorbed | Flux <br> $\left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right)$ | Kp <br> $(\mathrm{cm} / \mathrm{h})$ |  |  |
| Jet fuel (JP-8) | In vivo human volunteers |  |  | $\times 10^{-5}$ | $0.65 \pm 0.33$ |  |

\begin{tabular}{|c|c|c|c|c|c|}
\hline \multicolumn{6}{|l|}{Table 5: Undecane - summary} \\
\hline Application of undecane \& In vivo/in vitro animal-human \& \% dose absorbed \& \[
\begin{gathered}
\text { Flux } \\
\left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right)
\end{gathered}
\] \& \[
\begin{gathered}
\mathrm{Kp} \\
(\mathrm{~cm} / \mathrm{h})
\end{gathered}
\] \& Study \\
\hline Jet fuel (JP-8) \& In vitro pig skin dose levels: \(1 x\) dose \(2 x\) dose \(5 x\) dose \& \& \[
\begin{gathered}
\times 10^{-3} \\
0.03 \pm 0.00 \\
0.03 \pm 0.00 \\
0.04 \pm 0.01
\end{gathered}
\] \& \(\times 10^{-3}\)
\(0.0002 \pm 0.0000\)
\(0.0002 \pm 0.0000\)
\(0.0003 \pm 0.0000\) \& Muhammad et al., 2004 \\
\hline Jet fuel (JP-8) \& In vivo human volunteers \& \& \& \[
\begin{gathered}
\times 10^{-5} \\
0.045 \pm 0.023 \\
\hline
\end{gathered}
\] \& Kim et al., 2006 \\
\hline Jet fuel (JP-8) \& In vitro rat skin \& \& \[
\begin{gathered}
\times 10^{-3} \\
1.22 \pm 0.81
\end{gathered}
\] \& \[
\begin{aligned}
\& \times 10^{-3} \\
\& 0.025 \\
\& \hline
\end{aligned}
\] \& \[
\begin{aligned}
\& \text { McDougal et al., } \\
\& 2000
\end{aligned}
\] \\
\hline Jet fuel (JP-8) \& \begin{tabular}{l}
In vitro pig skin \\
After 1 day of pre-exposure \\
Control \\
Pre-exposed to jet fuel (JP-8) \\
After 4 day of pre-exposure \\
Control \\
Pre-exposed to jet fuel (JP-8)
\end{tabular} \& \& \(\mathrm{x} \mathrm{10-3}\)
\(0.07 \pm 0.01\)
\(0.16 \pm 0.05\)

$0.06 \pm 0.01$
$0.10 \pm 0.02$ \& $\mathrm{x} 10^{-3}$
$0.001 \pm 0.000$
$0.003 \pm 0.000$
$0.001 \pm 0.000$
$0.002 \pm 0.000$ \& Muhammad et al., 2005 <br>
\hline
\end{tabular}

| Table 6: Dodecane - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of dodecane | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vitro pig skin dose levels: <br> 1x dose <br> $2 x$ dose <br> $5 x$ dose |  | $\begin{gathered} \times 10^{-3} \\ 0.01 \pm 0.00 \\ 0.02 \pm 0.00 \\ 0.03 \pm 0.01 \end{gathered}$ | $\times 10^{-3}$ $0.0003 \pm 0.0001$ $0.0002 \pm 0.0001$ $0.0001 \pm 0.0000$ | Muhammad et al., 2004 |
| Jet fuel (JP-8) | In vivo human volunteers |  |  | $\begin{gathered} \times 10^{-5} \\ 0.16 \pm 0.056 \end{gathered}$ | Kim et al., 2006 |
| Jet fuel (JP-8) | In vitro rat skin |  | $\begin{gathered} \times 10^{-3} \\ 0.510 \pm 0.363 \end{gathered}$ | $\begin{aligned} & \times 10^{-3} \\ & 0.014 \end{aligned}$ | $\begin{aligned} & \text { McDougal et al., } \\ & 2000 \end{aligned}$ |

\begin{tabular}{|c|c|c|c|c|c|}
\hline \multicolumn{6}{|l|}{Table 6: Dodecane - summary} \\
\hline Various Jet fuels \& \begin{tabular}{l}
In vitro pig skin Jet fuel (JP-8) \\
Jet fuel (JP-A) \\
Jet fuel (JP-8) puddle \\
Jet fuel (JP-8(100))
\end{tabular} \& \[
\begin{aligned}
\& 0.63 \pm 0.04 \\
\& 0.29 \pm 0.04 \\
\& 0.27 \pm 0.07 \\
\& 0.35 \pm 0.04
\end{aligned}
\] \& \& \& \[
\begin{aligned}
\& \text { Riviere et al., } \\
\& 1999
\end{aligned}
\] \\
\hline Jet fuel (JP-8) \& \begin{tabular}{l}
In vitro pig skin \\
After 1 day of pre-exposure \\
Control \\
Pre-exposed to jet fuel (JP-8) \\
After 4 day of pre-exposure Control \\
Pre-exposed to jet fuel (JP-8)
\end{tabular} \& \& \(\times 10^{-3}\)
\(0.04 \pm 0.01\)
\(0.06 \pm 0.02\)
\(0.02 \pm 0.00\)
\(0.05 \pm 0.01\) \& \(\times 10^{-3}\)
\(0.0005 \pm 0.000\)
\(0.0009 \pm 0.000\)
\(0.0002 \pm 0.000\)
\(0.0008 \pm 0.000\) \& al., 2005 \\
\hline Jet fuel (JP-8) with various additives \& ```
In vitro pig skin
Jet-A
Jet-A + DIEGME
Jet-A + 8Q21
Jet-A + Stadis450
Jet-A + DIEGME + 8Q21
Jet-A + DIEGME + Stadis450
Jet-A + 8Q21 + Stadis 450
JP-8
``` \& \& \[
\begin{gathered}
x 10^{-4} \\
0.10 \pm 0.01 \\
0.06 \pm 0.003 \\
0.05 \pm 0.01 \\
0.10 \pm 0.01 \\
0.19 \pm 0.02 \\
0.25 \pm 0.05 \\
0.30 \pm 0.02 \\
0.09 \pm 0.01 \\
\hline
\end{gathered}
\] \& \(\times 10^{-4}\)
\(0.11 \pm 0.01\)
\(0.09 \pm 0.01\)
\(0.07 \pm 0.01\)
\(0.15 \pm 0.02\)
\(0.12 \pm 0.01\)
\(0.15 \pm 0.03\)
\(0.18 \pm 0.01\)
\(0.09 \pm 0.01\) \& \[
\begin{aligned}
\& \hline \begin{array}{l}
\text { Baynes } \\
2001
\end{array} \text { et al., }
\end{aligned}
\] \\
\hline Jet fuel (JP-8) with various additives \& ```
In vitro silastic membrane
JP-8 ( \(\mathrm{n}=5\) )
JP-8 + MDA ( \(\mathrm{n}=5\) )
JP-8 + BHT ( \(\mathrm{n}=5\) )
JP-8 + 8Q405 ( \(\mathrm{n}=4\) )
JP-8 + MDA + BHT ( \(n=5\) )
JP-8 + MDA + 8Q405 ( \(\mathrm{n}=4\) )
JP-8 + BHT + 8Q405 ( \(n=4\) )
JP-8(100) ( \(\mathrm{n}=5\) )
In vitro pig skin
JP-8 ( \(\mathrm{n}=5\) )
JP-8 + MDA \((\mathrm{n}=5)\)
\(\mathrm{JP}-8+\mathrm{BHT}(\mathrm{n}=5)\)
JP-8 + 8Q405 ( \(\mathrm{n}=4\) )
\(\mathrm{JP}-8+\mathrm{MDA}+\mathrm{BHT}(\mathrm{n}=5)\)
JP-8 + MDA + 8Q405 ( \(n=4\) )
``` \& \& \[
\begin{gathered}
x 10^{-4} \\
1.46 \pm 0.08 \\
0.70 \pm 0.03 \\
0.90 \pm 0.07 \\
0.75 \pm 0.04 \\
0.89 \pm 0.15 \\
0.8 \pm \pm 0.08 \\
0.74 \pm 0.24 \\
1.34 \pm 0.05 \\
\\
\\
0.090 \pm 0.01 \\
0.164 \pm 0.03 \\
0.123 \pm 0.01 \\
0.171 \pm 0.05 \\
0.077 \pm 0.01 \\
0.097 \pm 0.01
\end{gathered}
\] \& \(\times 10^{-4}\)
\(0.041 \pm 0.002\)
\(0.020 \pm 0.001\)
\(0.026 \pm 0.002\)
\(0.021 \pm 0.001\)
\(0.025 \pm 0.004\)
\(0.024 \pm 0.002\)
\(0.021 \pm 0.007\)
\(0.038 \pm 0.001\)

$0.0025 \pm 0.00$
$0.0047 \pm 0.00$
$0.0035 \pm \pm 000$
$0.0049 \pm 0.00$
$0.0022 \pm 0.00$

$0.0028 \pm 0.00$ \& $$
\begin{aligned}
& \hline \text { Muhammad et } \\
& \text { al., } 2004
\end{aligned}
$$ <br>

\hline
\end{tabular}

Table 6: Dodecane - summary

| Table 6: Dodecane - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \text { JP-8 + BHT + 8Q405 }(n=4) \\ & \text { JP-8(100) }(n=5) \end{aligned}$ |  | $\begin{aligned} & 0.079 \pm 0.01 \\ & 0.094 \pm 0.02 \end{aligned}$ | $\begin{aligned} & 0.0022 \pm 0.00 \\ & 0.0027 \pm 0.00 \end{aligned}$ |  |
| Table 7: Tridecane - summary |  |  |  |  |  |
| Application of tridecane | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vitro pig skin dose levels: <br> $1 x$ dose <br> $2 x$ dose <br> $5 x$ dose |  | $\times 10^{-3}$ $0.004 \pm 0.00$ $0.006 \pm 0.00$ $0.008 \pm 0.00$ | $\times 10^{-3}$ $0.0001 \pm 0.000$ $0.0001 \pm 0.000$ $0.0001 \pm 0.000$ | Muhammad et al., 2004 |
| Jet fuel (JP-8) | In vitro human skin In vitro pig skin |  | $\begin{gathered} \times 10^{-3} \\ 1.447 \pm 0.154 \\ 1.508 \pm 0.188 \\ \hline \end{gathered}$ |  | Kanikkannan et al., 2001a |
| Jet fuel (JP-8) | In vitro rat skin |  | $\begin{gathered} \times 10^{-3} \\ 0.334 \pm 0.194 \\ \hline \end{gathered}$ | $\begin{aligned} & \times 10^{-3} \\ & 0.015 \end{aligned}$ | $\begin{aligned} & \text { McDougal et al., } \\ & 2000 \end{aligned}$ |
| Jet fuel (JP-8) | In vitro pig skin <br> After 1 day of pre-exposure Control Pre-exposed to jet fuel (JP-8) <br> After 4 day of pre-exposure Control Pre-exposed to jet fuel (JP-8) |  | $\begin{gathered} \times 10^{-3} \\ 0.02 \pm 0.01 \\ 0.01 \pm 0.00 \\ \\ 0.00 \pm 0.00 \\ 0.02 \pm 0.01 \\ \hline \end{gathered}$ | $\mathrm{x} \mathrm{10}^{-3}$ $0.0003 \pm 0.000$ $0.0002 \pm 0.000$ $0.0001 \pm 0.000$ $0.0003 \pm 0.000$ | Muhammad et al., 2005 |
| Jet fuel (JP-8) | In vitro pig ear skin $\begin{aligned} & \text { JP-8 + } 100 \\ & \text { JP-8 + BHT } \\ & \text { JP-8 + MDA } \\ & \text { JP-8 + 8Q405 } \end{aligned}$ |  | $\times 10^{-3}$ $1.318 \pm 0.155$ $1.223 \pm 0.059$ $1.530 \pm 0.111$ $1.465 \pm 0.093$ | $\begin{aligned} & \times 10^{-4} \\ & 0.6102 \end{aligned}$ | Kanikkannan et al., 2001b |


| Table 8: Hexadecane - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of hexadecane | In vivo/in vitro animal-human | \% dose absorbed | Flux $\left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right)$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vivo pigs | 0.34 |  |  | Singh et al., 2003 |
| Jet fuel (JP-8) | In vitro pig skin Jet fuel (JP-8) | $0.18 \pm 0.08$ |  |  | $\begin{aligned} & \text { Riviere et al., } \\ & 1999 \end{aligned}$ |
| Jet fuel (JP-8) | In vitro human cadaver skin In vitro pig skin |  | $\begin{gathered} \times 10^{-6} \\ 1.586 \pm 0.000 \\ 1.980 \pm 0.000 \end{gathered}$ | $\begin{gathered} \times 10^{-5} \\ 3.60 \pm 0.00 \\ 4.60 \pm 0.00 \end{gathered}$ | Singh et al., 2002 |

Aromatic hydrocarbons

| Table 9: Benzene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of benzene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Benzene vapour in air | In vivo, rats |  | 0.0191 | $0.152 \pm 0.006$ | $\begin{aligned} & \text { McDougal et al, } \\ & 1990 \end{aligned}$ |
| Benzene in water | In vitro, human <br> Donor solution $\left(26^{\circ} \mathrm{C}\right)$ <br> Donor solution $\left(50^{\circ} \mathrm{C}\right)$ <br> Donor solution $\left(40^{\circ} \mathrm{C}\right)$ <br> Donor solution ( $15^{\circ} \mathrm{C}$ ) |  |  | $\begin{aligned} & 0.14 \pm 0.01 \\ & 0.26 \pm 0.05 \\ & 0.18 \pm 0.03 \\ & 0.10 \pm 0.04 \end{aligned}$ | Nakai et al, 1997 |
| Benzene in water | In vivo, rats <br> Previously frozen skin <br> Skin pre-treated with moisturizer <br> Skin pre-treated with baby oil <br> Skin pre-treated with insect repellant <br> Skin pre-treated with sunscreen |  |  | $\begin{aligned} & 0.18 \pm 0.02 \\ & 0.14 \pm 0.03 \\ & 0.16 \pm 0.02 \\ & 0.18 \pm 0.02 \\ & 0.24 \pm 0.04 \end{aligned}$ | Nakai et al, 1997 |
| Benzene in toluene Benzene in water | In vitro human skin <br> Benzene in toluene (0.01-0.50 \%) <br> Receptor fluid <br> Epidermis <br> Dermis <br> Benzene in water (0.10-0.50 \%) <br> Receptor fluid <br> Epidermis <br> Dermis | $\begin{aligned} & 0.08-0.12 \\ & 0.03-0.04 \\ & 0.01-0.01 \\ & 3.88-5.03 \\ & 0.36-0.42 \\ & 0.27-0.41 \\ & \hline \end{aligned}$ |  |  | Wester 200 Maibach, 2000 |
| Benzene in different vehicles | In vitro, human <br> Vehicle: <br> Benzene <br> Air <br> Water <br> Hexane <br> Hexadecane <br> Isooctane <br> Gasoline |  | $\begin{gathered} 1.861 \pm 0.95 \\ 0.92 \pm 0.33 \\ 0.194 \pm 0.044 \\ 0.106 \\ 0.044 \\ 0.167 \\ 0.062 \\ \hline \end{gathered}$ | $\begin{gathered} \times 10^{-3} \\ \\ 111.1 \pm 25.9 \\ 2.4 \\ 0.94 \pm 0.38 \\ 3.73 \pm 1.26 \\ 1.40 \pm 0.58 \end{gathered}$ | Blank et al, 1985 <br> Flux calculated using density of benzene to be $0.81 \mathrm{~g} / \mathrm{ml}$ |
| Neat benzene | In vitro, rats |  | 0.57 |  | Ahaghotu et al, 2005 |
| Gasoline | In vitro, human Gasoline 1 Gasoline 2 | $\begin{aligned} & 0.49 \\ & 0.63 \\ & \hline \end{aligned}$ | $\begin{gathered} \times 10^{-3} \\ 2.71 \pm 1.62 \\ 1.80 \pm 1.11 \end{gathered}$ | $\begin{gathered} \times 10^{-4} \\ 4.95 \\ 6.35 \end{gathered}$ | Adami et al, 2006 |


| Table 9: Benzene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Gasoline 3 <br> Mean $\pm$ SD | $\begin{gathered} 0.19 \\ 0.43 \pm 0.23 \end{gathered}$ | $\begin{aligned} & 1.47 \pm 0.53 \\ & 1.99 \pm 0.64 \end{aligned}$ | $\begin{gathered} 1.88 \\ 4.34 \pm 2.28 \end{gathered}$ |  |
| Table 10: Ethylbenzene - summary |  |  |  |  |  |
| Application of benzene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vitro pig skin <br> After 1 day of pre-exosure <br> Control <br> Pre-exposed to jet fuel (JP-8) <br> After 4 day of pre-exosure Control <br> Pre-exposed to jet fuel (JP-8) |  | $\begin{gathered} x 10^{-3} \\ 1.04 \pm 0.17 \\ 3.32 \pm 0.52 \\ \\ 0.61 \pm 0.15 \\ 2.04 \pm 0.17 \\ \hline \end{gathered}$ | $\begin{gathered} x 10^{-3} \\ 0.06 \pm 0.009 \\ 0.19 \pm 0.03 \\ \\ 0.035 \pm 0.009 \\ 0.12 \pm 0.009 \\ \hline \end{gathered}$ | Muhammad et al., 2005 |


| Table 11: Trimethylbenzene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of benzene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \\ \hline \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vitro pig skin <br> After 1 day of pre-exosure <br> Control <br> Pre-exposed to jet fuel (JP-8) <br> After 4 day of pre-exosure Control <br> Pre-exposed to jet fuel (JP-8) |  | $\begin{gathered} \times 10^{-3} \\ \\ 1.01 \pm 0.14 \\ 1.77 \pm 0.21 \\ \\ \\ 0.49 \pm 0.04 \\ 1.52 \pm 0.10 \\ \hline \end{gathered}$ | $\begin{gathered} x 10^{-3} \\ 0.056 \pm 0.008 \\ 0.10 \pm 0.01 \\ \\ 0.028 \pm 0.002 \\ 0.09 \pm 0.005 \\ \hline \end{gathered}$ | Muhammad et al., 2005 |


| Table 12: Toluene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of toluene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Aqueous toluene | In vivo, human volunteers Dermal only |  |  | $0.0115 \pm 0.0074$ | Thrall et al, 2002 |
| Neat toluene | In vitro, rats |  | 0.38 |  | Ahaghotu et al, 2005 |
| Toluene vapour mixture | In vivo, rats |  | 0.0206 | $0.721 \pm 0.007$ | McDougal et al, 1990 |
| Toluene | In vivo, guinea pigs | 16.5 |  |  | $\begin{aligned} & \text { Boman et al, } \\ & 1995 \end{aligned}$ |
| Toluene vapours | In vivo, human volunteers |  |  | $0.050 \pm 0.023$ | Kezic et al, 2000 |

Table 12: Toluene - summary

| Table 12: Toluene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Aqueous toluene | In vivo, rats | $43.8 \pm 9.6$ |  | $0.074 \pm 0.005$ | Thrall and Woodstock, 2002 |
| Jet fuel (JP-8) | In vitro, rat skin |  | $\begin{gathered} \times 10^{-3} \\ 0.535 \pm 0.094 \end{gathered}$ | $\begin{array}{r} \times 10^{-3} \\ 1.1 \\ \hline \end{array}$ | McDougal et al, 2000 |
| Jet fuel (JP-8) | In vitro, pig ear skin In vitro, human cadaver skin |  | $\begin{gathered} \times 10^{-3} \\ 0.119 \pm 0.004 \\ 0.095 \pm 0.009 \end{gathered}$ | $\begin{gathered} \times 10^{-4} \\ 2.47 \\ 1.97 \end{gathered}$ | Kanikkannan et al, 2001 |
| Gasoline | In vitro, human Gasoline 1 Gasoline 2 Gasoline 3 Mean $\pm$ SD | $\begin{gathered} 0.07 \\ 0.08 \\ 0.04 \\ 0.06 \pm 0.02 \\ \hline \end{gathered}$ | $\begin{gathered} \times 10^{-3} \\ 5.74 \pm 2.77 \\ 3.60 \pm 2.25 \\ 2.07 \pm 1.11 \\ 3.80 \pm 1.84 \end{gathered}$ | $\times 10^{-4}$ 0.722 0.82 0.403 $0.648 \pm 0.218$ | Adami et al, 2006 |
| Jet fuel (JP-8) | In vitro pig ear skin JP-8 + 100 <br> JP-8 + BHT <br> JP-8 + MDA <br> JP-8 + 8Q405 |  | $\times 10^{-3}$ $0.094 \pm 0.001$ $0.071 \pm 0.013$ $0.114 \pm 0.009$ $0.117 \pm 0.005$ | $\begin{aligned} & \times 10^{-4} \\ & 1.958 \end{aligned}$ | Kanikkannan et al., 2001b |
| Neat toluene | In vivo human volunteers |  | $\begin{gathered} \times 10^{-3} \\ 11.64 \pm 4.23 \\ \hline \end{gathered}$ |  | Kezic et al., 2001 |
| Toluene radiolabeled | In vivo, albino hairless mice | $15.4 \pm 2.0$ | $2.94 \pm 2.27$ |  | Susten et al, 1990 |
| Toluene in various solvents | ```In vitro, human Ventilation (ml/min), neat toluene O 90 400 900 Ventilation (ml/min), Butanol O 90 Ventilation (ml/min) Chloroform/MeOH 0 90``` | $\begin{gathered} 2.3 \pm 0.4 \\ 0.8 \pm 0.2 \\ 0.5 \pm 0.1 \\ 0.2 \pm 0.03 \\ \\ 6.6 \pm 0.2 \\ 0.7 \pm 0.3 \\ \\ \\ \\ \\ 1.1 \pm 0.5 \\ 1.1 \pm 0.3 \end{gathered}$ |  |  | Boman and Maibach, 2000 |
| Toluene in various vehicles | In vivo, mice <br> Vehicle: <br> Toluene <br> Methanol <br> Ethanol <br> 1-Propercutaneous absorptionnol |  |  | $\begin{gathered} x 10^{-3} \\ 0.0792 \\ 0.5904 \\ 0.1590 \\ 0.0714 \end{gathered}$ | Tsuruta et al, 1996 |

Table 12: Toluene - summary

| Table 12: Toluene - sumpary |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
|  | Isobutanol |  | 0.0714 |
|  | 1-Pentanol |  | 0.0864 |
|  | 1-Ooctanol |  | 0.0930 |
|  | 2-Metoxyethanol |  | 0.0990 |
|  | 2-Butoxyethanol |  | 0.1068 |
|  | Benzyl alcohol |  | 0.0834 |
|  | Cyclohexanol |  | 0.0768 |
|  | Ethylene glycol |  | 0.0564 |
|  | Propylene glycol |  | 0.1404 |
|  | Glycerol |  | 0.0864 |
|  | Ether |  | 0.1170 |
|  | Acetone |  | 0.1146 |
|  | DMSO |  | 0.714 |
|  | N,N-Dimethylacetamide |  | 0.3222 |
|  | N,N-Dimethylformamide |  | 0.5124 |
|  | Benzene |  |  |


| Table 13: Xylene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of xylene | In vivo/in vitro animal-human | \% dose absorbed | $\underset{\left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right)}{\text { Flux }}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Xylene | In vitro, perfused pig ear Whole blood Blood-WBC <br> Plasma <br> Buffer + BSA <br> Buffer - BSA |  | $\begin{aligned} & 0.0211 \pm 0.0028 \\ & 0.0188 \pm 0.0021 \\ & 0.0954 \pm 0.0191 \\ & 0.0143 \pm 0.0035 \\ & 0.0024 \pm 0.0006 \end{aligned}$ |  | deLange et al, 1994 |
| m-xylene vapour | In vivo, rats |  | 0.0151 | $0.723 \pm 0.003$ | McDougal et al, 1990 |
| Aqueous o-xylene | In vivo, rats In vivo, human volunteers |  |  | $\begin{aligned} & 0.058 \pm 0.009 \\ & 0.005 \pm 0.001 \end{aligned}$ | Thrall and Woodstock, 2003 |
| m-xylene vapour | In vivo, human volunteers |  |  | $0.025 \pm 0.012$ | Kezic et al, 2000 |
| Neat m-xylene | In vivo, human volunteers |  | $\begin{gathered} \times 10^{-3} \\ 2.40 \pm 0.89 \end{gathered}$ |  | Kezic et al, 2001 |
| Jet fuel (JP-8) | In vitro human cadaver skin In vitro pig skin |  | $\begin{gathered} \times 10^{-6} \\ 2.211 \pm 0.021 \\ 2.569 \pm 0312 \end{gathered}$ | $\begin{gathered} \times 10^{-5} \\ 8.33 \pm 0.01 \\ 9.68 \pm 0.01 \\ \hline \end{gathered}$ | Singh et al, 2002 |
| Gasoline | In vitro, human Gasoline 1 Gasoline 2 | $\begin{aligned} & 0.01 \\ & 0.01 \\ & \hline \end{aligned}$ | $\begin{gathered} \times 10^{-3} \\ 1.01 \pm 0.59 \\ 0.50 \pm 0.25 \end{gathered}$ | $\begin{aligned} & \times 10^{-4} \\ & 0.097 \\ & 0.051 \end{aligned}$ | Adami et al, 2006 |


| Table 13: Xylene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Gasoline 3 Mean $\pm$ SD | $\begin{gathered} 0.01 \\ 0.008 \pm 0.003 \end{gathered}$ | $\begin{aligned} & 0.71 \pm 0.56 \\ & 2.38 \pm 0.17 \\ & \hline \end{aligned}$ | $\begin{gathered} 0.104 \\ 0.084 \pm 0.029 \end{gathered}$ |  |
| Jet fuel (JP-8) | In vitro pig skin <br> After 1 day of pre-exosure Control Pre-exposed to jet fuel (JP-8) <br> After 4 day of pre-exosure Control Pre-exposed to jet fuel (JP-8) |  | $\times 10^{-3}$ $1.47 \pm 0.20$ $3.80 \pm 0.61$ $1.02 \pm 0.28$ $3.13 \pm 0.15$ | $\times 10^{-3}$ $0.085 \pm 0.011$ $0.218 \pm 0.040$ $0.059 \pm 0.016$ $0.180 \pm 0.009$ | Muhammad et al., 2005 |
| m-xylene vapour | In vivo, human volunteers 20 min exposure 45 min exposure 120 min exposure 180 min exposure |  | $\begin{gathered} \times 10^{-4} \\ 0.34 \pm 0.12 \\ 0.42 \pm 0.14 \\ 0.59 \pm 0.16 \\ 0.63 \pm 0.14 \end{gathered}$ | $\begin{aligned} & 0.059 \pm 0.016 \\ & 0.063 \pm 0.014 \\ & \hline \end{aligned}$ | Kezic et al, 2004 |
| Jet fuel (JP-8) | In vivo pigs | 0.12 |  |  | Singh et al., 2003 |
| Jet fuel (JP-8) | In vitro rat skin |  | $\begin{gathered} \times 10^{-3} \\ 0.795 \pm 0.238 \end{gathered}$ | $\begin{gathered} \times 10^{-3} \\ 0.17 \\ \hline \end{gathered}$ | McDougal et al., $2000$ |
| Neat xylene | In vitro, rats |  | 0.22 |  | Ahaghotu et al, 2005 |
| m-xylene in ethanol | In vivo, rats m -xylene alone m-xylene + sandy soil m-xylene + clay soil | $\begin{aligned} & 0.23 \pm 0.03 \\ & 0.15 \pm 0.03 \\ & 0.26 \pm 0.02 \end{aligned}$ |  |  | $\begin{aligned} & \text { Skowronski et al, } \\ & 1990 \end{aligned}$ |


| Table 14: Benzo[a]pyrene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of benzo[a]pyrene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Petroleum crude oil | In vivo rats crude oil crude oil in soil <br> In vitro rat skin crude oil crude oil in soil | $\begin{gathered} 35.3 \pm 2.6 \\ 9.2 \pm 1.2 \\ \\ 38.0 \pm 3.0 \\ 8.5 \pm 1.0 \end{gathered}$ |  |  | Yang et al., 1989 |
| Industrial coal-tar | In vitro pig ear skin |  | $\begin{aligned} & \times 10^{-4} \\ & 0.008 \end{aligned}$ |  | Van Rooij et al., |
| PAHs in contaminated soil | In vitro cadaver skin benzo[a]pyrene in soil at low dose benzo[a]pyrene in soil at medium dose |  | $\begin{gathered} \times 10^{-6} \\ 0.025 \\ 0.19 \end{gathered}$ |  | Roy et al., 1998 |

Table 14: Benzo[a]pyrene - summary

| Table 14: Benzo[a]pyrene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | benzo[a]pyrene in soil at high dose benzo[a]pyrene in soil extract at high dose |  | $\begin{aligned} & 1.00 \\ & 210 \end{aligned}$ |  |  |
| Mixture of PAHs | In vitro monkey skin Lubrication oil Acetone |  |  | $\begin{gathered} \times 10^{-3} \\ 0.23 \pm 0.20 \\ \hline \end{gathered}$ | Sartorelli et al., 1999 |
| Coal dust with acetone | In vitro human cadaver skin | $8.57 \pm 3.67$ |  |  | Sartorelli et al., 2001 |
| Native PAHs in Lampsblack in soil | In vitro human cadaver skin Concentration level (mg/kg) 38 <br> 111 <br> 135 <br> 632 <br> 817 <br> 915 <br> 1702 |  | $x 10^{-6}$ $0.05 \pm 0.08$ $0.05 \pm 0.06$ $0.10 \pm 0.03$ $0.20 \pm 0.08$ $0.30 \pm 0.08$ $0.20 \pm 0.08$ $0.30 \pm 0.08$ |  | Stroo et al., 2005 |


| Table 15: Pyrene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of pyrene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Industrial coal-tar | In vitro pig ear skin |  | $\begin{aligned} & \times 10^{-4} \\ & 0.008 \end{aligned}$ |  | Van Rooij et al., |
| Mixture of PAHs | In vitro monkey skin Lubrication oil Acetone |  |  | $\begin{gathered} \times 10^{-3} \\ 0.17 \pm 0.04 \\ 4.13 \pm 4.36 \\ \hline \end{gathered}$ | Sartorelli et al., 1999 |
| Coal dust with acetone | In vitro human cadaver skin | $51.98 \pm 14.97$ |  |  | $\begin{aligned} & \text { Sartorelli et al., } \\ & 2001 \end{aligned}$ |


| Table 16: Naphthalene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of naphthalene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Radiolabelled naphthalene | In vivo mice Naphthalene alone Naphthalenein sandy soil Naphthalene in clay soil | $\begin{aligned} & 0.50 \pm 0.04 \\ & 0.42 \pm 0.03 \\ & 0.63 \pm 0.03 \end{aligned}$ |  |  | Turkall et al., 1994 |
| Mixture of PAHs | In vitro monkey skin Lubrication oil acetone |  |  | $\begin{gathered} \times 10^{-3} \\ 1.87 \pm 1.31 \\ 6.31 \pm 2.49 \\ \hline \end{gathered}$ | Sartorelli et al., 1999 |

Table 16: Naphthalene - summary

| Table 16: Naphthalene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Jet fuel (JP-8) | In vivo human volunteers |  |  | $\begin{gathered} \times 10^{-3} \\ 0.053 \pm 0.038 \\ \hline \end{gathered}$ | Kim et al., 2006 |
| Mixture of aromatic and aliphatic hydrocarbons in hexadecane | In vitro pig skin Dose level: 1xdose 2xdose $3 x d o s e$ |  | $\begin{gathered} \times 10^{-3} \\ 0.43 \pm 0.12 \\ 1.24 \pm 0.26 \\ 3.63 \pm 0.24 \end{gathered}$ | $\times 10^{-3}$ $0.033 \pm 0.009$ $0.0485 \pm 0.0101$ $0.0569 \pm 0.0066$ | Muhammad et al., 2004 |
| Jet fuel (JP-8) | In vitro pig skin |  | $\begin{gathered} \times 10^{-3} \\ 0.376 \pm 0.017 \end{gathered}$ | $\begin{aligned} & \times 10^{-3} \\ & 0.181 \end{aligned}$ | Kanikkannan et al., 2001 |
| Jet fuel (JP-8) | In vitro rat skin |  | $\begin{gathered} \times 10^{-3} \\ 1.04 \pm 0.38 \end{gathered}$ | $\begin{gathered} \times 10^{-3} \\ 0.51 \end{gathered}$ | McDougal et al., 2000 |
| Various jet fuels | ```In vitro porcine (pig) skin JP-8 Jet-A JP-8 (Puddle) JP-8 (100)``` | $\begin{aligned} & 1.17 \pm 0.07 \\ & 1.49 \pm 0.18 \\ & 1.11 \pm 0.16 \\ & 1.63 \pm 0.29 \\ & \hline \end{aligned}$ |  |  | Riviere et al., 1999 |
| Jet fuel (JP-8) with various additives | ```In vitro porcine (pig) skin Jet-A Jet-A + DIEGME Jet-A + 8Q21 Jet-A + Stadis450 Jet-A + DIEGME + 8Q21 Jet-A + DIEGME + Stadis450 Jet-A + 8Q21 + Stadis450 JP-8``` |  | $x 10^{-3}$ $0.222 \pm 0.012$ $0.318 \pm 0.036$ $0.264 \pm 0.012$ $0.276 \pm 0.012$ $0.124 \pm 0.008$ $0.107 \pm 0.014$ $0.097 \pm 0.003$ $0.186 \pm 006$ | $x 10^{-3}$ $0.168 \pm 0.006$ $0.198 \pm 0.024$ $0.168 \pm 0.006$ $0.174 \pm 0.006$ $0.135 \pm 0.010$ $0.113 \pm 0.015$ $0.097 \pm 0.003$ $0.144 \pm 0.006$ | Baynes et al., 2001 |
| Jet fuel (JP-8(100)) with various additives | $\begin{aligned} & \text { In vitro pig skin and silastic } \\ & \text { membrane } \\ & \text { silastic membrane } \\ & \text { JP-8 }(n=5) \\ & \text { JP-8 + MDA }(n=5) \\ & \text { JP-8 + BHT }(n=5) \\ & \text { JP-8 + 8Q405 }(n=4) \\ & \text { JP-8 + MDA }+ \text { BHT }(n=5) \\ & \text { JP-8 + MDA }+8 Q 405(n=4) \\ & \text { JP-8 + BHT + 8Q405 }(n=4) \\ & \text { JP-8 }(100)(n=5) \\ & \text { pig skin } \\ & \text { JP-8 }(n=5) \\ & \text { JP-8 + MDA }(n=5) \\ & \text { JP-8 + BHT }(n=5) \\ & \text { JP-8 + 8Q405 }(n=4) \\ & \hline \end{aligned}$ |  | $\begin{gathered} \times 10^{-3} \\ 19.29 \pm 0.60 \\ 28.23 \pm 1.62 \\ 27.17 \pm 0.84 \\ 26.86 \pm 0.71 \\ 21.22 \pm 0.52 \\ 21.65 \pm 0.62 \\ 21.75 \pm 0.73 \\ 21.01 \pm 0.86 \\ \\ 2.21 \pm 0.27 \\ 2.63 \pm 0.09 \\ 2.48 \pm 0.28 \\ 2.17 \pm 0.16 \\ 2.88 \pm 0.35 \\ \hline \end{gathered}$ | $x+10^{-3}$ $1.78 \pm 0.06$ $2.30 \pm 0.13$ $2.22 \pm 0.07$ $2.19 \pm 0.06$ $1.99 \pm 0.05$ $2.03 \pm 0.06$ $2.04 \pm 0.07$ $1.93 \pm 0.08$ $0.21 \pm 0.03$ $0.20 \pm 0.01$ $0.19 \pm 0.02$ $0.17 \pm 0.01$ $0.23 \pm 0.03$ | Muhammad et al., 2004 |



| Table 17: Methylnaphthalenes - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of methylnaphthalene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vivo human volunteers <br> 1-methylnaphthalene <br> 2-methylnaphthalene |  |  | $\begin{gathered} \times 10^{-3} \\ 0.029 \pm 0.0059 \\ 0.032 \pm 0.0074 \\ \hline \end{gathered}$ | Kim et al., 2006 |
| Jet fuel (JP-8) | In vitro rat skin |  | $\begin{gathered} \times 10^{-3} \\ 1.55 \pm 0.52 \\ \hline \end{gathered}$ | $\begin{gathered} \times 10^{-3} \\ 0.16 \\ \hline \end{gathered}$ | McDougal et al., 2000 |


| Table 18: Dimethylnaphthalenes - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of dimethylnaphthalene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Mixture of aromatic and aliphatic hydrocarbons in hexadecane | In vitro pig skin 1xdose 2xdose 3xdose |  | $\begin{gathered} \times 10^{-3} \\ 0.13 \pm 0.01 \\ 0.23 \pm 0.05 \\ 0.58 \pm 0.09 \end{gathered}$ | $\begin{gathered} \times 10^{-3} \\ 0.0095 \pm 0.0007 \\ 0.0088 \pm 0.0020 \\ 0.0088 \pm 0.0014 \end{gathered}$ | Muhammad et al., 2004 |
| Jet fuel (JP-8) | In vitro rat skin |  | $\begin{gathered} \times 10^{-3} \\ 0.586 \pm 0.167 \end{gathered}$ | $\begin{gathered} \times 10^{-3} \\ 0.93 \\ \hline \end{gathered}$ | McDougal et al., $2000$ |
| Jet fuel (JP-8) | In vitro pigs <br> After 1 day of pre-exposure <br> Control <br> Pre-exposed to jet fuel (JP-8) |  | $\begin{gathered} \times 10^{-3} \\ 0.62 \pm 0.10 \\ 0.85 \pm 0.21 \end{gathered}$ | $\times 10^{-3}$ $0.02 \pm 0.004$ $0.03 \pm 0.008$ | Muhammad et al., 2005 |

Table 18: DimethyInaphthalenes - summary

| Table 18: Dimethylnaphthalenes - summary |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- |
|  | After 4 day of pre-exposure |  | $0.67 \pm 0.07$ | $0.03 \pm 0.003$ |
|  | Control | $0.99 \pm 0.09$ | $0.04 \pm 0.004$ |  |
|  | Pre-exposed to jet fuel (JP-8) |  |  |  |


| Table 19: Trimethylnaphthalenes - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of trimethylnaphthalene | In vivo/in vitro animal-human | \% dose absorbed | $\begin{gathered} \text { Flux } \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| Jet fuel (JP-8) | In vitro rat skin |  | $\begin{gathered} \times 10^{-3} \\ 1.25 \pm 0.50 \\ \hline \end{gathered}$ | $\begin{gathered} \times 10^{-3} \\ 0.13 \end{gathered}$ | McDougal et al., 2000 |


| Table 20: Tetramethylbenzene - summary |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Application of pyrene tetramethylnaphthale ne | In vivo/in vitro animal-human | \% dose | $\begin{aligned} & \text { Flux } \\ & \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{aligned}$ | $\begin{gathered} \mathrm{Kp} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Study |
| neat | In vitro rat skin |  | $<0.02$ |  | Ahaghotu et al., 2005 |

## APPENDIX 2 DETAILED DESCRIPTION OF EXPERIMENTAL CONDITIONS AND INTERPRETATION OF THE OUTCOMES



| Table 2 | Benzene, toluene, xylene, hexane |
| :--- | :--- |


| Authors |
| :--- |
| Title |


| Title | Dermal absorption of organic chemical vapors in rats and humans |
| :--- | :--- |
| Source | Fundamental and Applied Toxicology $14: 299-308,1990$ |


| Source | Fundamental and Applied Toxicology $14: 299-308,1990$ |  |  |  |
| :--- | :--- | :--- | :--- | :--- |
| Type / aim of | Test material/ species / | Exposure condition |  |  |



- In vivoln vivo $\quad$ - vapor of benzene,

| assessment of | toluene, xylene and |
| :--- | :--- |
| whole body | hexane | hexane

using masks to
respiratory exposure

| Permeability |  |
| :--- | :--- |
| ceefficient $(\mathrm{cm} / \mathrm{h})$ | Skin uptake (\%) |


| $0.723 \pm 0.003$ | 3.9 |
| :---: | :---: |
| $0.721 \pm 0.007$ | 3.7 |

${ }^{-}$
In general the results show that the permeability in rats is 2-4 times greater than in humans, based on literature data
Comment:
-The results for the individual compounds are calculated from the exposure to the mixture

| Table 2 Ben | Benzene, toluene, xylene, hexane |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors McD | McDougal et al. |  |  |  |  |  |
| Title $\quad$ Derm | Dermal absorption of organic chemical vapors in rats and humans |  |  |  |  |  |
| Source Fun | Fundamental and Applied Toxicology 14 :299-308, 1990 |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |
| - In vivoln vivo assessment of whole body dermal permeation of aliphatic and aromatic hydrocarbons. Comparison with published human data on vapor penetration | - vapor of benzene, toluene, xylene and hexane <br> -male Fisher-344 rats (205-273 g) <br> Analytical method: -GC-FID analysis | -whole body exposure to vapor mixture of chemicals using masks to prevent respiratory exposure <br> -exposure duration: <br> 4 h <br> Components measured: <br> -m-xylene ( $22.03 \mathrm{~g} / \mathrm{m}^{3}$ ) <br> -toluene ( $30.57 \mathrm{~g} / \mathrm{m}^{3}$ ) <br> -benzene ( $129.63 \mathrm{~g} / \mathrm{m}^{3}$ ) <br> -hexane ( $214.61 \mathrm{~g} / \mathrm{m}^{3}$ ) | -The vapor permeability was analyzed using physiologically based pharmacokinetic modeling according to Ramsey \& Andersen (1984) |  |  |  |
|  |  |  | Chemical | Flux (mg/cm ${ }^{2} / \mathrm{h}$ ) | Permeability coefficient (cm/h) | Skin uptake (\%) |
|  |  |  | m-xylene | 0.0151 | $0.723 \pm 0.003$ | 3.9 |
|  |  |  | toluene | 0.0206 | $0.721 \pm 0.007$ | 3.7 |
|  |  |  | benzene | 0.0191 | $0.152 \pm 0.006$ | 0.8 |
|  |  |  | hexane | 0.0065 | $0.031 \pm 0.004$ | 0.1 |
|  |  |  | In general the results show that the permeability in rats is 2-4 times greater than in humans, based on literature data |  |  |  |
|  |  |  | Commen -The resu mixture | dividual compoun | are calculated from | e exposure to the |


| Table 3 | Benzene, toluene, xylene, tetramethyl benzene |
| :--- | :--- |


| Table 3 Ben | Benzene, toluene, xylene, tetramethyl benzene | methyl benzene |  |
| :---: | :---: | :---: | :---: |
| Authors Ahag | hotu et al. |  |  |
| Title $\quad$ Effe | Effect of methyl substitution of benzene of |  |  |
| Source ${ }^{\text {Toxi }}$ | Toxicology Letters 159: 261-271, 2005 |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |
| - In vitro assessment of dermal permeation rate and retention of benzene and methylbenzenes in the skin | kerosene aromatic hydrocarbons <br> -dorsal skin excised from hairless rats after euthanasia (removed adhering fat and subcutaneous tissue) <br> - Franz static diffusion cells <br> Analytical method: -liquid scintillating counting | 1 ml of pure substance spiked with radiolabeled compounds <br> Exposure area: <br> $-0.636 \mathrm{~cm}^{2}$ <br> Exposure duration: <br> - 8 h <br> Components measured: <br> -benzene <br> -toluene <br> -xylene <br> -tetramethyl benzene | Steady-state flux of neat aromatic hydrocarbons <br> *estimated from the Figure 1 <br> - retention of all four chemicals in the stratum corneum is much higher than in epidermis and dermis at all time points (exact values are not available except for xylene, data are shown only graphically) <br> - higher retention of tetramethyl benzene than benzene in the stratum corneum showed effect of methyl substitution of benzene ring <br> - from the mass balance it was shown that for xylene the recovery ranged from 92 to $105 \%$, indicating a minimum evaporation of xylene during the experiment |
|  |  |  | Comment: <br> -The data are presented only graphically, exact flux values are given for benzene, toluene and xylene, but corresponding standard deviations or standard error are not presented <br> -No flux values are presented for tetramethyl benzene, possibly since the flux (in the graph presented) is much lower than for benzene, toluene and xylene |



Table 5 $\quad$ Toluene

| Table 5 Tolu | Toluene |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Authors ${ }^{\text {Bom }}$ | Boman et al. |  |  |  |
| Title $\quad$ Perc | Percutaneous absorption of organic solvents during intermittent exposure in guinea pigs |  |  |  |
| Source Acta | Acta Dermato Venereologica, 75: 114-119, 1995 |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results | Comment |
| -In vivoln vivo assessment of percutaneous absorption of three solvents with various physicochemical properties during brief intermittent exposure using an animal model | Test material: -toluene <br> -n-butanol <br> -1,1,1-trichloroethane <br> Experimental method and species: <br> - female guinea pigs (n =5) <br> Analytical method: <br> -GC-FID | -neat toluene <br> Intermittent exposure -duration of exposure: $8 \times 1$ min every 30 min -exposure area: $3.14 \mathrm{~cm}^{2}$ <br> Continuous exposure: -duration of exposure: 4 h <br> -exposure area: $3.14 \mathrm{~cm}^{2}$ <br> Components measured: -toluene in blood | For the approximation of the total dose absorbed during the exposure, area under the time vs. blood concentration curve (AUC) was calculated for intermittent and continuous exposure to toluene <br> The total dose absorbed of toluene was approximately $16.5 \%$ of continuous exposure (determined from graph, since the data were shown only graphically). |  |

Table 6 Toluene

Table 7 $\quad$ Toluene

| le 7 |  |  |  |
| :---: | :---: | :---: | :---: |
| Authors Bom | Boman and Maibach |  |  |
| Title $\quad$ Influ | Influence of evaporation and solvent mixtures on the absorption of toluene and n-butanol in human skin in vitro |  |  |
| Source Ann | Annal of Occupational Hygiene, 44: 125-135, 2000 |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |
| -In vitro percutaneous absorption of toluene. Influence of forced ventilation | Test material: -toluene <br> - human split-thickness skin ( $250 \mu \mathrm{~m}$ ) <br> flow through penetration-evaporation cell (adjustable ventilation above donor liquid) <br> The evaporation rate was determined gravimetrically (Gilbert, 1971) <br> Analytical method: -liquid scintillation counting | -radiolabelled toluene as neat toluene, toluene in butanol ( $50 \%$, $\mathrm{v} / \mathrm{v}$ ), and toluene in chloroform, methanol and butanol (25:33.3:16.7:25, v/v) <br> -applied volume: 200-300 $\mu \mathrm{l}$ <br> -duration of exposure: <br> 24 h <br> -exposure area: $1.0 \mathrm{~cm}^{2}$ <br> Components measured: -toluene | Influence of ventilation air flow rate and cosolvent on in vitro skin absorption of toluene <br> *CM $=50 / 50$ mixture with chloroform/methanol (2:1) <br> **Butanol $=50 / 50$ mixture with butanol <br> -the absorption of toluene varied between donors in the absence of ventilation ( $0 \mathrm{ml} / \mathrm{min}$ ) -the absorption of toluene was increased when applied in mixtures, being highest in mixture with butanol at ventilation air flow rate of $0 \mathrm{ml} / \mathrm{min}$ -at ventilation air flow rate of $90 \mathrm{ml} / \mathrm{min}$ there was a significant decrease of toluene absorption for all mixtures and -ventilation decreased absorption of toluene |



| Table 9 Tol | Toluene |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors ${ }^{\text {Thr }}$ | Thrall and Woodstock |  |  |  |  |  |
| Title $\begin{array}{l}\text { Eval } \\ \text { mo }\end{array}$ | Evaluation of the dermal absorption of aqueous toluene in F344 rats using real-time breath analysis and physiologically based pharmaco modelling |  |  |  |  |  |
| Source Jour |  |  |  |  |  |  |
| Type / aim of <br> study Test material/ species / <br> technique / detection |  | Exposure condition | Results |  |  |  |
| -In vivoln vivo assessment of the dermal absorption of toluene | -toluene <br> - male F344 rats ( $n=3$ ) <br> Analysis using real-time exhaled breath analysis and PBPK modelling <br> Analytical method: <br> -GC-FID <br> ASGDI-MS/MS | - approximately 2 ml of aqueous toluene was applied at two concentration levels: 0.5 $\mathrm{mg} / \mathrm{l}$ and $0.25 \mathrm{mg} / \mathrm{l}$ <br> -duration of exposure: <br> 5 h <br> -exposure area: $4.9 \mathrm{~cm}^{2}$ <br> Components measured: -toluene in exhaled air (breath) | Dermal expos <br> *Mean $\pm$ SE <br> *average val <br> -maximum peak of exposure obtained for th -both concent -the data indi | $\begin{gathered} \text { s (mean } \pm \text { SD, } \mathrm{n} \\ \begin{array}{c} \text { Exposure level } \\ \text { (mg/kg bw) } \end{array} \\ \hline 1.75 \pm 0.32 \\ 4.14 \pm 0.38 \\ \hline \end{gathered}$ <br> nd \% dose abso <br> ntration in exhal igh concentratio centration level e profiles showe absorption of to | $\mathrm{K}_{\mathrm{p}}$ <br> $(\mathrm{cm} / \mathrm{h})$$\|$$0.076 \pm 0.004$ <br> $0.070 \pm 0.004$ <br> $0.074 \pm 0.005$ <br> tween two conc <br> th was achieved ( $0.5 \mathrm{mg} / \mathrm{l}$ ), a $\mathrm{mg} / \mathrm{l}$ ) elimination ph hrough the rat sk | \% dose <br> absorbed <br> $45 \pm 4$ <br> $42 \pm 18$ <br> $43.8 \pm 9.6$ <br> tration levels <br> within 1 h after the end similar results were |


| Table 10 ${ }^{\text {10 }}$ Tolu | Toluene |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Authors Kled | Klede et al. |  |  |  |  |
| Title $\quad$ Tran | Transcutaneous penetration of toluene in rat skin a microdialysis study |  |  |  |  |
| Source ${ }^{\text {Exp }}$ | Experimental Dermatology, 14: 103-108, 2005 |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |
| -In vivoln vivo assessment of the permeation kinetics of toluene penetration after short and long exposure durations -Effect of exposure duration, tape stripping, and pre-treatment with barrier creams | -toluene | -200 $\mu \mathrm{l}$ of neat toluene was applied | Amount of toluene collected in the dialysate samples after dermal exposure (mean $\pm$ SEM) |  |  |
|  | - male Wistar rats(n = 82) |  | Skin pre-treatment | $\begin{gathered} \mu \mathrm{g} \\ \text { (15 min exposure) } \end{gathered}$ | $\begin{gathered} \mu \mathrm{g} \\ (240 \text { min exposure }) \end{gathered}$ |
|  |  | Exposure duration: <br> 15 min and 240 min | Dialysate (AUC) Control | $11.63 \pm 1.54$ |  |
|  | Technique: |  | Control tape stripped | $12.95 \pm 1.54(\mathrm{n}=8)$ | $34.55 \pm 7.27(n=8)$ |
|  | microdialysis |  | Cremor basalis control | $12.18 \pm 2.13(\mathrm{n}=7)$ | $40.36 \pm 6.83(\mathrm{n}=7)$ |
|  |  | Animals were sacrificed at the end of experiment (after 240 min for both exposures) | Cremor basalis tape stripped | $10.72 \pm 3.41(\mathrm{n}=6)$ | $32.68 \pm 9.92(\mathrm{n}=7)$ |
|  |  |  | Arretil control | $16.24 \pm 3.97(\mathrm{n}=6)$ | $40.28 \pm 6.05(\mathrm{n}=6)$ |
|  |  |  | Arretil tape stripped | $13.68 \pm 2.84(\mathrm{n}=7)$ | $38.23 \pm 8.50(\mathrm{n}=5)$ |
|  | Analytical method: <br> -GC-FPD <br> -GC-ECD | Exposure area: $0.9 \mathrm{~cm}^{2}$ <br> Components measured: <br> -Toluene in dialysate <br> -o-cresol in urine (collected during 240 mins for both exposure durations) | Urine (o-cresol) Control | $8.4 \pm 1.0$ | $12.7 \pm 1.4$ |
|  |  |  |  |  |  |
|  |  | Components measured: -Toluene in dialysate -o-cresol in urine (collected during 240 mins for both exposure durations) | -significant difference in AUC short and long exposure durat treatments. <br> -there was a significant differe exposure (ANOVA, p < 0.001 ) difference between different $p$ <br> -the urine o-cresol content ref different exposure durations comparison to the toluene co of 3 ). This could be explai incorporation of toluene in ad time points, which would pro precise data on the effect of exp | dialysate content of s, but no difference <br> in o-cresol content but according to the reatments (data were <br> ed the findings of tol ough to a lesser exte t found in the dialys by a slower elimi se tissue. The auth ly have yielded mor sure durations. | uene was observed between present among different pre- <br> ine after 15 min and 240 min ors there was no significant reported) <br> content in dialysate for the factor of difference of 1.5) in (average factor of difference process of toluene and did not collect urine at later rine o-cresol and give more |




| Table 13 Xyl | Xylene |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors Kezic | Kezic et al. |  |  |  |  |  |  |
| Title $\quad$ Per | Percutaneous absorption of m-xylene vapour in volunteers during pre-steady and steady state |  |  |  |  |  |  |
| Source Tox | Toxicology Letters, 153: 273-282, 2004 |  |  |  |  |  |  |
| Type / aim of study | of Test material/ species / technique / detection | Exposure condition | Results |  |  |  |  |
| -In vivoln vivo assessment of percutaneous absorption of m xylene vapour through human skin | -m-xylene <br> -human male volunteers $(n=6)$ | Dermal exposure: -body part: arm in glass chamber of 60 cm in length) | Maximum blood flux for different exposure durations |  |  |  |  |
|  |  |  |  | Maximum flux $\times 10^{-5}\left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right)$ |  |  |  |
|  |  |  | Exposure duration | 20 min | 45 min | 120 min | 180 min |
|  |  | -exposure concentration: | Volunteer 1 | 3.6 | 4.1 | 5.7 | 6.9 |
|  | Data analysis: -by linear system dynamics (LSD, Opdam, 1991)using a reference exposure via inhalation and (de)convolution to yield dermal permeation rate time courses <br> Analytical method: -GC-FID | 29.4 mg/m ${ }^{3}$ (range: | Volunteer2 | 2.8 | 2.1 | 3.0 | 3.4 |
|  |  | $24.9-34.0$ mg/m ${ }^{3}$ ) | Volunteer 3 | 9.0 | 11.0 | 12.0 | 12.1 |
|  |  |  | Volunteer 4 | 0.5 | 1.2 | 2.8 | 4.8 |
|  |  | -flow rate of xylene | Volunteer 5 | 1.2 | 2.9 | 3.1 | 3.9 |
|  |  | vapour: $31 / m \mathrm{~min}$ : | Volunteer 6 | 3.2 | 3.7 | 9.1 | 6.8 |
|  |  | -exposure duration: | Average (mean $\pm$ SD) | $3.4 \pm 1.2$ | $4.2 \pm 1.4$ | *5.9 $\pm 1.6$ | $\text { * } 6.3 \pm 1.4$ |
|  |  | $20,45,120$ and 180 min |  Apparent permeability coefficient ( $\left.\mathrm{K}_{\mathrm{p}}, \mathrm{cm} / \mathrm{h}\right)$ |  |  |  |  |
|  |  | -exposure area: | Average Kp <br> (mean $\pm$ SD) |  |  | $059 \pm 0.016$ | $063 \pm 0.014$ |
|  |  | Average: $1178 \mathrm{~cm}^{2}$ (range: $1100-1285 \mathrm{~cm}^{2}$ ) | *all fluxes were adjusted to an exposure concentration of $1 \mu \mathrm{~g} / \mathrm{cm}^{3}$ assuming a linear relationship between fluxes and concentration |  |  |  |  |
|  |  | Components measured: -xylene in exhaled breath | -permeability coefficient was calculated only for 120 and 180 min exposure since after 90 min of exposure permeation rates reached a constant value indicating attainment of pseudo steady-state permeation |  |  |  |  |



| Table 15 Xyle | Xylene and toluene |  |  |
| :---: | :---: | :---: | :---: |
| Authors Kez | Kezic et al. |  |  |
| Title $\quad$ Skin | Skin absorption of some vaporous solvents in volunteers |  |  |
| Source Inte | International Archive of Occupational and Environmental Health, 73: 415-422, 2000 |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |
| -In vivoln vivo assessment of percutaneous absorption rates of m-xylene, toluene and hexane vapour in humans | -m-xylene, toluene, hexane <br> -human volunteers ( $\mathrm{n}=$ 5) <br> Data analysis: -by linear system dynamics (LSD, Opdam, 1991)using a reference exposure via inhalation and (de)convolution to yield dermal permeation rate time courses <br> Analytical method: -GC-FID | Inhalation (reference) exposure: <br> -exposure concentration: C< occup. exp. limit in NL <br> -exposure duration: 10 min <br> Dermal exposure: -body part: arm in glass chamber of 60 cm in length) <br> -exposure concentration: m-xylene: $0.23 \mathrm{mmol} / \mathrm{l}$ toluene: $0.70 \mathrm{mmol} / \mathrm{I}$ hexane: $1.31 \mathrm{mmol} / \mathrm{l}$ (CV < 5\%) <br> -exposure duration: m-xylene: 20 min toluene: 20 min hexane: 30 min <br> -exposure area: range: $960-1070 \mathrm{~cm}^{2}$ <br> Components measured: -xylene, toluene and hexane in exhaled air | Absorption rates into the skin ( RATE $_{\text {skin }}$ ) and the maximum absorption rates into the blood (RATE blood, $\max$ ) (mean $\pm$ SEM) <br> Since the duration of exposure in this study was not long enough to reach steadystate, Kp could not be determined. However, since $\mathrm{K}_{\mathrm{p}}$ is always lower than the absorption rate into the skin and higher than or equal to the maximum absorption rate into the blood, the values given in the table represent the lower and upper limits for $\mathrm{K}_{\mathrm{p}}$, respectively <br> Both the absorption rates into the skin and into the blood were similar for xylene and toluene and they were approximately 10-20 times higher than obtained for hexane. |


| Table 16 Tolu | Toluene |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors Sus | Susten et al. |  |  |  |  |  |  |
| Title In v | In vivoln vivo percutaneous absorption studies of volatile organic solvents in hairless mice II. Toluene, ethylbenzene and aniline |  |  |  |  |  |  |
| Source Jou |  |  |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |  |
| -In vivoln vivo percutaneous absorption of toluene and ethylbenzene in hairless mice | - toluene, ethylbenzene <br> - male albino hairless mice ( $n=5$ ) | -exposure concentration: $5 \mu$ l of treatment solution containing radiolabelled toluene or ethylbenzene | Applied and absorbed amounts and absorption rates for toluene and ethylbenzene (mean $\pm$ SD) |  |  |  |  |
|  |  |  | Component | Amount applied (mg) |  | *Amount absorbed (mg) | Absorption rate $\left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right)$ |
|  |  |  | Toluene ethylbenzene | $\begin{aligned} & 3.89 \pm 0.25 \\ & 4.10 \pm 0.19 \end{aligned}$ |  | $\begin{aligned} & 0.0897 \pm 0.0708 \\ & 0.1486 \pm 0.1278 \end{aligned}$ | $\begin{aligned} & 2.94 \pm 2.27 \\ & 2.22 \pm 1.89 \end{aligned}$ |
|  | Analysis: <br> Mass balance <br> Analytical method: -liquid scintillation counting | -exposure duration: <br> Toluene: 113 sec <br> Ethylbenzene: 294 sec | *amount absorb excreta and car <br> Distribution of of (\% nominal dos | ed $=$ deter ass <br> the absorb recovered | mined from <br> d of ${ }^{14} \mathrm{C}$ la mean $\pm$ SE | radioactivity lev <br> lled aromatic s | ls found in expired breath, <br> olvents after dermal exposure |
|  |  | -exposure area: $0.8 \mathrm{~cm}^{2}$ <br> Components measured: -toluene and ethylbenzene in expired breath, excreta and carcass | Component | Absorbed total | Carcass | Expired breath | Excreta |
|  |  |  | Toluene $(\mathrm{n}=12)$ <br> Ethylbenzene $(\mathrm{n}=11)$ | $\begin{aligned} & 15.4 \pm 2.0 \\ & 15.5 \pm 2.0 \end{aligned}$ | $\begin{gathered} 11.0 \pm 3.0 \\ 4.5 \pm 1.0 \end{gathered}$ | $\begin{aligned} & 20.5 \pm 5.0 \\ & 14.3 \pm 6.0 \end{aligned}$ | $\begin{aligned} & 53.1 \pm 6.0 \\ & 65.6 \pm 5.0 \end{aligned}$ |
|  |  |  | -the amount of ethylbenzene absorbed was greater due to the longer exposure duration than toluene, although toluene had a greater absorption rate. |  |  |  |  |


| Table 17 | Xylene |
| :--- | :--- |
| Skow |  |


Table 18 Xylene




| Table 21 Ker | Kerosene |  |  |
| :---: | :---: | :---: | :---: |
| Authors ${ }^{\text {Tsu }}$ | Tsujino et al. |  |  |
| Title | Distribution of kerosene components in rats following dermal exposure |  |  |
| Source ${ }^{\text {a }}$ Inte | International Journal of Legal Medicine 116 :207-211, 2002 |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |
| -In vivoln vivo evaluation of the tissue distribution of kerosene components in rats following dermal exposure | -standard kerosene: aliphatic and aromatic hydrocarbons <br> - male Wistar rats (250300 g ) <br> Analytical method: -GC-MS analysis | Rats were exposed for 1 h abdominally using $2 x 2$ $\mathrm{cm}^{2}$ cotton piece soaked with 1 ml of standard kerosene and sacrificed immediately after the end of exposure without postexposure sampling <br> -rats were sacrificed at 0-12 h post-exposure <br> -exposure duration: <br> 1 h <br> Components measured: -aliphatic hydrocarbons $\mathrm{C}_{9}-\mathrm{C}_{16}$ <br> -aromatic hydrocarbons cumene pseudocumene mesitylene 1,2,3-trimethylbenzene | Experiment 1: <br> -the data were presented as ratios of peak areas of aliphatic and aromatic hydrocarbons to analytical internal standard (IS) in each blood, brain, lung, liver spleen, kidney, muscle and adipose tissue <br> -Ratios of aliphatic hydrocarbons to IS were significantly lower in kidney, liver and brain ( $p=0.002,0.01$ and 0.05 , respectively) compared to ratios of aromatic hydrocarbons to IS. Ratios of aliphatic hydrocarbons to is compared to ratios of aromatic hydrocarbons to IS were non-significantly lower in blood, spleen and the lungs ( $p>0.05$ ) while the ratios of aliphatic hydrocarbons to IS were non-significantly higher in muscle and adipose tissue than ratios of aromatic hydrocarbons to IS ( $p>$ 0.3 ). <br> Overall, ratios of aliphatic hydrocarbons to IS $(0.12 \pm 0.02)$ were significantly lowerr than ratios of aromatic hydrocarbons to IS ( $0.21 \pm 0.05$ ) (mean $\pm \mathrm{SE}, \mathrm{p}=0.04$ ) <br> These data showed that aromatic hydrocarbons were dermally absorbed to a greater degree than aliphatic hydrocarbons |
|  |  |  | Comment: <br> -not all data are shown <br> -only three animals were used |


| Table 23 Keros | Kerosene |  |  |
| :---: | :---: | :---: | :---: |
| Authors Hied | Hieda et al. |  |  |
| Title $\quad$ Skin | Skin analysis following dermal exposure to kerosene in rats: the effects of postmortem exposure and fire |  |  |
| Source Inte | International Journal of Legal Medicine 118: 41-46, 2004 |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |
| $\begin{aligned} & \text {-In vivo dermal } \\ & \text { absorption of } \\ & \text { kerosene in rats. } \\ & \text { Evaluation of } \\ & \text { usefulness of } \\ & \text { skin analysis for } \\ & \text { the forensic } \\ & \text { examination of } \\ & \text { cases involving } \\ & \text { postmortem } \\ & \text { dermal exposure } \\ & \text { to kerosene } \\ & \text { and/or fire } \end{aligned}$ | -kerosene aliphatic and aromatic hydrocarbons <br> -in vivo, male SpragueDawley rats (330-387 g) <br> -ante-mortem and postmortem exposure <br> -components measured: -aliphatic hydrocarbons $\mathrm{C}_{9}-\mathrm{C}_{16}$ <br> -aromatic hydrocarbons <br> 1,2,3-trimethylbenzene <br> 1,2,4-trimethylbenzene <br> 1,3,5-trimethylbenzene <br> Analytical method: -GC-MS analysis | -rats (ante-mortem) were exposed abdominally or on the back using cotton piece soaked with 4 ml of standard kerosene. <br> -after sacrificing, the trunk blood was collected <br> -rats (post-mortem) were exposed abdominally or on the back using cotton piece soaked with 4 ml of standard kerosene. <br> -exposure duration: 30 mins <br> -part of the exposed skin (ante- and post-mortem) was burned with a portable burner | -the concentration of aliphatic and aromatic hydrocarbons was consistently lower in back skin than in abdominal skin for both ante- and post-mortem exposure <br> -there was no difference in kerosene levels between ante-mortem and post-mortem exposure on abdominal skin and back skin <br> the kerosene concentration in mildly and severely burned skin was $84 \%$ and $28 \%$ of that in non-burned exposed skin |


| Table 24 | Kerosene |  |  |
| :---: | :---: | :---: | :---: |
| Authors ${ }^{\text {Fuji }}$ | Fujihara et al. |  |  |
| Title $\quad$ The | The levels of kerosene components in biological samples after repeated dermal exposure to kerosene in rats |  |  |
| Source Leg | Legal Medicine 6: 109-116, 2004 |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |
| In vivo dermal exposure of rats to kerosene. <br> Evaluation of accumulation of kerosene components after repeated (daily) dermal exposure | Standard kerosene aliphatic and aromatic hydrocarbons <br> male Sprague-Dawley rats (332-450 g) <br> Analytical method: GC-MS | -rats (divided into groups) were repeatedly exposed on the abdominal skin using cotton piece soaked with 4 ml of standard kerosene. <br> -exposure duration: Single exposure for 1 h and sacrificed at 0 and 24 $h$ after the end of exposure Repeated exposure for 5 consecutive days for 1 h each day and sacrificed at 0 and 24 h after the end of exposure period <br> Components measured: -aliphatic hydrocarbons $\mathrm{C}_{9}-\mathrm{C}_{16}$ <br> -aromatic hydrocarbons 1,2,3-trimethylbenzene 1,2,4-trimethylbenzene 1,3,5-trimethylbenzene | Blood: <br> There was no significant difference in the blood levels of aromatic hydrocarbons between groups which were sacrificed immediately after the end of exposure or between groups which were sacrificed at 24 h after the end of exposure, but there was a significant difference between groups sacrificed immediately and 24 h after the end of exposure being higher in the groups sacrificed immediately after the end of exposure <br> There were no significant differences in the blood level of aliphatic hydrocarbons between all groups <br> Skin: <br> The skin level of aromatic hydrocarbons was lower then the level of aliphatic hydrocarbons in all groups regardless of repeated or single dose. <br> Significantly high levels of aliphatic hydrocarbons were detected after the fifth exposure and immediate sacrificing of the animals compared to repeated exposure and sacrificing after 24 h and single dose and sacrificing at 0 and 24 h . <br> Comment: <br> The data were presented only graphically as total levels of aliphatic or aromatic hydrocarbons |



| Table 26 Jet | Jet fuel |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Authors ${ }^{\text {Sing }}$ | Singh et al. |  |  |  |
| Title | In vivo percutaneous absorption, skin barrier perturbation and irritation from JP-8 jet fuel components |  |  |  |
| Source ${ }^{\text {dru }}$ | Drug and chemical toxicology, 26: 135-146, 2003 |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |
| -In vivo percutaneous absorption of three components of JP-8 <br> Assesment of skin barrier pertubation and irritation by JP-8 components. | -jet fuel (JP-8) aliphatic and aromatic hydrocarbons <br> - weanling pigs <br> Analytical method: -liquid scintillation counting | -0.5 ml of JP-8 spiked with $1 \mu \mathrm{~mol}$ of each radiolabelled component: heptane, hexadecane and xylene <br> -duration of exposure: <br> 30 min <br> -exposure area: $3.14 \mathrm{~cm}^{2}$ <br> Components measured: -heptane <br> -hexadecane <br> -xylene | Percentage of <br> Component <br> Xylene <br> Heptane <br> Hexadecane <br> The percenta heptane and | plied dose absorbed into the skin <br> e applied dose absorbed into the skin was greater for aliphatic cane then aromatic xylene. |



Table 29 Jet fuel


| Table 30 | Jet fuel |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors Bay | Baynes et al. |  |  |  |  |  |
| Title | Mixture effects of JP-8 additives on the dermal disposition of Jet fuel components |  |  |  |  |  |
| Source ${ }^{\text {P }}$ Tox | Toxicology and Applied pharmacology, 175: 269-281, 2001 |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |
| -In vitro assessment of the influence if additives (DIEGME, 8Q21 and Stadis450) on dermal disposition of topically applied JP-8 | $\left.\begin{array}{l\|l\|}\hline \text {-jet fuel (JP-8) aliphatic and } \\ \text { aromatic hydrocarbons }\end{array} \quad \begin{array}{l}\text { Experiment 1: } \\ -10 \mu \mathrm{l} \text { of specified jet fuel } \\ \text { containing radiolabeled } \\ \text { - flow through diffusion cells } \\ \text { naphthalene (1.21\%) } \\ \text { and dodecane (4.70\%) }\end{array}\right\}$ |  | Absorption of marker compounds from jet fuels in experiment 1 (mean $\pm$ SE, $n=4$ for each fuel): |  |  |  |
|  |  |  | Component | $\begin{gathered} \text { Max flux } \times 10^{-4} \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{K}_{\mathrm{p}} \times 10^{-4} \\ (\mathrm{~cm} / \mathrm{h}) \\ \hline \end{gathered}$ | $\begin{aligned} & \mathrm{D} \times 10^{-4} \\ & \left(\mathrm{~cm}^{2} / \mathrm{h}\right) \\ & \hline \end{aligned}$ |
|  |  |  | ${ }^{14} \mathrm{C}$-Naphthalene |  |  |  |
|  |  |  | Jet-A | $2.22 \pm 0.12$ | $1.68 \pm 0.06$ | $0.95 \pm 0.05$ |
|  |  |  | Jet-A + DIEGME | $3.18 \pm 0.36$ | $1.98 \pm 0.24$ | $1.20 \pm 0.13$ |
|  |  |  | Jet-A + 8Q21 | $2.64 \pm 0.12$ | $1.68 \pm 0.06$ | $1.02 \pm 0.05$ |
|  |  |  | Jet-A + Stadis450 | $2.76 \pm 0.12$ | $1.74 \pm 0.06$ | $1.04 \pm 0.06$ |
|  |  |  | Jet-A + DIEGME + 8Q21 | $1.24 \pm 0.08$ | $1.35 \pm 0.10$ | $0.99 \pm 0.03$ |
|  |  |  | Jet-A + DIEGME + Stadis450 | $1.07 \pm 0.14$ | $1.13 \pm 0.15$ | $1.36 \pm 0.12$ |
|  |  |  | Jet-A + 8Q21 + Stadis450 | $0.97 \pm 0.03$ | $0.97 \pm 0.03$ | $1.04 \pm 0.05$ |
|  |  |  | ${ }^{\mathrm{JP}-8} \mathrm{C} \text {-Dodecane }$ | $1.86 \pm 0.06$ | $1.44 \pm 0.06$ | $0.99 \pm 0.05$ |


Table 31 Jet fuel

| rs | Comparative mixture effects of JP-8(100) additives on the dermal absorption and disposition of jet fuel hydrocarbons in different membrane mad systems |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Title $\quad$Com <br> sys |  |  |  |  |  |  |
| Source Tox | Toxicology Letters, 150: 351-365, 2004 |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |
| -In vitro assessment of the influence of additives (MDA, 8Q405 and BHT) on dermal absorption of topically applied JP-8 |  |  | Absorption parameters following dermal exposure to naphthalene in jet fuel mixtures (mean $\pm$ SEM) |  |  |  |
|  |  |  | Naphthalene | $\begin{aligned} & \text { Flux } \times 10^{-3} \\ & \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{aligned}$ | $\begin{gathered} \mathrm{K}_{\mathrm{p}} \times 10^{-3} \\ (\mathrm{~cm} / \mathrm{h}) \\ \hline \end{gathered}$ | $\begin{gathered} \text { Diffusivity } \\ \times 10^{-3}\left(\mathrm{~cm}^{2} / \mathrm{h}\right) \\ \hline \end{gathered}$ |
|  |  |  | $\begin{aligned} & \text { silastic membrane } \\ & \text { JP-8 }(n=5) \\ & \text { JP-8 + MDA }(n=5) \\ & \text { JP-8 + BHT }(n=5) \\ & \text { JP-8 + 8Q405 }(n=4) \\ & \text { JP-8 + MDA }+ \text { BHT }(n=5) \\ & \text { JP-8 + MDA }+8 \text { Q } 405(n=4) \\ & \text { JP-8 + BHT }+8 \text { Q405 }(n=4) \\ & \text { JP-8(100 })(n=5) \\ & \text { (pig skin } \\ & \text { JP-8 }(n=5) \\ & \text { JP-8 + MDA }(n=5) \\ & \text { JP-8 + BHT }(n=5) \\ & \text { JP-8 + 8Q405 }(n=4) \\ & \text { JP-8 + MDA + BHT }(n=5) \\ & \text { JP-8 + MDA + 8Q405 }(n=4) \\ & \text { JP-8 + BHT + 8Q405 }(n=4) \\ & \text { JP-8(100) }(n=5) \end{aligned}$ | $\begin{aligned} & 19.29 \pm 0.60 \\ & 28.23 \pm 1.62 \\ & 27.17 \pm 0.84 \\ & 26.86 \pm 0.71 \\ & 21.22 \pm 0.52 \\ & 21.65 \pm 0.62 \\ & 21.75 \pm 0.73 \\ & 21.01 \pm 0.86 \\ & 2.21 \pm 0.27 \\ & 2.63 \pm 0.09 \\ & 2.48 \pm 0.28 \\ & 2.17 \pm 0.16 \\ & 2.88 \pm 0.35 \\ & 2.64 \pm 0.47 \\ & 2.70 \pm 0.48 \\ & 2.24 \pm 0.25 \\ & \hline \end{aligned}$ | $\begin{aligned} & 1.78 \pm 0.06 \\ & 2.30 \pm 0.13 \\ & 2.22 \pm 0.07 \\ & 2.19 \pm 0.06 \\ & 1.99 \pm 0.05 \\ & 2.03 \pm 0.06 \\ & 2.04 \pm 0.07 \\ & 1.93 \pm 0.08 \\ & 0.21 \pm 0.03 \\ & 0.20 \pm 0.01 \\ & 0.19 \pm 0.02 \\ & 0.17 \pm 0.01 \\ & 0.23 \pm 0.03 \\ & 0.21 \pm 0.04 \\ & 0.21 \pm 0.04 \\ & 0.21 \pm 0.02 \end{aligned}$ | $\begin{gathered} 1.690 \pm 0.490 \\ 16.775 \pm 5.672 \\ 3.855 \pm 0.907 \\ 2.164 \pm 0.403 \\ 1.408 \pm 0.411 \\ 1.469 \pm 0.790 \\ 0.251 \pm 0.023 \\ 6.823 \pm 5.836 \\ \\ 0.455 \pm 0.049 \\ 0.337 \pm 0.023 \\ 0.324 \pm 0.039 \\ 0.330 \pm 0.042 \\ 0.176 \pm 0.015 \\ 0.158 \pm 0.009 \\ 0.161 \pm 0.008 \\ 0.402 \pm 0.037 \\ \hline \end{gathered}$ |
| Comment |  |  | Results |  |  |  |

Table 31 Jet fuel

Table 32 Jet fuel

| Table 32 Jet | Jet fuel |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors ${ }^{\text {Cha }}$ | Chao et al. |  |  |  |  |  |  |
| Title $\quad$ Der | Dermal exposure to jet fuel JP-8 significanty contributes to the production of urinary naphtols in fuel-cell maintenance workers |  |  |  |  |  |  |
| Source ${ }^{\text {Env }}$ | Environmental health perspectives, 114: 182-185, 2006 |  |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |  |
| -In vivo assessment of the contribution of dermal and inhalation exposure to JP-8 to the total body dose | -jet fuel (JP-8) | Assessment of inhalation exposure: | Dermal and inhalation absorption of naphthalene from jet fuels (geometric mean (GM) $\pm$ geometric SD (GSD), $\mathrm{n}=43-85$ ): |  |  |  |  |
|  | - humans ( $\mathrm{n}=85$ ) | -passive monitors | Indicator of exposure | GM | GSD | Min | $\operatorname{Max} \times 10^{+4}$ |
|  | (US Air Force fuel-cell | attached to workers shirt | Dermal naphthalene ( $\mathrm{ng} / \mathrm{m}^{2}$ ) | 4180 | 9.35 | 100 | 509 |
|  | maintenance workers) | collars | Breathing-zone naphthalene ( $\mathrm{ng} / \mathrm{m}^{3}$ ) | 614000 | 2.21 | 670 | 391 |
|  |  |  | Preexposure breath naphthalene ( $\mathrm{ng} / \mathrm{m}^{3}$ ) | 492 | 1.99 | 330 | 1.61 |
|  | Analytical method: | -duration of exposure: | Breath naphthalene ( $\mathrm{ng} / \mathrm{m}^{3}$ ) | 9230 | 2.88 | 667 | 7.58 |
|  | -GC-MS | 4 h during work shift | Preexsposure urinary 1-naphthol (ng/l) | 4200 | 3.77 | 242 | 3.90 |
|  |  |  | Urinary 1-naphthol (ng/l) | 28000 | 2.26 | 483 | 12.70 |
|  |  |  | Preexsposure urinary 2-naphthol (ng/l) | 4350 | 3.06 | 424 | 3.79 |
|  |  | Assessment of dermal | Urinary 2-naphthol (ng/l) | 38400 | 2.46 | 485 | 31.50 |
|  |  | exposure: <br> -tape striping of three body regions for each worker (three successive tapes per body region) <br> Components measured: -naphthalene in tape strips <br> - 1-naphtols and 2naphtols in urine | -the contribution of dermal and inhalation the total body dose of JP-8 were investiga <br> -for urinary 1-naphthol breathing-zone significant predictors (explaining 88.2 and - for urinary 2-naphthol dermal exposure were significant predictors (explaining respectively) <br> The results suggested that that derm significantly to urinary 2 -naphthol urinary which is explained as possibly caused by the skin by mixed-function oxygenase and | posure, sm using multip <br> thalene \% of total eathing-z 35.8 | king a e line <br> d sm arianc <br> nap <br> 13. <br> o na <br> to 1-n <br> metabo <br> zymes | othe regre <br> ng w respe halen of <br> thale hthol m of | covariates to on analysis <br> e the only vely) and smoking al variance, <br> contributed inary levels, phthalene in |



| Table 33 | Jet fuel |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors | Muhammad et al. |  |  |  |  |  |
| Title | Effect of in vivo jet fuel exposure on subsequent in vitro dermal absorption of indivi |  |  |  |  |  |
| Source | Journal of Toxicology and Envir | nmental Health, Part A , 68: | 719-737, 2005a |  |  |  |
| Type / aim of <br> study Test material/ species / <br> technique / detection Exposure condition |  |  | Results |  |  |  |
| Ethyl benzene, o-xylene and trimethyl benzene showed increase in absorbed amount proportional to the length of exposure. Similarly dodecane showed also increase in absorbed amount, however to a lesser extent. <br> Steady state flux and permeability coefficient were for all hydrocarbons, (except tridecane in 1-day pre-exposure experiment) higher in the skin pre-exposed to jet fuel than in control in both 1-day and 4-day preexposure experiments. <br> This result suggest that Jp-8 jet fuel pre-exposure of the skin produces changes in skin barrier function and alters the absorption parameters by increasing permeability coefficient and steady state flux as compared to controls. |  |  | Absorption of marker compounds from jet fuel after 4 day of pre-exposure (mean $\pm$ SEM, C = control pre-exposure, E = jet fuel pre-exposure, SS = steady-state) |  |  |  |
|  |  |  | Component | $\begin{gathered} \text { SS Flux } \times 10^{-3} \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{gathered}$ | $\begin{gathered} \mathrm{Kp} \times 10^{-3} \\ (\mathrm{~cm} / \mathrm{h}) \end{gathered}$ | Absorption ( $\mu \mathrm{g}$ ) |
|  |  |  | Naphthalene $_{\text {c }}$ | $4.81 \pm 0.36$ | $0.19 \pm 0.01$ | $12.030 \pm 0.989$ |
|  |  |  | Naphthalene ${ }_{\text {E }}$ | $7.57 \pm 0.47$ | $0.29 \pm 0.02$ | $19.835 \pm 1.453$ |
|  |  |  | $\mathrm{DMN}_{\mathrm{C}}$ | $0.67 \pm 0.07$ | $0.03 \pm 0.003$ | $1.179 \pm 0.155$ |
|  |  |  | $\mathrm{DMN}_{\mathrm{E}}$ | $0.99 \pm 0.09$ | $0.04 \pm 0.004$ | $2.056 \pm 0.221$ |
|  |  |  | Ethyl benzene ${ }_{\text {c }}$ | $0.61 \pm 0.15$ | $0.035 \pm 0.009$ | $0.637 \pm 0.136$ |
|  |  |  | Ethyl benzene ${ }_{\text {E }}$ | $2.04 \pm 0.17$ | $0.12 \pm 0.009$ | $2.594 \pm 0.250$ |
|  |  |  | $\mathrm{TMB}_{\mathrm{C}}$ | $0.49 \pm 0.04$ | $0.028 \pm 0.002$ | $0.837 \pm 0.090$ |
|  |  |  | $\mathrm{TMB}_{\mathrm{E}}$ | $1.52 \pm 0.10$ | $0.09 \pm 0.005$ | $3.821 \pm 0.267$ |
|  |  |  | $\mathrm{CHB}_{\mathrm{C}}$ | $0.29 \pm 0.02$ | $0.016 \pm 0.001$ | $0.470 \pm 0.042$ |
|  |  |  | $\mathrm{CHB}_{\text {E }}$ | $0.49 \pm 0.07$ | $0.025 \pm 0.004$ | $0.852 \pm 0.136$ |
|  |  |  | o-xylene ${ }_{\text {c }}$ | $1.02 \pm 0.28$ | $0.059 \pm 0.016$ | $1.156 \pm 0.263$ |
|  |  |  | o-xylene ${ }_{\text {E }}$ | $3.13 \pm 0.15$ | $0.180 \pm 0.009$ | $4.865 \pm 0.340$ |
|  |  |  | nonane ${ }_{\text {c }}$ | $0.03 \pm 0.01$ | $0.002 \pm 0.000$ | $0.080 \pm 0.015$ |
|  |  |  | nonane ${ }_{\text {E }}$ | $0.07 \pm 0.01$ | $0.005 \pm 0.000$ | $0.178 \pm 0.024$ |
|  |  |  | undecane ${ }_{\text {c }}$ | $0.06 \pm 0.01$ | $0.001 \pm 0.000$ | $0.135 \pm 0.022$ |
|  |  |  | undecane ${ }_{\text {E }}$ | $0.10 \pm 0.02$ | $0.002 \pm 0.000$ | $0.263 \pm 0.018$ |
|  |  |  | dodecane ${ }_{\text {c }}$ | $0.02 \pm 0.00$ | $0.0002 \pm 0.000$ | $0.032 \pm 0.007$ |
|  |  |  | dodecane ${ }_{\text {E }}$ | $0.05 \pm 0.01$ | $0.0008 \pm 0.000$ | $0.112 \pm 0.019$ |
|  |  |  | tridecane ${ }_{\text {c }}$ | $0.00 \pm 0.00$ | $0.0001 \pm 0.000$ | $0.0074 \pm 0.0019$ |
|  |  |  | tridecane ${ }_{E}$ | $0.02 \pm 0.01$ | $0.0003 \pm 0.000$ | $0.037 \pm 0.0093$ |

Table 34 $\quad$ Jet fuel

Table 35 Jet fuel


| Table 36 | PAH |
| :--- | :--- |
| Authors | Yang |




| Table 38 | PAH |
| :--- | :--- |
| Authors | Turk |



| Table 39 | PAH |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors ${ }^{\text {Van }}$ | Van Rooij et al. |  |  |  |  |  |
| Title $\quad$ Abs | Absorption of polycyclic aromatic hydrocarbons through human skin: differences between anatomical sites and individuals |  |  |  |  |  |
| Source Jou | Journal of Toxicology and Environmental Health, 38: 355-368, 1993 |  |  |  |  |  |
| Type / aim of <br> study Test material/ species / <br> technique / detection |  | Exposure condition | Results |  |  |  |
| -In vivo determination of dermal uptake of PAHs in humans <br> Contribution of various anatomical sites to uptake. <br> Assessment of interindividual variation | -therapeutical coal-tar ointment containing $10 \%$ coal-tar in a vehicle of zink oxide paste <br> - male volunteers ( $n=9$ ) <br> Analytical method: <br> -HPLC <br> -luminescence | Experiment 1 <br> Exposure surface <br> concentration <br> $2.5 \mathrm{mg} / \mathrm{cm}^{2}$ of coal-tar <br> -Exposure duration: 45 min <br> -Exposure area: $24 \mathrm{~cm}^{2}$ <br> Experiment 2 <br> Exposure surface concentration <br> $2.5 \mathrm{mg} / \mathrm{cm}^{2}$ of coal-tar <br> -Exposure duration: 6 h <br> -Exposure area: $400 \mathrm{~cm}^{2}$ <br> -Exposure sites: forehead, shoulder, volar forearm, palm site of hands, groin and ankle <br> Components measured: -disappearance of PAHs from the surface of the skin (experiment 1) -level of PAH metabolite 1-OH-pyrene in urine (experiment 2) | Skin absorption rate constants of PAHs (mean, $n=4$ )and excreted amount of 1-OHpyrene in urine (mean, $n=4$ ) at different anatomical sites |  |  |  |
|  |  |  | Site | Absorption rate constant (1/h) (from experiment 1) | Site | Excreted amount of 1-OH-pyrene (nmol) (from experiment 2) |
|  |  |  | Shoulder <br> Forearm <br> Forehead <br> Groin <br> Hand (palm) <br> Ankle <br> Mean $\pm$ SD | $0.135(0.069-0.196)$ $0.070(0.060-0.089)$ $0.065(0.046-0.083)$ $0.053(0.038-0.083)$ $0.037(0.026-0.050)$ $0.036(0.028-0.040)$ $0.066 \pm 0.037$ | Neck Calf Forearm Trunk Hand <br> Mean $\pm$ SD | $\begin{gathered} 14.6(10.1-23.8) \\ 13.9(7.0-23.0) \\ 11.3(5.0-21.2) \\ 10.8(7.8-15.0) \\ 7.7(6.0-11.1) \\ \\ 11.6 \pm 2.8 \\ \hline \end{gathered}$ |
|  |  |  | -The PAH absorption rate constants ranged from 0.026/h to 0.196/h. <br> -Two-way ANOVA showed significant effect of the anatomical site on the absorption rate constant but no significant effect of the individuals on the absorption rate constants. The variance explained $67 \%$ by skin site and only $7 \%$ by interindividual differences <br> -The total excreted amount of 1-OH-pyrene ranged from 5.0 to 23.8 nmol -two-way ANOVA showed that there are significant differences in the total excreted amount of 1-OH-pyrene between individuals, but no significant differences in the extent of urinary 1-OH-pyrene excretion between various skin sites -there was a significant difference between individuals for time needed in which half of the 1-OH-pyrene was excreted and it varied 8.2 to 18.9 h (data were not shown) -based on the absorption rate constants determined in the surface disappearance experiments, the authors estimated that at a low PAH (more comparable to occupational situations) dose 20-56\% will be absorbed after 6 h . |  |  |  |

Table 40 $\quad$ PAH

| Table 40 PAH | PAH |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors Van | Van Rooij et al. |  |  |  |  |  |
| Title $\quad$ Derm | Dermal absorption of polycyclic aromatic hydrocarbons in the blood-perfused pig ear |  |  |  |  |  |
| Source Jour | Journal of Applied Toxicology, 15: 193-2009, 1995 |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |
| -In vitro assessment of the validity of pyrene as a representative marker compound for the dermal absorption of other PAHs | -industrial coal-tar <br> - isolated blood perfused pig ear skin | Exposure surface concentration $11 \mathrm{mg} / \mathrm{cm}^{2}$ of coal-tar <br> -Exposure duration: $\leq 250$ min | The content of PAHs in coal tar, amount absorbed and their flux through pig ear skin during 200 min after coal tar application (means, or means $\pm$ SD) |  |  |  |
|  |  |  | Site | Content in coal tar (\%) | \#Relative amount absorbed $\times 10^{-6}$ $\left(\mathrm{mg} / \mathrm{cm}^{2}\right)$ | $\begin{aligned} & * \text { Flux } \times 10^{-4} \\ & \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{aligned}$ |
|  |  |  | Fluorene | 2.1 | $1.58 \pm 0.65$ | 0.71 |
|  |  |  | Phenanthrene | 6.8 | $2.14 \pm 0.68$ | 1.04 |
|  | Analytical method: | -Exposure area: | Anthracene | 3.7 | $0.32 \pm 0.09$ | 0.20 |
|  | -HPLC | $24 \mathrm{~cm}^{2}$ | Fluoranthene | 4.0 | $0.29 \pm 0.14$ | 0.21 |
|  |  |  | Pyrene | 2.1 | $0.20 \pm$-- | 0.12 |
|  |  |  | Benzo[b]fluoranthene | 0.9 | $<0.03 \pm 0.03$ | 0.007 |
|  |  | Components measured: | Benzo[k]fluoranthene | 0.4 | $<0.005 \pm 0.005$ | < 0.003 |
|  |  | -11 PAHs (see table | Benzo[a]pyrene | 0.9 | $<0.033 \pm 0.051$ | 0.008 |
|  |  | results) | Indenol[123-cd]pyrene | 0.6 | $<0.003 \pm 0.006$ | $<0.003$ |
|  |  |  | Dibenzo[ah]anthracene | 0.4 | $<0.008 \pm 0.008$ | $<0.003$ |
|  |  |  | *At 200 min after coal tar <br> \#Related to pyrene | pplication, th | exact data were no | shown in the article |
|  |  |  | -the use of pyrene as a the cumulative absorptio absorption of those with -it was suggested that th | arker of PAH of PAHs with her molecula situation is lik | bsorption through ower molecular we mass. <br> ly to occur in huma | g skin will underestimate ht, and overestimate the skin as well |



| Table 42 | PAH |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors Sart | Sartorelli et al. |  |  |  |  |  |  |
| Title | Dermal exposure assessment of polycyclic aromatic hydrocarbons: in vitro percutaneous penetration from lubricating oil |  |  |  |  |  |  |
| Source Inte | International Archive of Occupational and Environmental Health, 72: 528-532, 1999 |  |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |  |
| -In vitro percutaneous penetration of PAHs. Effect of the application form /vehicle | -mixture of PAHs <br> - static diffusion cells <br> -full thickness monkey skin <br> Analysis: <br> Mass balance <br> Analytical method: -HPLC | -Exposure surface concentration: PAHs were applied in lubrication oil or in acetone (ranging from 6.1 to $160 \mathrm{nmol} / \mathrm{cm}^{2}$ ) <br> -Exposure duration: Not clearly specified <br> -Exposure area: $-1.77 \mathrm{~cm}^{2}$ <br> Components measured: -13 PAHs (see table results) | Steady-state permeability coefficient (Kp), fluxes and lag time of the tested PAHs applied in lubrication oil or acetone (mean $\pm$ SD) |  |  |  |  |
|  |  |  | Component | $\mathrm{K}_{\mathrm{p}} \times 10^{-3}$ $(\mathrm{~cm} / \mathrm{h})$ lubricating oil | $\mathrm{K}_{\mathrm{p}} \times 10^{-3}$ $(\mathrm{~cm} / \mathrm{h})$ acetone | $t_{\text {lag }}$ <br> (h) <br> lubricating oil | $t_{\text {lag }}$ <br> (h) <br> Acetone |
|  |  |  | Naphthalene | $1.87 \pm 1.31$ | $6.31 \pm 2.49$ | $4.86 \pm 7.99$ | $1.18 \pm 0.01$ |
|  |  |  | Acenaphthalene | $1.72 \pm 1.76$ | $7.80 \pm 4.10$ | $8.37 \pm 3.44$ | $2.34 \pm 2.31$ |
|  |  |  | Fluorene | $1.64 \pm 1.66$ | $6.56 \pm 5.33$ | $5.70 \pm 3.02$ | $4.23 \pm 3.99$ |
|  |  |  | Anthracene | $0.93 \pm 0.98$ | $3.97 \pm 2.82$ | $17.55 \pm 4.73$ | $12.85 \pm 7.18$ |
|  |  |  | Phenanthrene | $0.50 \pm 0.28$ | $2.63 \pm 0.74$ | $15.15 \pm 3.10$ | $10.95 \pm 7.62$ |
|  |  |  | Pyrene | $0.17 \pm 0.04$ | $4.13 \pm 4.36$ | $13.38 \pm 8.91$ | $24.46 \pm 2.68$ |
|  |  |  | Benzo[a]anthracene |  | $1.72 \pm 2.60$ | * | $27.14 \pm 8.28$ |
|  |  |  | Chrysene | $0.22 \pm 0.12$ | $0.57 \pm 0.43$ | $26.12 \pm 3.34$ | $23.79 \pm 2.25$ |
|  |  |  | Benzo[b]fluoranthene |  | $0.09 \pm 0.04$ |  | $22.46 \pm 21.12$ |
|  |  |  | Benzo[k]fluoranthene |  | $0.09 \pm 0.04$ |  | $23.80 \pm 25.70$ |
|  |  |  | Benzo[a]pyrene |  | $0.23 \pm 0.20$ |  | $31.21 \pm 10.81$ |
|  |  |  | Dibenzo[ah]anthracene |  |  |  |  |
|  |  |  | Benzo[ghi]perilene |  |  |  |  |
|  |  |  | *below detection limit <br> The permeability results showed that the transport of PAHs through the skin is slower from lubricating oil than from acetone.This could be attributed to the liposolubility of PAHs and their affinity for oily liquids. <br> The permeability was significantly lower for naphthalene, acenaphthalene, fluorene, anthracene, phenanthrene and pyrene when applied in lubricating oil, except for chrysene where no significant difference was observed. <br> This was attributed to the fact that very often the concentration of chrysene was below the detection limit. |  |  |  |  |
|  |  |  |  |  |  |  |  |
|  |  |  | Comment: <br> No explicit information on exposure duration |  |  |  |  |


PAH

| able 44 PAH | PAH |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors Sar | Sartorelli et al. |  |  |  |  |  |  |
| Title | Dermal exposure assessment of polycyclic aromatic hydrocarbons: in vitro percutaneous penetration from coal dust |  |  |  |  |  |  |
| Source Tox | Toxicology and industrial Health, 17: 17-21, 2001 |  |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |  |
| -In vitro assessment of the dermal penetration of PAHs from coal dust through human skin comparison with the dermal penetration of pure PAH components | -coal dust <br> - dermatomed cadaver human abdominal skin (thickness $350 \mu \mathrm{~m}$ ) <br> -static diffusion cells | Exposure surface concentration: $-9 \mathrm{mg} / \mathrm{cm}^{2}$ of coal dust applied by adding a minimal amount of acetone for correct distribution on the skin surface -a mixture of seven PAHs was applied in $30 \mu$ of acetone without occlusion <br> -Exposure duration: 72 h <br> -Exposure area: $1.77 \mathrm{~cm}^{2}$ <br> Components measured: -PAHs (see table results) | Percentages of absorption of PAHs applied in acetone solution at (mean $\pm$ SD)Percentage at |  |  |  |  |
|  |  |  | Component | 6 h | 24 h | 48 h | 72 h |
|  |  |  | Phenanthrene | $9.58 \pm 4.70$ | $32.66 \pm 14.41$ | $39.26 \pm 19.83$ | 43.54土 |
|  |  |  | Pyrene | $1.86 \pm 2.05$ | $15.18 \pm 7.01$ | $37.51 \pm 8.06$ | 25.08 |
|  |  |  | Benzo[a]anthracene | $0.42 \pm 0.28$ | $4.54 \pm 1.78$ | $9.88 \pm 3.50$ | $51.98 \pm$ |
|  |  |  | Benzo[b]fluoranthene |  | $1.10 \pm 0.79$ | $3.71 \pm 1.74$ | 14.97 |
|  |  |  | Benzo[k] fluoranthene |  | $0.97 \pm 0.57$ | $3.45 \pm 1.54$ | $14.22 \pm 5.06$ |
|  |  |  | Benzo[a]pyrene |  | $1.40 \pm 0.78$ | $4.95 \pm 2.05$ | $6.75 \pm 3.46$ |
|  | Analytical method: -HPLC |  | Dibenzo[a,h]anthracene |  |  | $0.60 \pm 0.38$ | $\begin{aligned} & 6.19 \pm 2.88 \\ & 8.57 \pm 3.67 \\ & 1.94 \pm 1.34 \end{aligned}$ |
|  |  |  | *below detection limit |  |  |  |  |
|  |  |  | Since the concentration always under the detec assessed. <br> It was suggested that the could be attributed to the <br> Pure PAH compounds in Data on cumulative p dibenzo[a,h]anthracene penetration of phenanthre | PAHs in the n limit, no absence of d ysico-chemi <br> cetone cutaneous compared e was faster, | eceptor fluid a ercutaneous p <br> tectable penetr al properties of <br> enetration show to other PA decreasing afte | er exposure to netration of P <br> ation of PAHs coal <br> wed slower s, while the 24 h . | coal dust was AHs could be rom coal dust <br> enetration of percutaneous |



| Table 46 | Nonane, dodecane, tetradecane |  |  |
| :---: | :---: | :---: | :---: |
| Authors ${ }^{\text {a }}$ | Babu et al. |  |  |
| Title | Percutaneous absorption and skin irritation upon low-level prolonged dermal exposure to nonane, dodecane and tetradecane in hairless rats |  |  |
| Source $\quad$ Tox | Toxicology and Industrial Health, 20: 109-118, 2004 |  |  |
| Type / aim of study | Test material/ species technique / detection | Exposure condition | Results |
| -In vitro assessment of the percutaneous absorption of aliphatic hydrocarbons <br> evaluation of impact of prolonged dermal exposures to aliphatic hydrocarbons of varied chain length on skin irritation and biomarker responses. <br> Evaluation ofthe relation to the skin permeation and retention of these chemicals | -nonane <br> -decane <br> -tetradecane <br> - dorsal skin from male <br> CD hairless rats <br> Static diffusion cells <br> Analysis: <br> Mass balance <br> Analytical method: <br> -liquid scintillation <br> spectrometry <br> -HPLC | -Exposure concentration: 0.5 ml of mixture of aliphatic hydrocarbons containing their radiolabelled counterparts <br> -Exposure duration: <br> 8 h <br> -Exposure area: $0.636 \mathrm{~cm}^{2}$ <br> Components measured: -radiolabelled nonane, decane, tetradecane | The data showed that the steady-state flux of dodecane was about 3 folds higher than for nonane and about 77 fold higher than for tetradecane For all three aliphatic hydrocarbons the measured lag time was about 1 h . <br> The retention of chemicals was much higher in the stratum corneum than in the epidermis and dermis at all measured time points <br> The retention in the epidermis and dermis was highest for dodecane followed by nonane and tetradecane. <br> The relationship between molecular weight of the aliphatic hydrocarbons and skin retention in epidermis and dermis showed to be parabolic, which is in contrast to the stratum corneum where the relationship between molecular weight of aliphatic hydrocarbons (carbon chain length) and the absorption in the stratum corneum was linear. |
|  |  |  | Comment: <br> The data were presented only graphically |


| Table 47 Jet | Jet fuel |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Authors Kan | Kanikkannan et al. |  |  |  |  |  |  |
| Title Per | Percutaneous permeation and skin irritation of JP-8 + 100 jet fuel in a porcine model |  |  |  |  |  |  |
| Source ${ }^{\text {Tox }}$ | Toxicology Letters,119: 133-142, 2001b |  |  |  |  |  |  |
| Type / aim of study | Test material/ species / technique / detection | Exposure condition | Results |  |  |  |  |
| -In vitro assessment of percutaneous permeation of JP-8 + 100 <br> In vitro assessment of the effect of three performance additives (BHT, MDA and 8Q405 on percutaneous permeation of JP-8 across pig ear skin | Test material: <br> - JP-8 <br> - JP-8 + 100 <br> - JP-8 + BHT <br> - JP-8 + MDA <br> - JP-8 + 8Q405 <br> Experimental method and species: <br> - dermatomed pig ear skin $(500 \mu \mathrm{~m})$ <br> Static (Franz) diffusion cells <br> Analytical method: -liquid scintillating counting | -1 ml of jet fuels as described in column two spiked with radiolabelled tridecane, nonane, naphthlane and toluene <br> -duration of exposure: $24 \mathrm{~h}$ <br> -exposure area: $1.1 \mathrm{~cm}^{2}$ <br> Components measured: -tridecane <br> -nonane <br> -naphthlane <br> -toluene | Steady state (SS) flux (mean $\pm$ SD) of tridecane, nonane, naphthalene and toluene from JP-8, JP-8 + 100 and JP-8 + additives across pig ear skin |  |  |  |  |
|  |  |  | $\begin{aligned} & \begin{array}{l} \text { SS Flux } \times 10^{-3} \\ \left(\mathrm{mg} / \mathrm{cm}^{2} / \mathrm{h}\right) \end{array} \\ & \hline 1 \mathrm{P}^{*} 8^{*} \end{aligned}$ | Tridecane (\% (w/w) in jet fuel $=2.7 \%$ ) | Nonane (\% (w/w) in jet fuel = $1.1 \%$ ) | Naphthalene (\% (w/w) in jet fuel $=0.26 \%$ ) | Toluene (\% (w/w) in jet fuel $=0.06 \%$ ) |
|  |  |  | JP-8 + 100 JP-8 + BHT JP-8 + MDA JP-8 + 8Q405 | $1.318 \pm 0.155^{* *}$ $1.3223 \pm 0.059^{* *}$ $1.2530 \pm 0.11$ $1.465 \pm 0.093$ | $0.395 \pm 0.007^{* *}$ $0.396 \pm 0.011^{* *}$ $0.451 \pm 0.031$ $0.461 \pm 0.033$ | $\begin{aligned} & 0.419 \pm 0.033^{* *} \\ & 0.327 \pm 0.015^{* *} \\ & 0.386 \pm 0.020 \\ & 0.364 \pm 0.037 \end{aligned}$ | $\begin{aligned} & 0.094 \pm 0.001^{* *} \\ & 0.071 \pm 0.013^{* *} \\ & 0.114 \pm 0.009 \\ & 0.117 \pm 0.005 \end{aligned}$ |
|  |  |  | $\mathrm{K}_{\mathrm{p}} \times 10^{-4}(\mathrm{~cm}$ |  |  |  |  |
|  |  |  | JP-8 + 100 | 0.6102 | 0.4489 | 2.014 | 1.958 |
|  |  |  | *Data reproduced from Kanikkannan et al., 2001a (Table 25), **significantly different ( $\mathrm{p}<0.05$ ) compared to JP-8. <br> The permeation of tridecane was highest followed by nonane, naphthalene and toluene for all jet fuels. <br> The steady-state flux of all components showed to be significantly lower from JP-8 + 100 and JP-8 + BHT as compared to chemicals from JP-8, except the steady-state flux of naphthalene from JP-8 + 100 which was significantly higher than from JP-8. It was suggested that BHT minimizes the changes induced in the skin by inhibiting the formation of oxidative products and free radicals from JP-8. <br> MDA and 8Q405 showed no significant effect on the permeation of chemicals from JP-8. |  |  |  |  |
|  |  |  | Comment: <br> BHT - butylated hydroxytoluene (antioxidant), MDA - metal deactivator, <br> 8Q405 - detergent/dispersant <br> JP-8 +100 - (contains all three additives BHT, MDA and 8Q405) |  |  |  |  |

Table 48 Toluene, m-xylene

| Authors | Kezic et al. |
| :--- | :--- |


| Authors | Dermal absorption of neat liquid solvents on brief exposures in volunteers |
| :--- | :--- |




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[^0]:    ${ }^{1}$ The stratum corneum is the outermost layer of the skin and is made of several tight layers of dead, flat skin cells that shed about every 2 weeks; the stratum corneum is important for the barrier protection of the skin.

